



# Spatio-temporal patterns of genetic variation in *Arbacia lixula*, a thermophilous sea urchin in expansion in the Mediterranean

Rocío Pérez-Portela<sup>1,5</sup> · Owen S. Wangensteen<sup>2</sup> · Alex Garcia-Cisneros<sup>1</sup> · Claudio Valero-Jiménez<sup>3</sup> · Cruz Palacín<sup>4</sup> · Xavier Turon<sup>1</sup>

Received: 29 December 2017 / Revised: 23 May 2018 / Accepted: 24 May 2018 / Published online: 14 June 2018  
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## Abstract

The genetic structure of 13 populations of the amphiatlantic sea urchin *Arbacia lixula*, as well as temporal genetic changes in three of these localities, were assessed using ten hypervariable microsatellite loci. This thermophilous sea urchin is an important engineer species triggering the formation of barren grounds through its grazing activity. Its abundance seems to be increasing in most parts of the Mediterranean, probably favoured by warming conditions. Significant genetic differentiation was found both spatially and temporally. The main break corresponded to the separation of western Atlantic populations from those in eastern Atlantic and the Mediterranean Sea. A less marked, but significant differentiation was also found between Macaronesia (eastern Atlantic) and the Mediterranean. In the latter area, a signal of differentiation between the transitional area (Alboran Sea) and the rest of the Mediterranean was detected. However, no genetic structure is found within the Mediterranean (excluding Alboran) across the Siculo-Tunisian Strait, resulting from either enough gene flow to homogenize distance areas or/and a recent evolutionary history marked by demographic expansion in this basin. Genetic temporal variation at the Alboran Sea is as important as spatial variation, suggesting that temporal changes in hydrological features can affect the genetic composition of the populations. A picture of genetic homogeneity in the Mediterranean emerges, implying that the potential expansion of this keystone species will not be limited by intraspecific genetic features and/or potential impact of postulated barriers to gene flow in the region.

**Electronic supplementary material** The online version of this article (<https://doi.org/10.1038/s41437-018-0098-6>) contains supplementary material, which is available to authorized users.

✉ Rocío Pérez-Portela  
perezportela@gmail.com

- <sup>1</sup> Center for Advanced Studies of Blanes (CEAB, CSIC), Accés a la Cala Sant Francesc 14, Blanes, Girona, Spain
- <sup>2</sup> Norwegian College of Fishery Science, UiT The Arctic University of Norway, Tromsø, Norway
- <sup>3</sup> Laboratory of Phytopathology, Wageningen University, 6708 PB Wageningen, The Netherlands
- <sup>4</sup> Department of Evolutionary Biology, Ecology and Environmental Sciences, University of Barcelona and Research Institute of Biodiversity (IRBIO), Avda. Diagonal 643, Barcelona, Spain
- <sup>5</sup> Present address: Department of Biology, Geology, Physics and Inorganic Chemistry, Rey Juan Carlos University, C/ Tulipán s/n, 28932 Móstoles, Spain

## Introduction

*Arbacia lixula* (Linnaeus 1758) is a warm-temperate water species occurring from the western Atlantic in Brazil (Tommasi 1964) to the other side of the Atlantic where it is present in the Macaronesian archipelagos (Mortensen 1935; Lessios et al. 2012), African Atlantic coast from Gibraltar to Angola, and the Mediterranean Sea (Tortonese 1965). Marine species with amphiatlantic distributions (i.e., those inhabiting both eastern and western Atlantic shorelines) provide interesting tests of the permeability of the mid-Atlantic dispersal barrier. Barring cases of cryptic speciation (e.g. Carmona et al. 2011), historical, hydrological and developmental features are usually called for to explain trans-Atlantic dispersal. In this sense, *Arbacia* is an interesting genus with fossil record dating from the Palaeocene (Kroh and Smith 2010). Its five extant species occur in the eastern Pacific and both sides of the Atlantic (Lessios et al. 2012). The two Atlantic species, *A. punctulata* (western

Atlantic) and *A. lixula* (amphiatlantic) diverged some 1.5–3.3 mya at both sides of the mid-Atlantic barrier (Lessios et al. 2012), likely by a range expansion event from western to eastern Atlantic of the lineage that would become *A. lixula*, which nevertheless crossed again the mid-Atlantic barrier to establish the present-day Brazilian populations (Lessios et al. 2012; Wangensteen et al. 2012).

*Arbacia lixula* is an ecosystem engineer species (i.e., those that change availability of resources to other species; Jones et al. 1994, 1997), capable of transforming littoral communities into barren grounds due to its grazing activity (Bulleri et al. 1999; Gianguzza et al. 2011; Bonaviri et al. 2011). Mitochondrial genetic data (Wangensteen et al. 2012) and the absence of fossil records (Stefanini 1911; Mortensen 1935; Madeira et al. 2011) support the idea of a relatively recent colonization of this sea urchin in the Mediterranean Sea, likely during the last interglacial period (Wangensteen et al. 2012). The Mediterranean is a semi-enclosed sea subject to important anthropogenic impacts (e.g. Lejeune et al. 2010; Coll et al. 2010). In turn, these threats interact in complex ways with the ongoing climate change that favours the progressive tropicalization of this sea (Francour et al. 1994). Among the key drivers of structure and function in littoral Mediterranean communities is the grazing activity of sea urchins, which induce regime shifts between macroalgal beds and sea urchin barrens (Bonaviri et al. 2011). Human-derived impacts can exacerbate the risk and irreversibility of such dramatic changes (Ling et al. 2014).

The thermophilous nature of *A. lixula* has long been recognized (Kempf 1962; Tortonese 1965), and this species is listed among those being currently favoured by the warming of the Mediterranean (Wangensteen 2013). Its abundance has been increasing in several areas of this sea in the past (Petit et al. 1950; Boudouresque et al. 1989; Francour et al. 1994). Its reproduction is enhanced by high temperatures (Gianguzza et al. 2011, Wangensteen et al. 2013b) and larval development features indicate that warming, modulated by other factors such as pH and food availability, may favour *A. lixula* development (Privitera et al. 2011; Wangensteen et al. 2013a; Gianguzza et al. 2014; Visconti et al. 2017). Although recent results showed a regression of marine invertebrate populations at the coast of Israel (eastern Mediterranean) due to the whole ecosystem collapsing (Yerulam et al. 2015; Rilov 2016), the general scenario is a progressive increase of abundance of *A. lixula* in most areas of the Mediterranean (Privitera et al. 2011; Wangensteen et al. 2013a; Visconti et al. 2017), which will result in significant changes in ecosystem functioning.

