

SCIENTIFIC REPORTS



OPEN

Improvement in the biochemical and chemical properties of badland soils by thorny bamboo

Received: 05 September 2016

Accepted: 08 December 2016

Published: 19 January 2017

Yo-Jin Shiau¹, Hsueh-Ching Wang¹, Tsai-Huei Chen², Shih-Hau Jien³, Guanglong Tian⁴ & Chih-Yu Chiu¹

Badland soils—which have high silt and clay contents, bulk density, and soil electric conductivity—cover a large area of Southern Taiwan. This study evaluated the amelioration of these poor soils by thorny bamboo, one of the few plant species that grows in badland soils. Soil physiochemical and biological parameters were measured from three thorny bamboo plantations and nearby bare lands. Results show that bamboo increased microbial C and N, soil acid-hydrolysable C, recalcitrant C, and soluble organic C of badland soils. High microbial biomass C to total organic C ratio indicates that soil organic matter was used more efficiently by microbes colonizing bamboo plantations than in bare land soils. High microbial respiration to biomass C ratio in bare land soils confirmed environmentally induced stress. Soil microbes in bare land soils also faced soil organic matter with the high ratio of recalcitrant C to total organic C. The high soil acid-hydrolysable C to total organic C ratio at bamboo plantations supported the hypothesis that decomposition of bamboo litter increased soil C in labile fractions. Overall, thorny bamboo improved soil quality, thus, this study demonstrates that planting thorny bamboo is a successful practice for the amelioration of badland soils.

The badland soil that is derived from mudrock (a class of fine-grained siliciclastic sedimentary rock) is unfavourable for plant growth because of its high clay and calcium carbonate contents¹. One of the common physiochemical properties of the badland soil is the low water infiltration, thus, most rainfalls on badland fields drain via surface runoff, leading to soil erosion and nutrient loss². Moreover, in badland soils, sodium and chlorine have been found to concentrate near soil surface during the dry season, which creates repulsive forces among soil particles, causing the soil to become rock-hard. On the other hand, during the rainy season, soils swell and soften upon water saturation, which has been shown to accelerate surface erosion³. Concentrated sodium and chlorine in this type of soil also increase soil electric conductivity, rendering it inhospitable to plant growth, resulting in bare landscapes in most badland ecosystems.

In Southern Taiwan, Plio-Pleistocene badland soils, consisting of up to 90% silt and clay combined, occupy more than 10,000 ha and are devoid of vegetation¹. Soil crusting is common on the surface of this badland soil, which increases its bulk density and penetration resistance. Previous research⁴ has indicated that root growth could be inhibited when penetration resistance exceeds 14 kg cm⁻², where the penetration resistance of the bare soil in the badlands of Southern Taiwan has been shown to be high (14.5 ± 1.47 kg cm⁻²)⁵.

Thorny bamboo (*Bambusa stenostachya*), a dense clumping bamboo with an average height of 12–15 m, is one of the only plant species that are able to grow in these uninhabitable soils. Research has shown that bamboo plants absorb nutrients from deep soils and retain soil water in infertile environments, as they are known to have deeper rhizomes and roots than many other non-woody plants⁶. A previous study reported that the ecological parameters of a highly degraded soil in India were significantly improved as a result of the increase in soil organic matter (SOM) by planting bamboos⁷.

Soil organic matter consists of different C fractions such as soluble, acid hydrolysable and recalcitrant C. In addition to the overall SOM pools increased by bamboo plantations, humification and the composition of SOM may be a good reflection of the impact of the plant on SOM⁸. During humification, organic compounds in plant

¹Biodiversity Research Center, Academia Sinica, Taipei 11529, Taiwan. ²Taiwan Forestry Research Institute, Taipei 10066, Taiwan. ³Department of Soil and Water Conservation, National Pingtung University of Science and Technology, Pingtung 912-01, Taiwan. ⁴Environmental Monitoring and Research Division, Monitoring and Research Department, Metropolitan Water Reclamation District of Greater Chicago, 6001 W. Pershing Road, Cicero, IL 60804, USA. Correspondence and requests for materials should be addressed to C.-Y.C. (email: bochiu@sinica.edu.tw)

Site	Vegetation	pH	EC ¹ (mS/cm)	Water Content (%)	Clay (%)	Sand (%)	Silt (%)	Texture	Bulk density	Total organic C (%)	Total N (%)	Classification
1	Bare land	8.63 ^a	7.59 ^b	7.2 ^d	43.5 ^a	20.2 ^c	36.3 ^a	clay	1.55 ^a	0.34 ^c	0.07 ^d	Typic Eutrustepts
	Bamboo	8.20 ^{ab}	0.80 ^c	15.9 ^b	30.2 ^{ab}	39.1 ^a	30.7 ^a	clay loam	1.23 ^b	0.90 ^b	0.13 ^{bc}	Typic Dystrustepts
2	Bare land	8.42 ^{ab}	12.59 ^a	8.7 ^{cd}	28.9 ^b	26.7 ^{bc}	44.4 ^a	clay loam	1.81 ^a	0.45 ^c	0.08 ^d	Typic Eutrustepts
	Bamboo	6.15 ^c	0.66 ^c	20.5 ^a	29.5 ^b	36.3 ^{ab}	34.2 ^a	clay loam	1.11 ^b	1.77 ^a	0.21 ^a	Typic Dystrustepts
3	Bare land	8.33 ^{ab}	13.33 ^a	10.0 ^c	27.1 ^b	25.5 ^{bc}	47.4 ^a	clay loam	1.80 ^a	0.47 ^c	0.10 ^{cd}	Typic Eutrustepts
	Bamboo	8.17 ^b	0.65 ^c	17.0 ^b	29.0 ^b	26.8 ^{bc}	44.2 ^a	clay loam	1.23 ^b	1.11 ^b	0.14 ^b	Typic Dystrustepts

Table 1. Soil properties (0–10 cm) under three thorny bamboo plantations and adjacent bare lands in Southern Taiwan. ¹EC = Electrical Conductivity. Values with the same superscripted letters in each column are not significantly different at $p = 0.05$ based on the Tukey's HSD comparison.

Site	Vegetation	S _b OC _{KCl} (μg/g soil)	NH ₄ ⁺ _{KCl} (μg/g soil)	NO ₃ ⁻ _{KCl} (μg/g soil)	S _b ON _{KCl} (μg/g soil)	TDN _{KCl} (μg/g soil)	S _b OC _{KCl} / TOC (%)	S _b ON _{KCl} / TN (%)
1	Bare land	27.8 ^c	4.4 ^{bc}	3.9 ^b	5.4 ^a	13.7 ^c	0.84 ^a	0.79 ^a
	Bamboo	41.9 ^b	7.5 ^{bc}	4.4 ^b	3.5 ^a	15.4 ^c	0.49 ^{bc}	0.32 ^a
2	Bare land	26.5 ^c	4.0 ^c	8.5 ^{ab}	12.2 ^a	24.7 ^{bc}	0.60 ^b	1.81 ^a
	Bamboo	58.0 ^a	20.6 ^a	15.0 ^a	11.3 ^a	46.9 ^a	0.33 ^c	0.58 ^a
3	Bare land	27.4 ^c	6.5 ^{bc}	6.4 ^b	10.4 ^a	23.3 ^{bc}	0.59 ^b	1.08 ^a
	Bamboo	41.8 ^b	10.9 ^b	4.9 ^b	15.0 ^a	30.8 ^b	0.38 ^c	1.29 ^a