Under this scenario, it is of utmost importance to ascertain the genetic structure of *A. lixula*. In a previous study, Wangensteen et al. (2012) identified

phylogeographic patterns in *A. lixula* using sequences of the mitochondrial gene cytochrome oxidase I (COI). That study identified three haplogroups in worldwide populations, one of them shared between eastern and western Atlantic populations. The mitochondrial structure of the species appeared to be shaped by Pleistocene demographic expansions, isolation between the eastern Atlantic, western Atlantic and Mediterranean Sea, and genetic homogeneity across the Mediterranean. Nevertheless, the lack of genetic differentiation across the Mediterranean basin (Wangensteen et al. 2012; Deli et al. 2017) needs to be compared with nuclear markers to confirm the information on gene-flow patterns and genetic signals in this species. Mitochondrial DNA only retains half of the species' evolutionary history (Avice 2000), and due to the potential differential selection (Silva et al. 2014; Consuegra et al. 2015) and stochasticity of the coalescence processes between nuclear and mitochondrial DNA, these two types of markers can show different evolutionary signatures (e.g. Glynn et al. 2015; Garcia-Cisneros et al. 2016; Pérez-Portela et al. 2017). Therefore, combining both mitochondrial and nuclear information should provide complementary information to unravel both recent and historical processes shaping the genetic structure of *A. lixula*.

Population analyses should additionally include information about temporal changes in genetic make-up to understand whether the structure observed is stable over contemporary time periods. Currently, there is still a scarce number of temporal genetic studies in marine species, despite being a fundamental information for interpreting their long-term genetic distribution (e.g. Pérez-Portela et al. 2012; Pineda et al. 2016; Pascual et al. 2016). It is known that the stochasticity of reproduction, recruitment and survival of larvae and juveniles can potentially change the genetic composition of populations over the generations (e.g. Calderón et al. 2012; Aglieri et al. 2014; Couvray and Coupé 2018). Additionally, temporal variation across oceanographic discontinuities can also promote variation of gene-flow patterns over time (Olivar et al. 2003; Calderón et al. 2012). An outstanding example of interannual oceanographic variation is that across the Atlantic-Mediterranean transition, associated with shifts in Atlantic and Mediterranean water contributions across the Alboran Sea (Renault et al. 2012; Oguz et al. 2014). These marine circulation variations determine different levels of genetic mixing between Atlantic and Mediterranean genetic stocks over the years (Pascual et al. 2016). Therefore, spatio-temporal structuring patterns can provide valuable information about the future evolution of the populations, identifying connectivity patterns over time, and reservoirs of genetic diversity, among other important features.

In the present work, we use hypervariable nuclear microsatellite loci to investigate in detail the genetic

structure of *A. lixula* across most of its distribution range using the same samples analysed by Wangensteen et al. (2012), but also extending these analyses to a temporal perspective. With new nuclear markers and samples, we specifically tested: (a) the disruptive effect of major oceanographic breaks, including the mid-Atlantic barrier, as well as migration patterns across them, which were used to determine the coherence of genetic divergence patterns between the nuclear and mitochondrial data, and (b) the relevance of the genetic change over time in two sites at the Alboran Sea (Atlantic-Mediterranean transition) and in another non-transitional Mediterranean site, which were sampled at two time points. We were particularly interested in inferring spatio-temporal population structure at the Atlanto-Mediterranean transition where other marine invertebrates have shown significant interannual variation in genetic structure (Pascual et al. 2016). The data generated in this study can be useful to infer present-day and future processes in the ongoing expansion of this keystone engineer species.

## Materials and methods

### Sample collection and microsatellite genotyping

#### Spatial genetic structure

The collection sites included 13 different localities across most of the distribution range of the species: two localities on the western Atlantic (Brazil), three sites on the eastern Atlantic: Cape Verde, Canary Islands and Azores (Macaronesian Islands), five at the western Mediterranean (including two populations from the transitional zone at the Alboran Sea), and three in the eastern Mediterranean (see details in Fig. 1 and Table 1). Samples analysed correspond to a subset of 278 out of 604 individuals previously sequenced (mitochondrial COI gene) by Wangensteen et al. (2012) between 2009 and 2011, with an additional location from Sicily collected by SCUBA diving for the present study at the end of 2011. This sampling scheme included several major oceanographic breaks and/or transitions with observed disruptive effect in populations of other echinoderms (e.g. Calderón et al. 2008; Pérez-Portela et al. 2013, 2017; Taboada and Pérez-Portela 2016; Garcia-Cisneros et al. 2016, 2017): the mid-Atlantic barrier that divides the eastern and western Atlantic; the Gibraltar Strait that marks the geographical partition between the Atlantic Ocean and Mediterranean Sea; the Almeria-Oran front, described as the biogeographical break between the Atlantic and Mediterranean basins in most marine species; and the Siculo-Tunisian Strait between the eastern and western Mediterranean sub-basins.

### Temporal genetic trends

For testing potential changes in genetic structure and diversity over time, three of the Mediterranean populations sampled in 2009 were re-sampled for this study in 2014: Colera at the northwestern Mediterranean, and La Herradura and Torremuelle at the Alboran Sea–Atlantic–Mediterranean transition. These sites were selected because we were specifically interested in exploring the potential effect of interannual oceanographic variation on populations' divergence at the Atlantic–Mediterranean transition, an area where *A. lixula* populations displayed significant mitochondrial differences (Wangensteen et al. 2012) despite the short geographical distances separating them to other Atlantic and Mediterranean sites. We analysed the two Alboran sites for which samples from 2009 were available (Wangensteen et al. 2012) and one northwestern Mediterranean site far away from this Atlantic–Mediterranean transition for comparison with the first two sites.

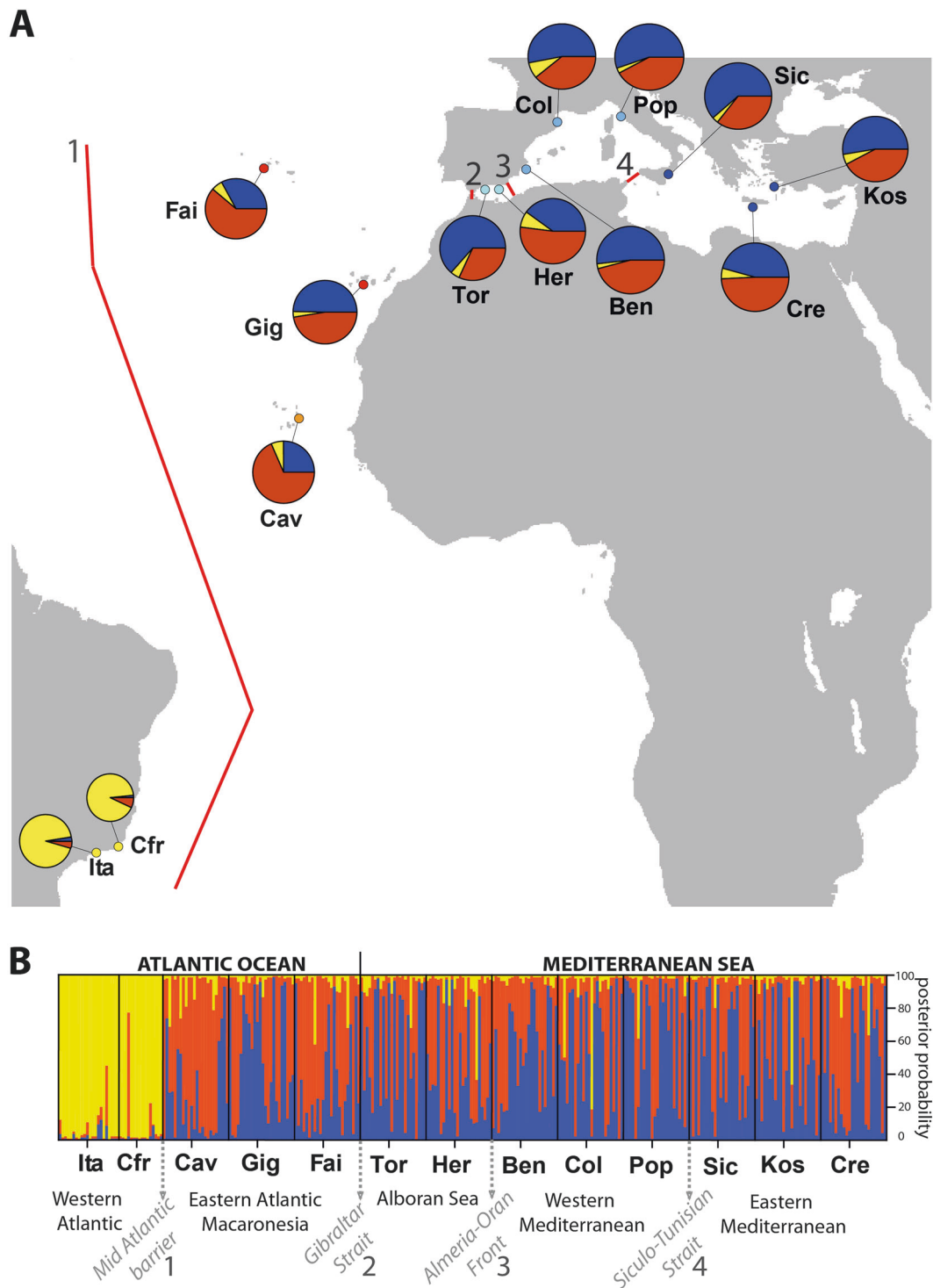
Tissue samples were collected and fixed as described in Wangensteen et al. 2012. Total DNA was extracted from 302 individuals for the “spatial” study, plus 77 individuals of the 2014 sampling used for the “temporal” study. The REDEExtract-N-Amp Tissue PCR kit (from Sigma-Aldrich, [www.sigmaaldrich.com/](http://www.sigmaaldrich.com/)) was used, following the protocol described by the manufacturer. All individuals were genotyped at ten microsatellite loci (ALM2, ALM4, ALM5, ALM7, ALM8, ALM9, ALM11, ALM14, ALM15 and ALM17) described in Garcia-Cisneros et al. (2013).

Amplification of fragments containing microsatellites was performed by polymerase chain reaction (PCR) in a final volume of 10 µl, containing 5 µl of ReadyMix Taq PCR Reaction Mix (Sigma-Aldrich), 2–8 ng of DNA, 0.4 µl (10 µM) of each primer (forward and reverse) and 3.2 µl of ultrapure water. Samples were amplified in a thermocycler (Bio-Rad MyCycler, <http://www.bio-rad.com>) with an initial 2 min denaturation step at 94 °C, and 35 amplification cycles: 45 s at 94 °C, 50 s at the locus-specific annealing temperature (51–58 °C; see Garcia-Cisneros et al. 2013) and 40 s at 72 °C, followed by 4 min of final extension at 72 °C.

Successful amplifications were genotyped in an automated sequencer (Applied Biosystems, [www.thermofisher.com](http://www.thermofisher.com)) in the Science and Technology Centres of the University of Barcelona (CCiTUB). Allele length was estimated relative to the internal size standard 70–500 ROX (Bioventures) using the software Peak-Scanner v 1.0 (Applied Biosystems).

### Data analyses

The number of alleles per population, observed heterozygosity ( $H_o$ ), expected heterozygosity ( $H_e$ ), inbreeding coefficients ( $F_{IS}$ ), and number of private alleles per



**Fig. 1** Collection sites and general genetic structure in *Arbacia lixula*. **a** Sampling sites and pie charts of the percentage of each cluster ( $K = 3$ ) per site obtained from STRUCTURE as represented below, and **b** STRUCTURE barplot with the posterior probabilities of individual assignment of the most probable number of clusters for the whole data

set ( $K = 3$ ). Clusters are represented by different colours (yellow, red and blue), and each bar represents a different individual. Numbers and red lines represent the four different marine barriers considered: Mid-Atlantic barrier (1), Gibraltar Strait (2), Almeria-Oran front (3) and Siculo-Tunisian Strait (4)

geographical area were calculated using GenAlix v 6.41 (Peakall and Smouse 2006) and Genepop v 4.2 webserver (Raymond and Rousset 1995). The exact test for departure

from Hardy–Weinberg equilibrium (HWE) was performed in Arlequin v 3.5.1.2 (Excoffier et al. 2005). The potential correlation between the  $F_{IS}$  and number of missing data per