Table 2. Concentrations of soluble N and soluble organic C in 2 M KCl extracts, S_bON/TN ratio, and S_bOC/TOC ratio in the top 10 cm of soil from thorny bamboo plantations and adjacent bare land in Southern Taiwan. [Soil soluble organic carbon (S_bOC); ammonium (NH₄⁺); nitrate (NO₃⁻); Soil soluble organic nitrogen (S_bON); total dissolved nitrogen (TDN); total organic carbon (TOC); total nitrogen (TN)]. Values with the same superscripted letters in each column are not significantly different at $p = 0.05$ based on the Tukey's HSD comparison.

litter could be transformed into humic substances, whereupon their structures become significantly altered compared to the original forms in plant materials⁹. To the best of our knowledge, the mechanisms by which thorny bamboo grows in badland ecosystems, and the interactions between the growth of the bamboo, soil nutrients, and microorganisms in this hostile soil environment are still not well understood. Thus, the objective of the study was to find out the difference in physical, chemical, and biological parameters of soil between bamboo plantations and adjacent bare lands in Southern Taiwan. We hypothesized that the presence of bamboo in these soils would result in significant increases in SOM and soil nutrients, and that the increase in SOM would improve soil physicochemical properties and increase soil microbial biomass.

Results

The analysis of soil properties from the three sampling locations is shown in Table 1. Soils were mostly loam to clay loam, containing high silt and/or clay content. Soil was slightly alkaline (pH = 8.3) at all locations except at the bamboo plantation in Site 2, where the soil was slightly acidic (pH = 6.2). Generally, the soil was distinctively different between bamboo plantations and bare land soils. Soil electric conductivity was significantly lower in bamboo plantation soils (0.7 mS cm⁻¹) than in bare land soils (11.2 mS cm⁻¹) ($p < 0.05$), and soil water content was higher at bamboo plantations than in bare land soils ($p < 0.05$).

The soluble organic carbon (S_bOC) content, measured using potassium chloride (KCl) extracts (S_bOC_{KCl}), was much higher at bamboo plantations than in bare land soils ($p < 0.05$) (Table 2). Site 2 had the highest S_bOC_{KCl} among the bamboo plantations, which was more than 2 times higher compared to bare land soil ($p < 0.05$). A similar trend was observed for S_bOC using the hot water extraction method (S_bOC_{HW}). The mean S_bOC_{HW} was nearly 6 times higher in bamboo plantation soils than bare land soils ($p < 0.05$), while the highest S_bOC_{HW} was measured in the bamboo plantation soil from Site 2 (Table 3). The S_bOC values from all locations were higher from samples extracted using the hot water method compared to samples extracted using KCl. The bare land soils from the three sampling locations had similar levels of S_bOC, irrespective of the extraction method used.

The difference in extractable nitrogen (N) between bamboo plantations and bare land soils depended on N species. The concentrations of KCl-extracted NH₄⁺ (NH₄⁺_{KCl}) and total dissolved nitrogen (TDN_{KCl}) at the three sites were 2–4 times higher at bamboo plantations than in bare land soils ($p < 0.05$), whereas, KCl-extracted NO₃⁻ (NO₃⁻_{KCl}) and soluble organic nitrogen (S_bON_{KCl}) were similar between bamboo plantations and bare land soils. Similarly, the concentration of NH₄⁺_{HW} and TDN_{HW} were 3–10 times higher in bamboo plantation soils than in bare land soils ($p < 0.05$), whereas NO₃⁻_{HW} and S_bON_{HW} were not significantly different between bamboo plantations and bare land soils, with the exception of S_bON_{HW} from Site 2, where NO₃⁻_{HW} and S_bON_{HW} were higher in bamboo plantation soils than bare land soils. The S_bOC_{HW} to total organic carbon (TOC) ratio (S_bOC_{HW}/TOC) was found to be higher in bamboo plantation soils than in bare land soils, however, the opposite trend was observed for S_bOC_{KCl}/TOC. No differences in the S_bON_{HW} or S_bON_{KCl} to total nitrogen (TN) ratios (S_bON_{HW}/TN or S_bON_{KCl}/TN) were observed between bamboo plantations and bare land soils.

Site	Vegetation	S _b OC _{HW} (μg/g soil)	NH ₄ ⁺ _{HW} (μg/g soil)	NO ₃ ⁻ _{HW} (μg/g soil)	S _b ON _{HW} (μg/g soil)	TDN _{HW} (μg/g soil)	S _b OC _{HW} /TOC (%)	S _b ON _{HW} /TN (%)
1	Bare land	62.8 ^c	1.5 ^c	2.3 ^b	0.9 ^b	3.8 ^d	1.88 ^c	0.11 ^a
	Bamboo	215.8 ^b	9.4 ^b	2.8 ^{ab}	4.7 ^b	16.9 ^{bc}	2.40 ^b	0.43 ^a
2	Bare land	70.3 ^c	1.9 ^c	4.4 ^{ab}	1.0 ^b	6.6 ^{cd}	1.58 ^c	0.22 ^a
	Bamboo	543.7 ^a	16.4 ^a	13.4 ^a	16.4 ^a	41.5 ^a	3.01 ^a	0.72 ^a
3	Bare land	78.6 ^c	2.5 ^c	4.1 ^{ab}	3.9 ^b	9.9 ^{cd}	1.69 ^d	0.37 ^a
	Bamboo	282.8 ^b	14.1 ^{ab}	3.7 ^{ab}	9.5 ^{ab}	27.3 ^b	2.54 ^b	0.74 ^a

Table 3. Concentrations of soluble N and soluble organic C in hot water extracts, S_bON/TN ratio, and S_bOC/TOC ratio in the top 10 cm of soil from thorny bamboo plantations and adjacent bare land in Southern Taiwan. [Soil soluble organic carbon (S_bOC); ammonium (NH₄⁺); nitrate (NO₃⁻); Soil soluble organic nitrogen (S_bON); total dissolved nitrogen (TDN); total organic carbon (TOC); total nitrogen (TN)]. Values with the same superscripted letters in each column are not significantly different at $p = 0.05$ based on the Tukey's HSD comparison.

Site	Vegetation	C _{mic} (μg C/g soil)	Respiration rate (μg C/g soil/hr)	N _{mic} (μg N/g soil)	Mineralisable N (μg N/g soil/d)	C _{mic} /TOC (%)	N _{mic} /TN (%)	Respiration/C _{mic} (g CO ₂ -C/g microbial-C/h)
1	Bare land	4.3 ^c	1.33 ^c	1.8 ^d	0.00 ^c	0.13 ^b	0.26 ^b	0.22 ^a
	Bamboo	40.0 ^b	3.13 ^{ab}	44.5 ^c	0.87 ^b	0.51 ^a	3.46 ^a	0.11 ^b
2	Bare land	3.5 ^c	0.88 ^c	2.2 ^d	0.22 ^{bc}	0.08 ^b	0.28 ^b	0.25 ^a
	Bamboo	69.0 ^a	3.73 ^a	84.4 ^a	3.04 ^a	0.40 ^a	4.37 ^a	0.05 ^b
3	Bare land	5.8 ^c	1.28 ^c	5.2 ^d	0.00 ^c	0.13 ^b	0.56 ^b	0.31 ^a
	Bamboo	35.9 ^b	3.79 ^b	63.0 ^b	0.79 ^{bc}	0.33 ^a	4.89 ^a	0.08 ^b

Table 4. Soil microbial biomass, potentially mineralisable nitrogen, and microbial quotient in soils from thorny bamboo plantations and adjacent bare lands in Southern Taiwan. [Microbial biomass carbon (C_{mic}); microbial biomass nitrogen (N_{mic}); total organic carbon (TOC); total nitrogen (TN)]. Values carrying the same letters in each column are not significantly different at $p = 0.05$ based on the Tukey's HSD comparison.

Soil microbial biomass C (C_{mic}) and N (N_{mic}) were found to be 7–10 times higher in bamboo plantation soils than in bare land soils ($p < 0.05$) (Table 4). Soil respiration rates and potentially mineralisable N were also higher at bamboo plantations than in bare land soils. The amount of total mineralisable N was almost negligible in bare land soils, although a slight respiration rate was detected. The C_{mic}/TOC and N_{mic}/TN ratios were higher at bamboo plantations than in bare land soils ($p < 0.05$), whereas respiration/C_{mic} was higher in bare land soils than in bamboo plantation soils ($p < 0.05$).

The S_bOC_{KCl} and S_bOC_{HW} in both bamboo plantations and bare land soils were significantly correlated with TOC and C_{mic}, respectively (Fig. 1). However, values for S_bOC_{KCl} and S_bOC_{HW} in the bare land soils were clustered and showed less dependence on TOC or C_{mic}. In this study S_bOC_{HW} was more strongly correlated with both TOC and C_{mic} than S_bOC_{KCl}, indicating that S_bOC_{HW} could be a better reflection of the effects of bamboo on soil TOC and C_{mic} than S_bOC_{KCl}. The overall S_bON_{HW} was significantly correlated with both TN and N_{mic} but not with S_bON_{KCl} (Fig. 2). The strong correlations between S_bOC_{HW} and TOC, as well as S_bON_{HW} and TN, imply that hot water may help to release more organic substrates from the soils.

The ΔlogK (logarithmic ratio of light absorbance of humic acids at 400 and 600 nm; ΔlogK = log(A₄₀₀/A₆₀₀)) values observed in soils from thorny bamboo plantations were significantly higher than those of bare land soils (Table 5). The contents of soil acid-hydrolysable C pool I (soil organic C extracted with 5 N H₂SO₄; AHPI-C), acid-hydrolysable C pool II (soil organic C extracted with 26 N H₂SO₄; AHPII-C), and recalcitrant C pool (residue after acid-hydrolysis; RP-C) were significantly higher in soils from thorny bamboo plantations compared to those of bare land soils. Considering that the TOC contents differed between the two types of soils, the ratios of acid-hydrolysable and recalcitrant contents to TOC were calculated. The ratios of AHPI-C and AHPII-C to TOC (AHPI-C/TOC and AHPII-C/TOC) were higher in bamboo plantation soils, while the ratio of RP-C to total organic C (RP-C/TOC) was higher in bare land soils.

Discussion

In this study, we compared physicochemical properties of bare land soils and those planted with thorny bamboo. The data from this study shows significantly higher soil C_{mic}, N_{mic}, respiration rates, total mineralisable N, S_bOC_{HW}/TOC, and C_{mic}/TOC in soils from the bamboo plantations compared to bare land soils, mostly due to fast growth and high biomass production of bamboos^{10,11}. In the present study, thorny bamboo resulted in an increase in TOC content, which also improved the labile SOM (i.e. S_bOC) of the previously bare lands. For both KCl and hot water extraction methods, S_bOC was 2–7 times higher in soils from bamboo plantations compared to soils from the adjacent bare lands, indicating the strong contribution of thorny bamboo growth to soil C pools.

Soil organic matter is one of the most important indicators of soil quality¹², as it improves soil physicochemical properties (e.g. decreasing soil bulk density and increasing soil porosity), and increases soil water content

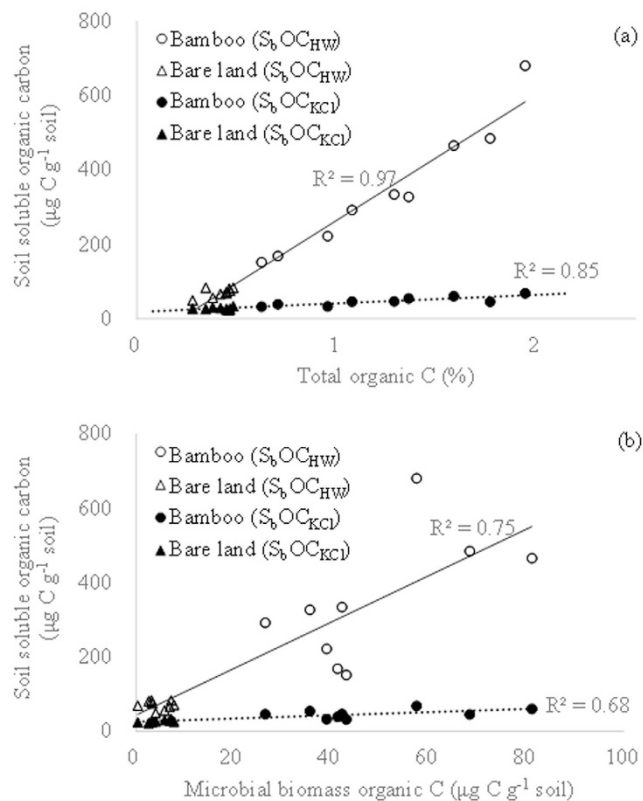


Figure 1. Relationship between soil soluble organic carbon in 2 M KCl extracts ($S_b\text{OC}_{\text{KCl}}$) and soil total organic carbon (a) and that between hot water extracts ($S_b\text{OC}_{\text{HW}}$) and soil microbial biomass carbon (b) in thorny bamboo plantations and adjacent bare lands.

as a result high water holding capacity of SOM^{13,14}. These benefits provided by SOM may explain the overall improvement of soils at bamboo plantations in the present study, where soil bulk density decreased by ~30% and soil water content increased by 10%. Moreover, the highest TOC in the bamboo plantation from Site 2 helped to improve the soil quality to the extent that Site 2 was observed to have the lowest bulk density and the highest water content of all the sites in the present study.

The higher $S_b\text{OC}_{\text{HW}}/\text{TOC}$ observed in bamboo plantation soils suggested that the observed SOM changes in bamboo plantation soils was not only in the total C pool as a result of litter from the bamboo plants, but also labile fractions as a result of bamboo litter decomposition. The lower $S_b\text{OC}_{\text{KCl}}$ and $S_b\text{OC}_{\text{HW}}$ of bamboo plantation soils in the current study than those in a previous study conducted in a mountainous area of Central Taiwan¹⁵ could be due to the fact that soils in the Central Taiwan contained higher TOC. In addition, the considerable lower $S_b\text{OC}/\text{TOC}$ from both KCl and hot water extractions of bamboo plantation soils in the current study than those in a mountainous area of Central Taiwan¹⁵ implies that most of the TOC in badland soils was strongly bound to the soil particles as a result of the high silt and clay contents.

The organic materials and dense roots provided by bamboo have shown to reduce overland flow velocity and infiltration rates^{16,17}, thereby reducing soil erosion¹⁸. The export of nutrients associated with sediment loss was found to decrease as a result of bamboo plantations¹⁹, resulting in an overall accumulation of soil nutrients in bamboo plantations, as well as an improvement in soil quality²⁰. The fact that the NH_4^+ , $S_b\text{ON}$, and TDN measured in bamboo plantation soils in the current study were distinctively higher in comparison to bare land soils may be due to the less soil erosion in bamboo plantation soils.