**Table 1** Genetic descriptors of *Arbacia lixula*

| Code     | Locality               | Geographical area                     | <i>N</i> | <i>H<sub>o</sub></i> | <i>H<sub>e</sub></i> | <i>F<sub>IS</sub></i> | Allele         |
|----------|------------------------|---------------------------------------|----------|----------------------|----------------------|-----------------------|----------------|
| Cre      | Crete                  | Eastern Mediterranean                 | 24       | 0.565 ± 0.195        | 0.849 ± 0.076        | 0.323 ± 0.07 *        | 13.2 ± 1.23    |
| Kos      | Kos                    | Eastern Mediterranean                 | 24       | 0.597 ± 0.236        | 0.854 ± 0.059        | 0.289 ± 0.08*         | 12.6 ± 1.26    |
| Sic      | Sicilia                | Eastern Mediterranean                 | 24       | 0.568 ± 0.213        | 0.879 ± 0.051        | 0.343 ± 0.08 *        | 12.8 ± 1.39    |
| Pop      | Populonia              | Western Mediterranean                 | 24       | 0.522 ± 0.222        | 0.842 ± 0.059        | 0.365 ± 0.09 *        | 11.2 ± 0.96    |
| Col      | Colera (2009)          | Western Mediterranean                 | 24       | 0.562 ± 0.198        | 0.878 ± 0.029        | 0.347 ± 0.07 *        | 13.4 ± 1.02    |
| Ben      | Benidorm               | Western Mediterranean                 | 24       | 0.563 ± 0.215        | 0.864 ± 0.058        | 0.331 ± 0.08 *        | 12.1 ± 1.07    |
| Her      | La Herradura (2009)    | Alboran Sea                           | 24       | 0.552 ± 0.134        | 0.819 ± 0.103        | 0.300 ± 0.06 *        | 11.9 ± 1.06    |
| Tor      | Torremuelle (2009)     | Alboran Sea                           | 24       | 0.571 ± 0.205        | 0.845 ± 0.065        | 0.300 ± 0.09 *        | 11.8 ± 1.20    |
|          |                        | MEDITERRANEAN                         | 192      | 0.563 ± 0.090        | 0.863 ± 0.052        | –                     | (10.94 ± 0.80) |
| Fai      | Faials, Azores         | Macaronesian Island, eastern Atlantic | 24       | 0.619 ± 0.202        | 0.834 ± 0.064        | 0.241 ± 0.08 *        | 11.7 ± 1.01    |
| Gig      | Los Gigantes, Tenerife | Macaronesian Island, eastern Atlantic | 24       | 0.569 ± 0.232        | 0.852 ± 0.051        | 0.289 ± 0.08 *        | 13.1 ± 1.26    |
| Cav      | Boavista, Cape Verde   | Macaronesian Island, eastern Atlantic | 24       | 0.585 ± 0.222        | 0.872 ± 0.057        | 0.318 ± 0.08 *        | 14.3 ± 1.28    |
|          |                        | EASTERN ATLANTIC                      | 72       | 0.591 ± 0.200        | 0.856 ± 0.051        | –                     | (11.10 ± 0.71) |
| Cfr      | Cabo Frío              | Brazil, western Atlantic              | 16       | 0.625 ± 0.282        | 0.804 ± 0.122        | 0.201 ± 0.11 *        | 9.3 ± 1.06     |
| Ita      | Itaipu                 | Brazil, western Atlantic              | 22       | 0.579 ± 0.209        | 0.819 ± 0.113        | 0.253 ± 0.06 *        | 11.0 ± 1.12    |
|          |                        | WESTERN ATLANTIC                      | 38       | 0.593 ± 0.205        | 0.803 ± 0.100        | –                     | (9.76 ± 0.91)  |
| Col 2014 | Colera (2014)          | Western Mediterranean                 | 25       | 0.625 ± 0.082        | 0.869 ± 0.015        | 0.267 ± 0.09 *        | 13.0 ± 0.91    |
| Her 2014 | La Herradura (2014)    | Alboran Sea                           | 27       | 0.588 ± 0.083        | 0.858 ± 0.017        | 0.298 ± 0.09 *        | 12.6 ± 1.18    |
| Tor 2014 | Torremuelle (2014)     | Alboran Sea                           | 25       | 0.627 ± 0.069        | 0.861 ± 0.014        | 0.255 ± 0.08 *        | 12.7 ± 1.02    |

Codes of the populations including the year of collection for the populations used for the temporal analyses, localities, geographical area, sample size (*N*), heterozygosity observed (*H<sub>o</sub>*), heterozygosity expected (*H<sub>e</sub>*), inbreeding coefficient (*F<sub>IS</sub>*), and allele diversity—mean number of alleles per population or allele richness per marine geographical area in brackets (calculated by rarefaction from the minimum sample size). Standard error is also presented for genetic descriptors

\*Significant HW disequilibrium ( $p < 0.01$ )

population was explored to understand the impact of missing data on this statistic.

### Spatial genetic structure

We used different approaches based on Bayesian clustering, genetic distances and discriminant analyses of principal components. Whereas methods based on genetic distances (e.g. *F<sub>ST</sub>*) are affected by the populations' Hardy–Weinberg disequilibrium, and assume absence of linkage disequilibrium among all loci within populations, other multivariate methods are free from these two assumptions. Therefore, we compared here different methods to minimize potential bias of using only one approach.

The software STRUCTURE v 2.3.4 (Pritchard et al. 2000) was used to infer an optimal number of homogeneous genetic units (*K*) based on Bayesian clustering analyses. It was run with the whole data set, with a *K* number from 1 to 16, and 200,000 Markov chain Monte Carlo (MCMC) steps were performed following 80,000 burn-in iterations in ten independent replicates under the “admixture model” and the “correlated allele frequencies mode” implemented by the software. The same strategy was separately applied to selected subsets of the populations in order to obtain a finer-

scale analysis within major marine areas: (a) the eastern Atlantic and Mediterranean populations to better explore genetic partition across the Atlantic–Mediterranean arch and, (b) only Mediterranean sites to investigate potential divergence within this basin and across the Almeria-Oran Front and the Siculo-Tunisian Strait. The most likely value of “real” clusters was identified comparing the rate of change in the likelihood of *K*. The optimal *K* values were determined using the ad hoc statistic  $\Delta K$  (Evanno et al. 2005). Ten independent replicates per run were averaged using the clumpak server (Kopelman et al. 2015), and results were graphically represented with the same software.

Genetic clusters were also delineated using “find.clusters” of the adegenet package for R software (Jombart 2008; Team R Core 2013) using a K-means clustering algorithm. A range of cluster numbers was chosen and the optimal number was selected using a Bayesian Information Criterion (BIC). Group assignment probabilities were then displayed with the “compplot” function of adegenet. As before, further analyses were performed with “find.clusters” considering only eastern Atlantic and Mediterranean populations and, finally, only Mediterranean populations. Additionally, we ran a discriminant analysis of principal components (DAPC, Jombart et al. 2010) using populations

as groups with the adegenet package. This method allows the visual identification of genetic clusters of individuals and can outperform Bayesian clustering approaches in detecting genetic substructure (Jombart et al. 2010). The optimal number of principal components (PC) from the PCA step passed onto the discriminant analysis was determined by the cross-validation method, and by comparison of a-scores for a set of increasing numbers of PCs and a spline interpolation using the “a-score” function of adegenet. DAPCs were performed separately for the whole data set, for the eastern Atlantic plus Mediterranean populations, and for the Mediterranean populations alone.

The software Arlequin was used to estimate population distances with the  $F_{ST}$  statistic between pairs of populations based on an allele infinite model. The Jost's  $D_{est}$  estimator (Jost 2008) was also obtained with the package DEMETics in R (Gerlach et al. 2010). A false discovery rate correction was applied for the  $p$  values (Benjamini–Yekutieli method, Narum 2006) to account for multiple tests. The genetic dissimilarity matrices generated with both estimators were represented with cluster analyses and heatmaps obtained with the gplots package for R (Warnes et al. 2016).

To test the concordance between nuclear and mitochondrial genetic distances, we performed correlation analyses for  $F_{ST}$  and  $D_{est}$  matrixes obtained from microsatellite loci (this study) and COI sequences (COI distance matrixes obtained from Wangenstein et al. 2012).

Null allele frequencies were estimated following the Expectation Maximization (EM) algorithm implemented in FreeNA (Chapuis and Estoup 2007). Using this information, the corrected estimations of  $F_{ST}$  values were calculated applying the *ENA* and *INA* methods with the same software.

Analyses of molecular variance (AMOVA) were computed using an allele infinite model, and their significance tested with 20,000 permutations in Arlequin. For the AMOVAs we grouped populations in different sets according to the  $F_{ST}$  results, geographical origin and known oceanographic barriers. We initially tested differences among western Atlantic, eastern Atlantic and Mediterranean Sea, considering two major marine breaks: the mid-Atlantic barrier and the Gibraltar Strait. In a second analysis we removed populations from western Atlantic and compared east Atlantic populations with Mediterranean populations. We then compared the populations from the Alboran Sea with the rest of the Mediterranean to test differentiation across the Almeria-Oran front. Finally, we analysed only Mediterranean populations excluding Alboran Sea, comparing the eastern and the western sub-basins to explore the potential disruptive effect of the Siculo-Tunisian Strait.