The N_{mic} was found to be positively correlated with TN, as soils with higher levels of N support higher microbial biomass²¹. The higher N_{mic}/TN ratios at the bamboo plantations in this study indicate that bamboo plantations may have a higher capacity for N retention through the synthesis of N in microbial biomass. Potentially mineralisable N has been considered an active fraction of soil organic N²². Thus, the high potentially mineralisable N in the bamboo plantation soils in our study implies that the bamboo plantations contained high levels of active soil organic N.

Labile SOM ($S_b\text{OC}$ and $S_b\text{ON}$) are the most readily available energy source for soil microbial growth²³. In the present study, $S_b\text{OC}$ was positively correlated with C_{mic} in bamboo plantation soils. The positive relationships between $S_b\text{OC}$, TOC, and C_{mic} implies that $S_b\text{OC}$ may be derived from the decomposition of soil TOC as a result of microbial activity²⁴. The increase in C_{mic} , owing to the readily available energy in $S_b\text{OC}$, was found to increase the soil respiration rates of bamboo plantation soils. Moreover, higher $C_{\text{mic}}/\text{TOC}$ and $S_b\text{OC}_{\text{HW}}/\text{TOC}$ ratios indicate that the available organic substrates were used more efficiently by microbes in soils planted with bamboos²⁵.

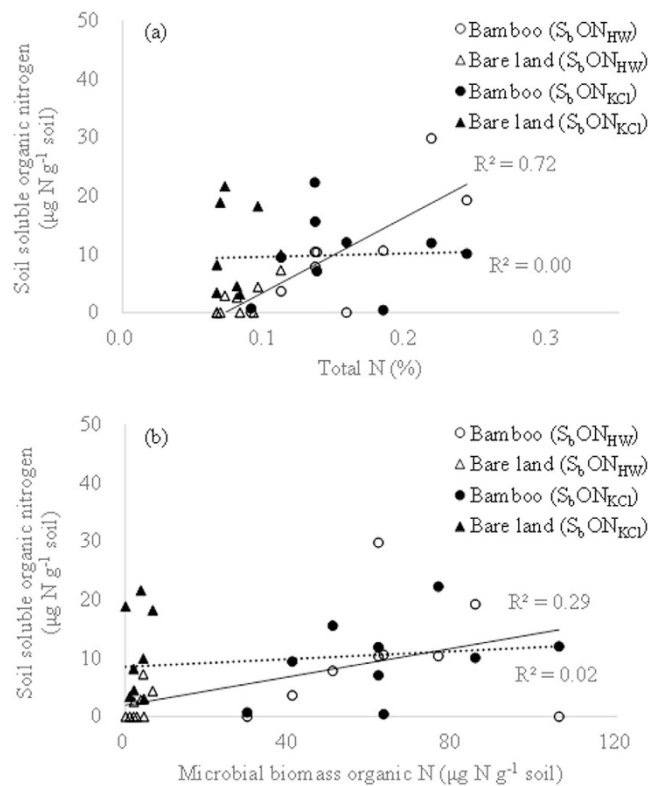


Figure 2. Relationship between soil soluble organic nitrogen in 2 M KCl extracts ($S_b ON_{KCl}$) and soil total organic nitrogen (a) and that between hot water extracts ($S_b ON_{HW}$) and soil microbial biomass nitrogen (b) in thorny bamboo plantations and adjacent bare lands.

Site	Vegetation	$\Delta \log K$	AHPI-C (mg C/g soil)	AHPII-C (mg C/g soil)	RP-C (mg C/g soil)	AHPI-C/TOC	AHPII-C/TOC	RP-C/TOC
1	Bare land	0.35 ^b	0.31 ^c	0.30 ^c	5.70 ^{bc}	0.04 ^c	0.04 ^d	0.75 ^a
	Bamboo	0.85 ^a	3.79 ^b	2.10 ^b	8.88 ^{ab}	0.28 ^b	0.16 ^{ab}	0.68 ^{ab}
2	Bare land	0.48 ^b	0.27 ^c	0.60 ^c	5.31 ^{bc}	0.04 ^c	0.09 ^{cd}	0.76 ^a
	Bamboo	0.73 ^a	7.72 ^a	3.66 ^a	11.05 ^a	0.42 ^a	0.20 ^a	0.59 ^b
3	Bare land	0.37 ^b	0.18 ^c	0.32 ^c	4.74 ^c	0.03 ^c	0.05 ^d	0.68 ^{ab}
	Bamboo	0.73 ^a	3.20 ^b	1.52 ^{bc}	8.01 ^{abc}	0.24 ^b	0.12 ^{bc}	0.65 ^{ab}

Table 5. The degree of humification as indicated by $\Delta \log K$, acid-hydrolysable pool I (AHPI-C) and II carbon (AHPII-C), recalcitrant pool carbon (RP-C), and ratios of labile and recalcitrant carbon to total organic carbon (TOC) of soils from three thorny bamboo plantations and adjacent bare lands in Southern Taiwan. Values carrying the same letters in each column are not significantly different at $p = 0.05$ based on the Tukey's HSD comparison.

The respiration/ C_{mic} ratio has been used in microbial studies to indicate the ecological efficiency (i.e. energy required to support the metabolism of per unit C_{mic} in soil) of soil microbial communities^{26,27}. A high respiration/ C_{mic} ratio indicates the inefficient use of energy, while a low respiration/ C_{mic} ratio indicates high efficiency, and that higher quantities of C are utilized for biomass production^{26,28,29}. In the present study, soil respiration appeared to be higher in bamboo plantation soils than in bare land soils. However, the respiration/ C_{mic} ratio showed a reverse trend, and was significantly lower in soils planted with bamboo than in bare land soils ($p < 0.05$). This result implies that the lower respiration/ C_{mic} ratio of bamboo plantation soils could be attributed to high microbial biomass rather than microbial activity. Moreover, the higher respiration/ C_{mic} ratio could be a good indicator that microbes were under greater stress in bare land soils.

The value of $\Delta \log K$ is a soil index to evaluate the humification status of SOM, and it generally decreases with increasing SOM humification. The lower $\Delta \log K$ values observed in bare land soils indicate a high degree of humification, as there is little input of fresh litter into the soil. The growth of bamboo adds fresh litter to soils and causes the accumulation of SOM^{30,31}. Wang, *et al.*³² noted that bamboo litter contains a high portion of O-alkyl-C, a C functional group that can be easily decomposed by soil microbes, which is thought to contribute to increases in the labile organic C of soils. The high AHPI-C/TOC and AHPII-C/TOC ratios of soils planted with