The potential effect of genetic isolation of populations by geographical distance, independently of oceanographic barriers, was assessed for the whole data set, and separately for different population subsets (eastern Atlantic and

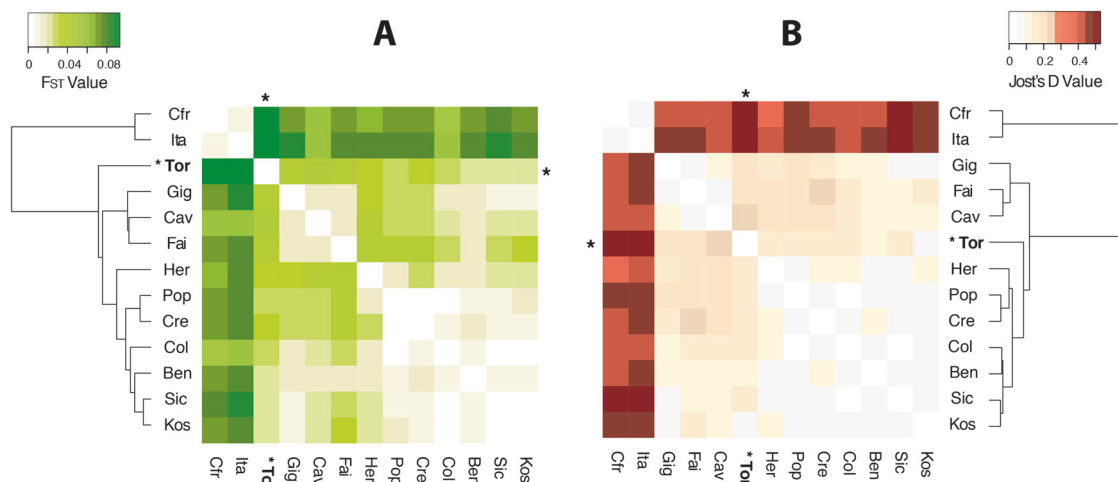
Mediterranean Sea, and only the Mediterranean Sea), using the correlation of linearized genetic distances ( $F_{ST}/1 - F_{ST}$ ) with geographical distances (as measured in Wangenstein et al. 2012) between localities. The significance of the correlations was tested by a Mantel test, as implemented in Arlequin with 20,000 permutations per analysis.

To estimate gene flow among marine areas, we calculated mutation-scaled effective migration rates ( $M$ ) based on Bayesian inference using the software MIGRATE v 3.6.11 (Beerli 2006; Beerli and Felsenstein 2001). We estimated asymmetric  $M$  among the three major geographical areas: the western Atlantic (Brazilian sites), eastern Atlantic (Macaronesian islands) and the Mediterranean Sea. Migration estimates per generation can be expressed as  $4Nm$  for nuclear markers, in which  $N$  is the effective population size and  $m$  the immigration rate. Three preliminary runs were performed to infer initial parameters and check convergence before performing a final run. For the latter, we used a Brownian motion mutation model with constant mutation rate for all loci, three different replicates with one long chain, 3,000,000 iterations (9,000,000 final sampled parameters) with the first 30,000 iterations discarded, and an adaptive heating scheme of four different temperature chains.

### Temporal genetic trends

For the three populations sampled in 2009 and again in 2014 (Colera, Torremuelle and La Herradura), we computed a DAPC representation using populations from each sampling year as groups (with the adegenet package in R) and pairwise tests using  $F_{ST}$  (calculated with Arlequin) and  $D_{est}$  (calculated with DEMETics) as described above.

We also estimated effective population sizes ( $N_e$ ) for these three populations (Colera, Torremuelle and La Herradura) using the temporal method, based in shifts in allele frequencies between samples taken a number of generations apart (Jorde and Ryman 2007). We used NeEstimator v.2.01 (Do et al. 2014) to calculate  $N_e$  based on allele frequency changes between the two sampling years using three different estimators that differ in precision and bias (Do et al. 2014): those of Nei and Tajima (1981), Pollak (1983), and Jorde and Ryman (2007). We considered a generation per year (Wangenstein et al. 2013b) and removed alleles below a frequency threshold of 0.05 to reduce random error (likely at the cost of a slight downward bias in the estimates, Do et al. 2014). *Arbacia lixula* has overlapping generations, which adds complexity to the computation of  $N_e$  estimates originally devised for discrete generations. Ideally, a correction should be made on measures of temporal change in allele frequency that incorporates the different contributions of the co-existing cohorts (Jorde and Ryman 1995). Calculating this correction



**Fig. 2** Genetic differentiation between sites in *Arbacia lixula*. **a** Heatmap and dendrogram based on pairwise  $F_{ST}$  values, and **b** heatmap and dendrogram based on pairwise  $D_{est}$  values. Values of  $F_{ST}$  and

$D_{est}$  and their associated  $p$  values are included as Supplementary Table S2. \*Highlights the different clustering of Torremuelle (Tor) with other marine areas between heatmaps

requires precise biological knowledge of the cohort structure, age-specific survival rates, and age-specific reproduction rates (e.g., Calderón et al. 2009), parameters that were not available for *A. lixula*. We nevertheless applied temporal methods without correction as, first, we sampled the sea urchins randomly with respect to age and, second, we sampled at a wide interval of generations (five generations apart, from 2009 to 2014). Jorde and Ryman (1995) showed how sampling over long time intervals greatly reduces the bias in temporal methods for overlapping generations. In any case, our estimates should still be useful for comparative purposes among populations, as biological parameters are unlikely to be very different between populations and, therefore, any remaining bias should be similar.

## Results

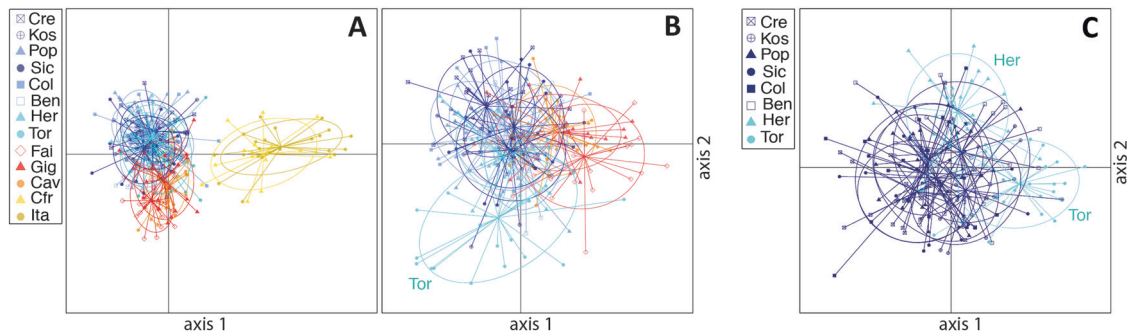
The ten microsatellite loci were highly polymorphic, with a total number of alleles ranging between 16 (locus ALM11) and 38 (locus ALM4). Details of genetic descriptors for each locus and population are presented as supplementary material (Table S1). Populations of *A. lixula* were in general characterized by high genetic diversity and a large number of alleles (mean number per locus ranged from 9.3 to 14.3 alleles, Table 1). Allele richness, used to compare allelic diversity among marine areas with large differences in sample size, showed that the eastern Atlantic retained the highest richness, followed by the Mediterranean and the western Atlantic areas. Regarding private alleles, the eastern Atlantic showed the lowest value, with only 6.77% (13 alleles) of private alleles, whereas the Mediterranean and western Atlantic had 14.2% (31 alleles) and 10.7% (14

alleles) of private alleles, respectively (Supplementary Fig. S1).