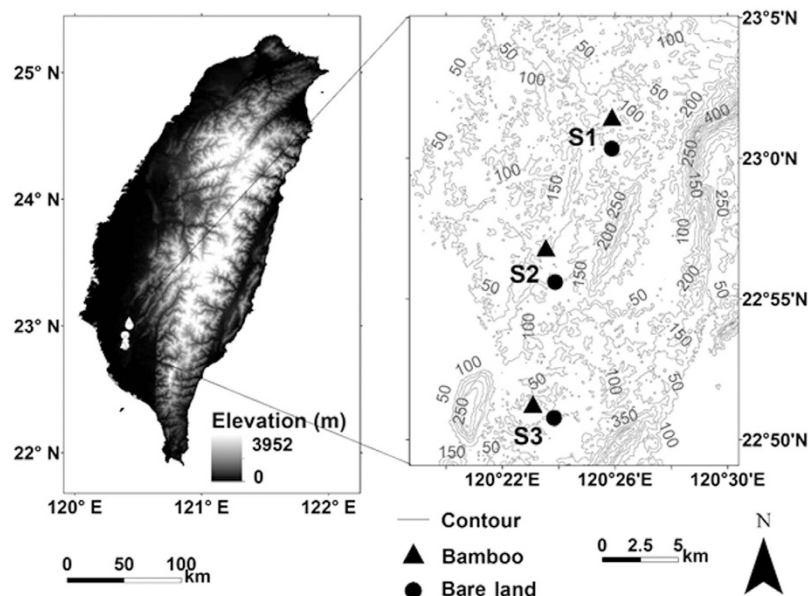


Figure 3. Map of soil sampling locations in Southern Taiwan. Elevation data used to create Fig. 3 was obtained using the Advanced Spaceborne Thermal Emission and Reflection Radiometer (ASTER) digital elevation model (DEM) at a resolution of 30 m (<https://asterweb.jpl.nasa.gov/gdem.asp>), while contour lines were created using the ArcGIS v.10.0 software package (ESRI [<http://www.esri.com/>], Redlands, CA, USA). (Triangles: thorny bamboo sites; circles: bare land sites).

bamboo support the notion that the decomposition of bamboo litter increases acid-hydrolysable C pools. The acid-hydrolysable C pool is small in size and has a rapid turnover, which responds rapidly to changes in C supply and affects microbial activity^{33,34}. The higher quantity of acid-hydrolysable C pools in bamboo plantation soils could be one of indicators of their high soil quality. The high RP-C/TOC ratio of bare land soils indicates that the C of the SOM composition is relatively more recalcitrant (i.e. high molecular weight C, irregular structure, and long turnover), and therefore has a higher resistance to chemical degradation and decomposability³⁵.

In addition, the fact that the acid-hydrolysable C and S_bOC displayed similar trends between the sampling sites implies that both methods extracted similar C pools³⁶. The observation that acid-hydrolysable C was 10–20 times higher than S_bOC at all sites may be due to the fact that acid-hydrolysable C contained more slow-turnover C than hot water- or KCl-extracted C³⁷. Hot water-extracted soil organic C has been typically considered to be readily metabolisable³⁸, while acid-hydrolysable C has been shown to be composed of bioreactive C in soil, even if it is not readily used by microbes³⁹. Therefore, acid-hydrolysable C is less sensitive to environmental changes, while both acid-hydrolysable C and S_bOC represent the labile fractions of soil C pools³⁶.

In conclusion, this study demonstrates that planting of thorny bamboos in uninhabitable badland soils appears to be a successful practice for soil amelioration. The thorny bamboo helped to increase SOM and improve soil physicochemical properties that contribute to soil improvement.

Methods

Site and soil sampling. This study was carried out at three locations in southwestern Taiwan—Site 1 at Zuozhen, Tainan (120°26'E, 23°0'N), Site 2 at Longqi, Tainan (120°23' E, 22°54'N), and Site 3 at Tianliao, Kaohsiung (120°24'E, 25°51'N)—with an average altitude of 100–300 m (Fig. 3).

The lands chosen as sampling locations for this study are owned by the Forestry Bureau and accessed for public. Research proposals were approved and granted by the Ministry of Science and Technology, Taiwan. Field and laboratory studies did not involve any animal husbandry, nor any protected or endangered biological species.

At each site, it involved badland soils with and without bamboos. In badland, desiccation cracks occur on the slope surfaces extending to a few centimeters in the dry season. During the rainy season, the soil expands and closes the cracks to form crusts in 1–2 cm thick³. Soils in the three studied locations have high clay and silt contents (clay: 27–44%; silt: 31–47%), primarily inherited from the parent mudrock. Based on U.S. Soil Taxonomy⁴⁰, soils in the badland lacking of vegetation were classified as Typic Eutrustepts, and those with the bamboo as Typic Dystrudepts. Thorny bamboo, which was cultivated in the early 1900s⁴¹ by local residents for bamboo shoots and stems, is the predominant plant on the north-facing sides of hogbacks (ridges with a sharp summit and steeply sloping sides) (Fig. 4). The south-facing sides of hogbacks are composed of bare land soils with no plant coverage. Interestingly, the downstream valley near these badlands (only a few km away) are well cultivated as orchards with various fruit trees, such as mango, jujube, and banana, indicating that original mudstone soils can be ameliorated for agricultural uses after certain improvements. At each site, three transects (replications), 50 m apart, were demarcated at both north-facing and south-facing sides of the hogbacks. Thus, there were a total of 18 transects (composited samples) for the study (2 vegetation types: bamboo and bare land soil × 3 replications × 3 locations

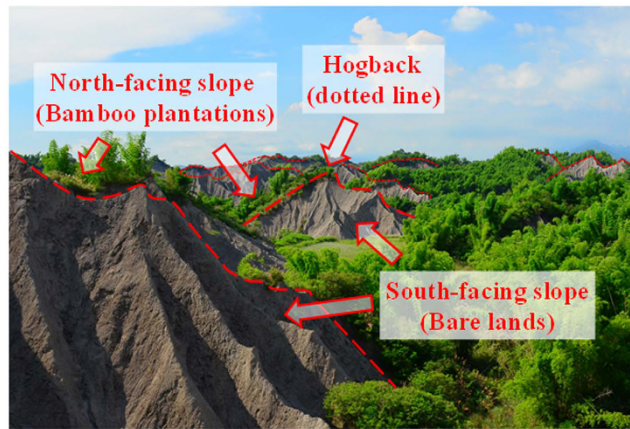


Figure 4. Landscape of mudstone badlands at Zuozhen in Tainan City, Southern Taiwan. South-facing slopes are barren while north-facing slopes are covered with thorny bamboo (photo by CYC, corresponding author).

for each soil type). A soil sample, composed of 8 soil cores, was collected in each transect by a soil auger (4.25 cm in diameter) and stored in plastic bag. The collected soil samples were stored at 4 °C until laboratory analysis.

Laboratory analyses. *General properties of soil.* Soil pH was measured in a 1:1 soil to water suspension using a portable Jenco 6009 pH/mV meter. Soil electrical conductivity was measured using a saturated extract method⁴². Soil total organic carbon (TOC) was analysed using a Fisons NA1500 elemental analyser (ThermoQuest Italia, Milan, Italy) after removing carbonate with 1 N HCl. Briefly, 0.3 g of soil was weighted out, mixed with 0.5 ml of HCl and air dried for 12 h before analysing with the elemental analyser. Soil TN was determined using the Kjeldahl digestion method⁴³. Deionized water was added to the digested solution to a volume of 100 ml and analysed using the previously described indophenol method⁴⁴ and measured spectrophotometrically (UV-1201, Shimadzu Corp., Kyoto, Japan). To determine the soil water content, soil was oven dried at 105 °C for 24 h and weighed for water loss. Soil texture was determined using the hydrometer method⁴⁵. Soil bulk density was determined using the weight and volume of soil samples collected via cores.

Soil extractable nutrients. Soil nutrient contents, including NH_4^+ , NO_3^- , S_bON , and S_bOC , were measured using two different extraction methods, KCl extraction and hot water extraction, described in Chen *et al.*⁴⁶ with modification. For the KCl extraction, 5 g of air-dried soil from each replicate sample was weighted out, placed in a 250-ml conical flask, and injected with 50 ml of 2 M KCl. The flask was sealed with plastic paraffin film and shaken for 60 min at 150 rpm. The solution was collected by filtering the slurry using filter paper (Whatman No. 5).

For the hot water extraction method, 6 g of air-dried soil samples were weighed, placed in a 50-ml centrifuge bottle, and injected with 30 ml of distilled deionized water. The bottle was placed in an 80 °C water bath and incubated for 18 h, subsequently shaken for 5 min at 150 rpm, and the solution was collected by filtering the slurry using the same size filter paper as KCl extracts. Extracted NH_4^+ was analysed using the indophenol method with a spectrophotometer at 635 nm wavelength (UV-1201, Shimadzu Corp., Kyoto, Japan). The NO_3^- was determined using the cadmium reduction method⁴⁷ with a flow injection analyser (SP-8001, Metertech Inc., Taipei City, Taiwan). Soil TDN was digested to NO_3^- using the persulfate method⁴⁸ and was then analysed as aforementioned with a flow injection analyser. The S_bON was calculated by subtracting NO_3^- and NH_4^+ from TDN. The S_bOC was measured using a TOC analyser (1010, O.I. Analytical, Texas, USA).

Soil microbial biomass C and N. Soil C_{mic} and N_{mic} were analysed using a chloroform-fumigated extraction method⁴⁹. Briefly, four 10 g aliquots of soil were weighed out from each sample and placed in a 250-ml conical flask. Then, soil samples were equally separated into two sets of 36 samples. One set of soil samples were fumigated using chloroform for 24 hours and extracted with 50 ml of potassium sulphate (0.5 M K_2SO_4). The other set of soil samples were directly extracted with 50 ml of K_2SO_4 (0.5 M). The C in K_2SO_4 solutions was determined using a TOC analyser and N was analysed using a previously described protocol⁵⁰. The C_{mic} and N_{mic} were calculated as the differences between the fumigated and unfumigated K_2SO_4 -extractable C and N by multiplying coefficients ($K_{\text{EC}} = 2.22$ and $K_{\text{EN}} = 4.95$) [(fumigated C (N) – unfumigated C (N)) × coefficients]^{51,52}.

Soil respiration. Soil respiration rate was estimated from the total amount of CO_2 -C during a 3-day incubation using an alkali method^{15,53}. Briefly, 20 ml of 0.05 M NaOH was injected into a 250-ml flask and a plastic tube and 20 g of fresh soil sample was placed in the flask. After 3 days of incubation at 25 °C, the 20 ml NaOH solution was titrated using 0.05 M HCl with phenolphthalein and BaCl_2 .

Total mineralisable N. Total mineralisable N was determined using the waterlogging incubation method⁵⁴. Briefly, 5 g of moist field soil was weighed out from each sample, placed in a 250-ml centrifuge bottle with 25 ml

of distilled deionized water, and incubated and shaken in an incubator at 40 °C for 7 d. After incubation, 25 ml of 4 M KCl was injected into the centrifuge bottle and shaken for 1 h at 150 rpm. Subsequently, the bottle was centrifuged at 2000 rpm for 20 min and 50 ml of water was extracted and filtered for the determination of NH_4^+ concentration.

Extraction of humic acids for photometric analysis. One gram of air-dried soil from each sample was weighed and shaken at 150 rpm with 30 ml of 0.1 N NaOH at 100 °C for 30 min. The Na_2SO_4 (2 ml) was added and the samples were subsequently centrifuged at $10,000 \times g$ for 15 min. Precipitates were further extracted with a mixed solution of 20 ml of 0.1 N NaOH and 3% Na_2SO_4 , and centrifuged at $10,000 \times g$ for 15 min, twice. The extractant was quantified to 100 ml with deionized water and then acidified with 1 ml concentrated H_2SO_4 (98%). Precipitates were subsequently dissolved in 30 ml of 0.01 N NaOH to determine the ratio of humic acids⁹ using a spectrophotometer (Hitachi U-2000). The light absorbance of humic acids was measured at 400 and 600 nm to determine their molecular fractions, as the smaller fractions of humic substrates tend to absorb shorter wavelengths of light⁵⁵. The degree of humification was determined by calculating the $\Delta\log K$ value, which is the logarithmic ratio of the absorbance at 400 and 600 nm ($\Delta\log K = \log(A_{400}/A_{600})$)⁵⁶. The $\Delta\log K$ is an inverse index of the condensation of the aromatic network in the macromolecules of humic acids, representing the degree of humification.

Determination of acid-hydrolysable and recalcitrant C. Acid hydrolysis is a chemical fractionation method used to separate acid-hydrolysable SOM and recalcitrant SOM pools (unhydrolysable)³³. The two-step acid hydrolysis is used to isolate and quantify labile and recalcitrant substances of SOM with sulphuric acid (H_2SO_4) as the extractant⁵⁷. Briefly, 0.5 g of each soil sample was hydrolysed with 20 ml of 5 N H_2SO_4 at 105 °C for 30 min in Pyrex flasks with Allihn condensers. Samples were centrifuged at $20,000 \times g$ for 10 min and decanted to collect the hydrolysable SOM from each sample. Residues were flushed with 20 ml of deionized water, and the extract was added to the previous hydrolysate. The C from the hydrolysate was regarded as the AHPI-C. The remaining residual SOM was hydrolysed with 2 ml of 26 N H_2SO_4 overnight while shaking continuously at room temperature. The concentration of H_2SO_4 in the sample was brought down to 2 N by dilution and hydrolysed for 3 h by heating at 105 °C. The resulting hydrolysate was considered as AHPII-C. The remaining SOM in the sample, regarded as RP-C, was flushed with 30 ml of deionized water and dried at 60 °C in a pre-weighed crucible. AHPI-C and AHPII-C were measured using a TOC analyser (1010, O.I. Analytical, Texas), while the RP-C was determined by an NSC elemental analyser (Fisons NA1500).

All analyses in the study were performed in duplicate. All results have been converted on dry weight basis.

Statistical analyses. Differences in soil nutrients between bare lands and bamboo plantations at all locations were tested for significance using a one-way analysis of variance (one-way ANOVA). When the one-way ANOVA revealed interactions between sites and soil physicochemical properties, Tukey's honestly significant difference (HSD) test was applied to further test the means of all pairs of dependent variables. Relationships between two soil properties were analysed using bivariate regression analysis. All statistical analyses were performed using JMP (Version Pro 10, SAS Institute, Cary, NC, USA). The level of significance was set at 0.05 for all tests.