In all populations observed heterozygosity was lower than expected, as demonstrated by the significant values of the  $F_{IS}$ , with significant deviation from the HWE in all populations ( $p < 0.001$ ) (see Table 1). All microsatellite loci considered individually had overall positive values of  $F_{IS}$ , significant in all cases ( $F_{IS}$  values  $> 0.11$ ) except in the locus ALM2 ( $F_{IS} = 0.021$ ,  $p = 0.157$ ). A low overall percentage of missing data (2.25%), distributed across all microsatellites but mostly concentrated in the Brazilian populations, makes unlikely that null alleles underlie this general deficit of heterozygotes. Interestingly, the two populations showing the highest percentage of missing data also displayed the lowest  $F_{IS}$  values, also suggesting that missing data are not related to positive and significant  $F_{IS}$  (Supplementary Fig. S2).

The Bayesian analyses detected an optimal  $K$  value of 3 based on the  $\Delta K$  plot (Supplementary Fig. S3). The composition of the different populations in terms of these three genetic groups (sum of individual membership probabilities to each group) is represented in form of pie charts in Fig. 1a. One of the three genetic clusters detected sharply separated the populations from the western Atlantic (yellow group in Fig. 1), while the rest of populations were mainly composed of the other two genetic clusters. In most individuals, however, the most probable group had a membership probability above 75%, with few admixed individuals (Fig. 1b).

The situation is similar when genetic groups are delineated using the “find.clusters” function in adegenet. The number of clusters (BIC criterion) that better explains our data is 6 (Fig. S4), but the plot of membership probabilities



**Fig. 3** DAPCs of *Arbacia lixula*. Graphs represent DAPC results from three different analyses: **a** the whole data set, **b** eastern Atlantic and Mediterranean sites, and **c** only Mediterranean sites (including the Alboran Sea). On the graph, points represent different individuals, and

point patterns different sampling sites. Blue pattern = Mediterranean sites; red-orange pattern = eastern Atlantic; and yellow pattern = western Atlantic

shows clear differences between western Atlantic and all other Atlantic and Mediterranean samples, and some differentiation between the eastern Atlantic (Macaronesia) and Mediterranean based on group membership (Fig. S4). Hence, both the Bayesian clustering analysis and “find clusters” function detected a strong disruptive effect of the mid-Atlantic barrier and a smaller effect of the eastern Atlantic (Macaronesia)–Mediterranean transition. Analyses performed separately for the different marine areas, the whole data set, eastern Atlantic and Mediterranean and only Mediterranean Sea, did not provide additional information (results not shown).

Results from FreeNA showed that, in most cases, the correction of  $F_{ST}$  values was minimal and the significance of the  $F_{ST}$  statistic did not change in any case. Therefore, we consider that null alleles do not have a large effect on genetic distance estimations in this study, and that uncorrected values can be used for further analyses. The values of population differentiation using  $F_{ST}$  and  $D_{est}$  estimators are shown in Table S2 and graphically depicted as dendrograms and heatmaps in Fig. 2. Both estimators provide basically the same information, and are highly correlated ( $r = 0.979$ ,  $p < 0.001$ ). Moreover, they are highly correlated with previous genetic distance results from mitochondrial DNA obtained from Wangenstein et al. (2012) ( $r = 0.832$  and  $r = 0.816$ ,  $p < 0.01$  for  $F_{ST}$  and  $D_{est}$  values, respectively), showing congruent results between microsatellites and COI. Pairwise comparisons using microsatellite loci showed significant differentiation in all comparisons involving Brazilian populations (western Atlantic) with the rest, indicating a strong disruptive effect of the mid-Atlantic barrier. In addition, 17 comparisons (out of 24) between eastern Atlantic (Macaronesian) and Mediterranean populations were significant with both estimators, and 6 comparisons (out of 12) of the Alboran Sea populations (La Herradura and Torremuelle) with the rest of the Mediterranean were also significant for both indices, suggesting limited gene flow across two additional marine barriers: the

Gibraltar Strait and the Almeria-Oran front. Furthermore, the two sites from the Alboran Sea, the transition area between the eastern Atlantic and Mediterranean Sea, were significantly different from each other for both indices. Only one significant pairwise difference within Macaronesia was found between Los Gigantes (Gig—Canary Islands) and Boavista (Cav—Cape Verde Islands) with Jost’s estimator ( $D_{est}$ ). No significant divergence was found in any comparison within the western Atlantic. Within the Mediterranean Sea, no significant divergence was detected between sites, discarding the Siculo-Tunisian Strait as a genetic barrier in this species.

The heatmaps and dendrograms show clearly the distinction between western Atlantic populations and the remaining ones. Among the latter, the Macaronesian populations (eastern Atlantic- Faials, Los Gigantes and Boavista) formed a cluster, while Mediterranean populations appeared well mixed, with no inter-basin structure, although Alboran Sea populations (Torremuelle and La Herradura) were in general slightly more differentiated. In particular, the Torremuelle population was somewhat more divergent and was separated from the rest of Mediterranean populations ( $D_{est}$ ) or even external to the eastern Atlantic plus Mediterranean clusters with the  $F_{ST}$  estimator (Fig. 2).

The spatial representation of the DAPC considering all populations (Fig. 3a, 51 PCs retained) showed again this pattern of separation between western Atlantic and eastern Atlantic plus Mediterranean in the first axis, while along the second axis the populations of the Macaronesian archipelagos are separated, albeit with some overlap, from the Mediterranean populations.