References

- Lee, D., Lin, H. & Wu, J. The basic properties of mudstone slopes in southwestern Taiwan. *Journal of GeoEngineering* **2**, 81–95 (2007).
- Chen, S. Z. *A study on the erosion characteristics of the mudstone in the southwestern Taiwan*. (Disaster Prevention Investigation, National Science Council, Taipei, Taiwan, 1984).
- Higuchi, K., Chigira, M. & Lee, D. High rates of erosion and rapid weathering in a Plio-Pleistocene mudstone badland, Taiwan. *Catena* **106**, 68–82 (2013).
- Becel, C., Vercambre, G. & Pages, L. Soil penetration resistance, a suitable soil property to account for variations in root elongation and branching. *Plant and Soil* **353**, 169–180, doi: 10.1007/s11104-011-1020-7 (2012).
- Hseu, Z., Jien, S., Chien, W. & Liou, R. Impacts of biochar on physical properties and erosion potential of a mudstone slopeland soil. *The Scientific World Journal*, doi: 10.1155/2014/602197 (2014).
- Sharma, B. D., Hore, D. K., Pandey, G. & Wadhwa, B. M. Genetic resources of bamboos in the NE region of India. *Indian Journal of Forestry* **15**, 44–51 (1992).
- Behari, B., Agarwal, R., Singh, A. & Banerjee, S. Vegetation development in a degraded area under bamboo based agro-forestry system. *Indian Forester* **126**, 701–720 (2000).
- Beyer, L. Soil microbial biomass and organic matter composition in soils under cultivation. *Biology and Fertility of Soils* **19**, 197–202, doi: 10.1007/bf00336159 (1995).
- Gerzabek, M., Danneberg, O. & Kandeler, E. In *Method in soil biology* (eds F. Schinner, R. Ohlinger, E. Kandeler & R. Margesin) 116–119 (Springers 1996).
- Nath, A. J. & Das, A. K. Carbon pool and sequestration potential of village bamboos in the agroforestry system of northeast India. *Tropical Ecology* **53**, 287–293 (2012).
- Wang, J. *et al.* Estimating aboveground biomass and carbon sequestration of Moso bamboo grown under selection cutting after two years. *Quarterly Journal of Forest Research* **32**, 35–44 (2010).
- Reeves, D. W. The role of soil organic matter in maintaining soil quality in continuous cropping systems. *Soil & Tillage Research* **43**, 131–167, doi: 10.1016/s0167-1987(97)00038-x (1997).
- Manns, H. R. & Berg, A. A. Importance of soil organic carbon on surface soil water content variability among agricultural fields. *Journal of Hydrology* **516**, 297–303, doi: 10.1016/j.jhydrol.2013.11.018 (2014).
- Hudson, B. D. Soil organic-matter and available water capacity. *Journal of Soil and Water Conservation* **49**, 189–194 (1994).
- Huang, C.-Y., Jien, S.-H., Chen, T.-H., Tian, G. & Chiu, C.-Y. Soluble organic C and N and their relationships with soil organic C and N and microbial characteristics in moso bamboo (*Phyllostachys edulis*) plantations along an elevation gradient in Central Taiwan. *Journal of Soils and Sediments* **14**, 1061–1070, doi: 10.1007/s11368-014-0870-z (2014).
- Rhoton, F. E., Shipitalo, M. J. & Lindbo, D. L. Runoff and soil loss from midwestern and southeastern US silt loam soils as affected by tillage practice and soil organic matter content. *Soil & Tillage Research* **66**, 1–11, doi: 10.1016/s0167-1987(02)00005-3 (2002).
- Tapia-Vargas, M., Tiscareno-Lopez, M., Stone, J. J., Oropeza-Mota, J. L. & Velazquez-Valle, M. Tillage system effects on runoff and sediment yield in hillslope agriculture. *Field Crops Research* **69**, 173–182, doi: 10.1016/s0378-4290(00)00139-8 (2001).