A DAPC graph excluding the Brazilian populations (Fig. 3b, 28 PCs retained) also showed a separation of the Macaronesian populations along the first axis, with overlap of the inertia ellipses. Torremuelle appeared also partially separated from the rest on the second axis. Finally, a DAPC considering only the Mediterranean populations (Fig. 3c, 26 PCs retained) showed less differentiation than the previous



**Table 2** Analyses of the molecular variance (AMOVA) in *A. lixula*

| Source of variation                                | d.f. | Fixation index   | % variation | <i>P</i> value | Grouping  |
|--|------|------------------|-------------|----------------|---|
| <b>Brazil, Macaronesia, Mediterranean</b>          |      |                  |             |                |   |
| Among group (3 groups, 13 populations)             | 2    | $F_{CT}$ : 0.033 | 3.30        | 0.000 **       | 1. Ita, Cfr<br>2. Fai, Gig, Cav                               |
| Among populations within groups                    | 10   | $F_{SC}$ : 0.008 | 0.80        | 0.000 **       | 3. Ben, Col, Pop, Sic, Cre, Kos, Tor, Her                     |
| Within populations                                 | 298  | $F_{IS}$ : 0.306 | 29.32       | 0.000 **       |   |
| Within individuals                                 | 302  | $F_{IT}$ : 0.334 | 66.58       | 0.000 **       |   |
| Total  | 603  |                  |             |                |   |
| <b>Macaronesia–Mediterranean</b>                   |      |                  |             |                |   |
| Among group (2 groups, 11 populations)             | 1    | $F_{CT}$ : 0.013 | 1.30        | 0.006 **       | 1. Fai, Gig, Cav<br>2. Ben, Col, Pop, Sic, Cre, Kos, Tor, Her |
| Among populations within groups                    | 9    | $F_{SC}$ : 0.009 | 0.87        | 0.000 **       |   |
| Within populations                                 | 253  | $F_{IS}$ : 0.309 | 30.32       | 0.000 **       |   |
| Within individuals                                 | 264  | $F_{IT}$ : 0.325 | 67.51       | 0.000 **       |   |
| Total  | 527  |                  |             |                |   |
| <b>Mediterranean–Alboran Sea</b>                   |      |                  |             |                |   |
| Among group (2 groups, 8 populations)              | 1    | $F_{CT}$ : 0.005 | 0.54        | 0.088          | 1. Ben, Col, Pop, Sic, Cre, Kos                               |
| Among populations within groups                    | 6    | $F_{SC}$ : 0.006 | 0.60        | 0.002 **       | 2. Tor, Her   |
| Within populations                                 | 184  | $F_{IS}$ : 0.320 | 31.61       | 0.000 **       |   |
| Within individuals                                 | 192  | $F_{IT}$ : 0.327 | 67.26       | 0.000 **       |   |
| Total  | 383  |                  |             |                |   |
| <b>Western Mediterranean–Eastern Mediterranean</b> |      |                  |             |                |   |
| Among group (2 groups, 6 populations)              | 1    | $F_{CT}$ : 0.001 | 0.13        | 0.393          | 1. Ben, Col, Pop<br>2. Sic, Cre, Kos                          |
| Among populations within groups                    | 4    | $F_{SC}$ : 0.001 | 0.03        | 0.472          |   |
| Within populations                                 | 138  | $F_{IS}$ : 0.326 | 32.58       | 0.000 **       |   |
| Within individuals                                 | 144  | $F_{IT}$ : 0.327 | 67.26       | 0.000 **       |   |
| Total  | 287  |                  |             |                |   |

Four different groupings and subsets of populations are represented. Populations comprising each group are listed on the right column

\* Significant at  $p < 0.05$ ; \*\* Significant at  $p < 0.01$

graphs. The two populations from the Alboran Sea appeared somewhat offset from the others, Torremuelle at one extreme of the first axis, La Herradura at one extreme along the second axis. No differentiation was apparent among populations of eastern and western Mediterranean, which showed interspersed centroids and widely overlapping inertia ellipses.

The results of the AMOVA analyses are coherent with the results from clustering and ordination methods (Table 2). An AMOVA considering as groups the Brazilian (western Atlantic), Macaronesian (eastern Atlantic), and Mediterranean populations (thus including the whole data set) showed low but highly significant percentage of variation between groups and among populations within groups. The same outcome was found when excluding

western Atlantic populations and considering the Macaronesian (eastern Atlantic) and the Mediterranean populations as different groups. However, in an analysis comparing the Alboran Sea with the rest of the Mediterranean populations the “among group” component explained only 0.54% of the variance and was not significant, while the among populations within groups component was still significant ( $p = 0.002$ ). Finally, if we restrict the analysis to the Mediterranean populations excluding the Alboran Sea and compare western with eastern Mediterranean populations, the “among group” and the “among populations within groups” components were not significant ( $p = 0.393$  and  $p = 0.472$ , respectively), pointing to a lack of gene-flow restriction across the Siculo-Tunisian Strait. In all cases, most of the variation was contained within populations

(29.32–32.58%) and, particularly, within individuals ( $F_{IT}$ ) (66.58–67.51%) (Table 2).

Assessing the hypothesis of isolation by distance through the Mantel test revealed significant correlation between genetic and geographic distances ( $r = 0.859$ ,  $p < 0.001$ ) when considering all populations. The correlation was weaker, but still significant, when removing the Brazilian populations ( $r = 0.384$ ,  $p = 0.025$ ), and no correlation was found when considering just the Mediterranean Sea ( $r = 0.189$ ,  $p = 0.179$ ) (see correlation graphs in Supplementary Fig. S5).

**Table 3** Migration patterns in *A. lixula*

| Migration direction      | Mean          | Confidence interval   |
|--------------------------|---------------|-----------------------|
| Mediterranean→E Atlantic | <b>24.182</b> | <b>(5.333–42.667)</b> |
| E Atlantic→Mediterranean | 13.042        | (0.000–29.333)        |
| Mediterranean→W Atlantic | <b>18.336</b> | <b>(0.667–35.333)</b> |
| W Atlantic→Mediterranean | 7.145         | (0.000–24.000)        |
| E Atlantic→W Atlantic    | 8.629         | (0.000–24.667)        |
| W Atlantic→E Atlantic    | 14.919        | (0.000–33.333)        |

Bold values represents confidence intervals that do not include zero  
Mutation-scaled effective migration rates ( $M$ ) between three marine areas: western Atlantic (W Atlantic), Eastern Atlantic (E Atlantic) and Mediterranean Sea. Mean values of  $M$  (Mean) and the 95% interval of confidence are presented in the table

The results of migration patterns between western Atlantic, eastern Atlantic and Mediterranean Sea are presented in Table 3. Migration outputs showed a general overlapping of the 95% confidential intervals around the  $M$  estimates between areas. Only  $M$  estimations from the Mediterranean to eastern Atlantic, and from the Mediterranean to the western Atlantic, which were also the highest values of  $M$  (mean 24.182 and 18.336, respectively), did not include zero within the confidence interval. These results may suggest a potential pattern of asymmetric and long distance migration that mainly occurs westwards. All the other estimations presented lower values of the  $M$  mean, ranging from 7.145 to 14.919, and wide confidence intervals that always included zero (Table 3).