18. Tian, G. M. *et al.* Effect of different vegetation systems on soil erosion and soil nutrients in red soil region of southeastern China. *Pedosphere* **13**, 121–128 (2003).
19. Gao, C. *et al.* Nitrogen export from an agriculture watershed in the Taihu Lake area, China. *Environmental Geochemistry and Health* **26**, 199–207, doi: 10.1023/B:EGAH.0000039582.68882.7f (2004).
20. Tu, L.-H. *et al.* Litterfall, litter decomposition, and nutrient dynamics in two subtropical bamboo plantations of China. *Pedosphere* **24**, 84–97 (2014).
21. Yang, L., Zhang, F., Gao, Q., Mao, R. & Liu, X. Impact of land-use types on soil nitrogen net mineralization in the sandstorm and water source area of Beijing, China. *Catena* **82**, 15–22, doi: 10.1016/j.catena.2010.04.004 (2010).
22. Curtin, D. & Campbell, C. A. In *Soil sampling and methods of analysis* (eds M. R. Carter & E. G. Gregorich) 599–606 (CRC, Boca Raton, 2007).
23. Liu, J. *et al.* Seasonal soil CO₂ efflux dynamics after land use change from a natural forest to Moso bamboo plantations in subtropical China. *Forest Ecology and Management* **262**, 1131–1137, doi: 10.1016/j.foreco.2011.06.015 (2011).
24. Kalbitz, K. *et al.* Changes in properties of soil-derived dissolved organic matter induced by biodegradation. *Soil Biology & Biochemistry* **35**, 1129–1142, doi: 10.1016/S0038-0717(03)00165-2 (2003).
25. Manzoni, S., Taylor, P., Richter, A., Porporato, A. & Agren, G. I. Environmental and stoichiometric controls on microbial carbon-use efficiency in soils. *New Phytologist* **196**, 79–91, doi: 10.1111/j.1469-8137.2012.04225.x (2012).
26. Anderson, T. H. & Domsch, K. H. Application of ecophysiological quotients (qCO₂ and qD) on microbial biomass from soils of different crop histories. *Soil Biology & Biochemistry* **22**, 251–255, doi: 10.1016/0038-0717(90)90094-g (1990).
27. Degens, B. P. Decreases in microbial functional diversity do not result in corresponding changes in decomposition under different moisture conditions. *Soil Biology and Biochemistry* **30**, 1989–2000, doi: 10.1016/S0038-0717(98)00071-6 (1998).
28. Anderson, T. H. Microbial eco-physiological indicators to assess soil quality. *Agriculture Ecosystems & Environment* **98**, 285–293, doi: 10.1016/S0167-8809(03)00088-4 (2003).
29. Francaviglia, R. *et al.* Soil quality and vulnerability in a Mediterranean natural ecosystem of Central Italy. *Chemosphere* **55**, 455–466, doi: 10.1016/j.chemosphere.2003.10.066 (2004).
30. Tu, L.-H. *et al.* Decomposition of different litter fractions in a subtropical bamboo ecosystem as affected by experimental nitrogen deposition. *Pedosphere* **21**, 685–695, doi: 10.1016/S1002-0160(11)60171-9 (2011).
31. Watanabe, T., Fukuzawa, K. & Shibata, H. Temporal changes in litterfall, litter decomposition and their chemical composition in Sasa dwarf bamboo in a natural forest ecosystem of northern Japan. *Journal of Forest Research* **18**, 129–138, doi: 10.1007/s10310-011-0330-1 (2013).
32. Wang, H.-C., Tian, G. & Chiu, C.-Y. Invasion of moso bamboo into a Japanese cedar plantation affects the chemical composition and humification of soil organic matter. *Scientific Reports* doi: 10.1038/srep32211 (2016).
33. McLaughlan, K. K. & Hobbie, S. E. Comparison of labile soil organic matter fractionation techniques. *Soil Science Society of America Journal* **68**, doi: 10.2136/sssaj2004.1616 (2004).
34. Wang, Q., Xiao, F., Zhang, F. & Wang, S. Labile soil organic carbon and microbial activity in three subtropical plantations. *Forestry* **86**, 569–574, doi: 10.1093/forestry/cpt024 (2013).
35. Kleber, M. What is recalcitrant soil organic matter? *Environmental Chemistry* **7**, 320–332, doi: 10.1071/EN10006 (2010).
36. Huang, Z., Xu, Z., Chen, C. & Boyd, S. Changes in soil carbon during the establishment of a hardwood plantation in subtropical Australia. *Forest Ecology and Management* **254**, 46–55, doi: 10.1016/j.foreco.2007.07.021 (2008).
37. McLaughlan, K. K. & Hobbie, S. E. Comparison of labile soil organic matter fractionation techniques. *Soil Science Society of America Journal* **68**, 1616–1625 (2004).
38. Nishiyama, M., Sumikawa, Y., Guan, G. & Marumoto, T. Relationship between microbial biomass and extractable organic carbon content in volcanic and non-volcanic ash soil. *Applied Soil Ecology* **17**, 183–187, doi: 10.1016/S0929-1393(01)00131-7 (2001).
39. Xu, J. M., Cheng, H. H., Koskinen, W. C. & Molina, J. A. E. Characterization of potentially bioreactive soil organic carbon and nitrogen by acid hydrolysis. *Nutrient Cycling in Agroecosystems* **49**, 267–271, doi: 10.1023/a:1009763023828 (1997).
40. Soil Survey Staff. *Keys to Soil Taxonomy*. (USDA-Natural Resources Conservation Service, Washington, DC, 2014).
41. Zeng, J.-L. Morning breeze and lingering moon: Zuozhen Caoshan Moon World Photo Album. (Tainan Garvornment, 2000).
42. Rhoades, J. D. Soluble Salts. in *Methods of Soil Analysis, Part 2. Chemical and Microbiological Properties - Agronomy Monograph 9* (eds A. L. Page, R. H. Miller, & D. R. Keeney) 167–178 (Agronomy Society of America and Soil Science Society of America, Madison, WI, USA 1982).
43. Bremner, J. M. & Mulvaney, C. S. Nitrogen - Total in *Methods of Soil Analysis, Part 2. Chemical and Microbiological Properties - Agronomy Monograph 9* (eds A. L. Page, R. H. Miller, & D. R. Keeney) 595–624 (Agronomy Society of America and Soil Science Society of America, Madison, WI, USA, 1982).
44. Scheiner, D. Determination of ammonia and Kjeldahl nitrogen by indophenol method. *Water Research* **10**, 31–36, doi: 10.1016/0043-1354(76)90154-8 (1976).
45. Gee, G. W. & Bauder, J. W. Particle-size Analysis in *Methods of Soil Analysis: Part 1—Physical and Mineralogical Methods SSSA Book Series* (ed Arnold Klute) 383–411 (Soil Science Society of America, American Society of Agronomy, Madison, WI, USA, 1986).
46. Chen, C. R., Xu, Z. H., Zhang, S. L. & Keay, P. Soluble organic nitrogen pools in forest soils of Subtropical Australia. *Plant and Soil* **277**, 285–297, doi: 10.1007/s11104-005-7530-4 (2005).
47. O'Dell, J. W. Determination of nitrate-nitrite nitrogen by automated colorimetry Report No. 600/R-93-100, (U.S. Environmental Protection Agency, Cincinnati, Ohio, USA, 1993).
48. Sollins, P. *et al.* In *Standard Soil Methods for Long-Term Ecological Research* (eds G. P. Robertson, D. C. Coleman, C. S. Bledsoe, & P. Sollins) 89–105 (Oxford University Press, Oxford, UK, 1999).
49. Witt, C., Gaunt, J. L., Galicia, C. C., Ottow, J. C. G. & Neue, H. U. A rapid chloroform-fumigation extraction method for measuring soil microbial biomass carbon and nitrogen in flooded rice soils. *Biology and Fertility of Soils* **30**, 510–519, doi: 10.1007/s003740050030 (2000).
50. Amato, M. & Ladd, J. N. Assay for microbial biomass based on ninhydrin-reactive nitrogen in extracts of fumigated soils. *Soil Biology & Biochemistry* **20**, 107–114, doi: 10.1016/0038-0717(88)90134-4 (1988).
51. Inubushi, K., Brookes, P. C. & Jenkinson, D. S. Soil microbial biomass C, N and ninhydrin-N in aerobic and anaerobic soils measured by the fumigation-extraction method. *Soil Biology & Biochemistry* **23**, 737–741, doi: 10.1016/0038-0717(91)90143-8 (1991).
52. Wu, J., Joergensen, R. G., Pommerening, B., Chaussod, R. & Brookes, P. C. Measurement of soil microbial biomass C by fumigation extraction - An automated procedure. *Soil Biology & Biochemistry* **22**, 1167–1169, doi: 10.1016/0038-0717(90)90046-3 (1990).
53. Gömöryová, E., Štřelcová, K., Škvarenina, J., Bebej, J. & Gömöry, D. In *Bioclimatology and Natural Hazards* (ed Střelcová *et al.*) 251–259 (Springer Science, Berlin, Germany, 2009).
54. Waring, S. A. & Bremner, J. M. Effect of soil mesh-size on estimation of mineralizable nitrogen in soils. *Nature* **202**, 1141–8, doi: 10.1038/2021141a0 (1964).
55. Kalsi, P. S. *Spectroscopy of Organic Compounds*. 6th edn, 16–19 (New Age International Pvt Ltd Publishers, New Delhi, India, 2004).
56. Kumada, K. *Chemistry of soil organic matter*. 214 (Tokyo: Japan Scientific Societies Press, Tokyo, Japan, 1987).
57. Rovira, P. & Vallejo, V. R. Labile and recalcitrant pools of carbon and nitrogen in organic matter decomposing at different depths in soil: an acid hydrolysis approach. *Geoderma* **107**, 109–141, doi: 10.1016/S0016-7061(01)00143-4 (2002).

Acknowledgements

The authors thank the Ministry of Science and Technology of Taiwan, Republic of China, for financially supporting this research under contract number MOST 104-2621-B-001-005 and MOST 105-2621-B-001-007. The authors are also grateful to Ms Pei-Yi Yu and Ms Yun-chien Hsu from the Biodiversity Research Center, Academia Sinica, Taipei, Taiwan, for soil analyses.

Author Contributions

Y.J.S. and H.C.W. performed statistical analyses and built statistical models. T.H.C. initiated the ecological study of bamboo plantations in the badlands and helped to set up the experimental design. SHJ helped to conduct the soil survey and collect soil samples. G.T. helped in analysing and interpreting data. C.Y.C. originally formulated the idea and developed the methodology. Y.J.S., H.C.W., G.T. and C.Y.C. wrote the manuscript. All authors read and approved the final manuscript.

Additional Information

Competing financial interests: The authors declare no competing financial interests.

How to cite this article: Shiau, Y.-J. *et al.* Improvement in the biochemical and chemical properties of badland soils by thorny bamboo. *Sci. Rep.* 7, 40561; doi: 10.1038/srep40561 (2017).

Publisher's note: Springer Nature remains neutral with regard to jurisdictional claims in published maps and institutional affiliations.



This work is licensed under a Creative Commons Attribution 4.0 International License. The images or other third party material in this article are included in the article's Creative Commons license, unless indicated otherwise in the credit line; if the material is not included under the Creative Commons license, users will need to obtain permission from the license holder to reproduce the material. To view a copy of this license, visit <http://creativecommons.org/licenses/by/4.0/>

© The Author(s) 2017