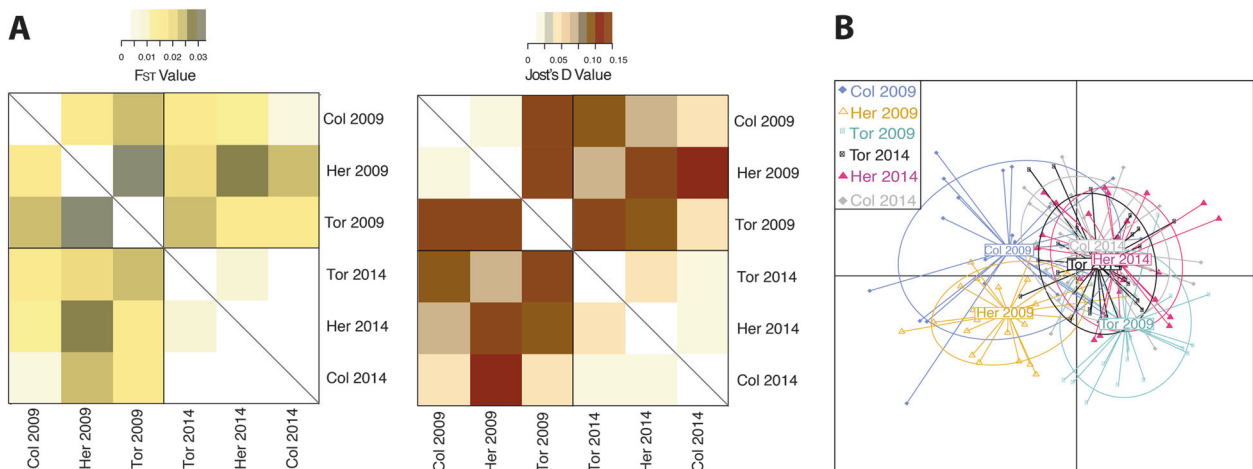
**Temporal genetic trends**

For the three populations that were re-sampled in 2014 (Torremuelle and La Herradura at the Alboran Sea, and Colera at the northwestern Mediterranean), the discerned genetic diversity was higher than that recorded in 2009, in terms of observed heterozygosity and mean allele number (except Colera for the latter parameter). Likewise,  $F_{IS}$  values were lower, likely indicating less inbreeding (Table 1). Both  $F_{ST}$  and  $D_{est}$  estimators showed significant genetic differentiation between Torremuelle and the other two

**Table 4** Estimated values of effective population size using three different temporal method estimates

| Per pop/Temporal methods | Pollak           | Nei/Tajima       | Jorde/Ryman         |
|--------------------------|------------------|------------------|---------------------|
| Colera                   | 177.3 (56.5–∞)   | 268.3 (67–∞)     | 387.9 (264.5–534.5) |
| La Herradura             | 33.9 (18.1–66.6) | 38.2 (19.9–78.8) | 35.3 (23.2–50)      |
| Torremuelle              | 34.2 (19.3–63.3) | 38.8 (21.4–75.5) | 38.1 (26.3–52.2)    |

Values within parentheses represent the 95% confidence interval of the estimates



**Fig. 4** Genetic differences over time in *Arbacia lixula*. **a** Heatmaps based on  $F_{ST}$  (left) and  $D_{est}$  (right) values between years and three sites, and **b** DAPC analyses for the three populations at two time points (2009 and 2014)

populations in 2009 ( $p < 0.015$ ), whereas La Herradura and Colera did not show significant differences between them in 2009. In 2014, the three populations displayed no significant differences in genetic structure (Supplementary Table S3). Genetic distances also revealed that the north-western Mediterranean population of Colera did not significantly change in genetic structure between 2009 and 2014, whereas both populations at the Alboran Sea, Torremuelle and La Herradura, demonstrated significant differences in genetic structure between 2009 and 2014. Therefore, Alboran Sea populations significantly changed their genetic structure over time (Supplementary Table S3 for  $F_{ST}$  and  $D_{est}$ , and Fig. 4). Mean differentiation values between years in the three populations ( $F_{ST}$ :  $0.040 \pm 0.015$ ,  $D_{est}$ :  $0.103 \pm 0.029$ , mean  $\pm$  SE) were higher, but of the same order, than mean genetic divergence detected in the spatial study among the Mediterranean populations ( $F_{ST}$ :  $0.015 \pm 0.002$ ,  $D_{est}$ :  $0.087 \pm 0.007$ , mean  $\pm$  SE).

A heatmap representation of the  $F_{ST}$  and  $D_{est}$  values (Fig. 4a) highlighted this pattern of marked interannual differences, but showed also that the three populations were more divergent among them in 2009 than in 2014. A DAPC representation (Fig. 4b, 20 PCs retained) revealed this same pattern: the three populations were more separated in 2009 (particularly Tor), but clustered tightly in 2014.

Considering one generation per year, the different estimators of effective population size (Table 4) revealed low values in all populations (approximate range 30–400 individuals). There were consistently higher sizes in the northern population of Colera (177.3–387.9 individuals, according to the different methods) than in the Alboran sea populations of La Herradura (33.9–38.3) and Torremuelle (34.2–38.8). The three estimators yielded remarkably similar estimates (and confidence intervals) in the southern populations, but varied by a factor of ca. 2 for the Colera population, for which defined confidence intervals could be obtained only with the unbiased Jorde/Ryman's estimator.

## Discussion

The amphiatlantic sea urchin, *Arbacia lixula* displayed significant nuclear divergence among the western Atlantic, eastern Atlantic and Mediterranean Sea. Additionally, variable structure across the transitional area of the Alboran Sea was also detected, which can be attributed to the interannual variation in the oceanographic circulation across this area.

Populations of *A. lixula* showed a high degree of genetic diversity. There was, however, a strong deficit of heterozygotes in all populations, with significant departure from HWE. This is unexpected for species with long pelagic larval duration. However, Addison and Hart (2005),

reviewing data for 124 marine invertebrates, showed a prevalence of positive  $F_{IS}$  values even in species with planktonic larvae. It can be explained by several factors, such as null alleles, mating among relatives, or unrecognized spatial and temporal structure within samples (Wahlund effects). The scarcity of null alleles indicates that our result is not an artefact of the markers. A potential explanation in our case is that assortative mating occurs linked to different gamete recognition proteins. Bindin, the sperm protein implicated in the fertilization of the egg, is well known in sea urchins (Metz et al. 1998; Zigler and Lessios 2003; Zigler et al. 2005; Lessios et al. 2012). Calderón and Turon (2010) showed that assortative mating linked to selected positions in the bindin gene of *Paracentrotus* explained inter-cohort differentiation. Such non-random mating structures, as well as the presence of spatial breeding groups, linked to stochasticity in reproductive success, patchiness in gamete distribution and the collective dispersal of genetically related larvae in the plankton (e.g. Broquet et al. 2013; Couvray and Coupé 2018), can explain the lack of HWE detected in *Arbacia*. In *A. lixula*, as in many other species, most genetic diversity was retained within populations and individuals (e.g. Calderón et al. 2008; Garcia-Cisneros et al. 2016).

Our nuclear results showed a sharp divergence between the western and eastern Atlantic areas, likely due to the combined effect of isolation by distance and the strong disruptive effect of the deep mid-Atlantic barrier. This sharp genetic divergence is similar to the one observed in other amphiatlantic echinoderms with large dispersal potential (e.g. Lessios et al. 2001; Garcia-Cisneros et al. 2017). The nuclear divergence in *A. lixula* was also largely congruent with COI mitochondrial data, but historical migration patterns and allele frequencies highlighted interesting insights in its phylogeography. Lessios et al. (2012) and Wangenstein et al. (2012) hypothesized the colonization of the western Atlantic coast, across the mid-Atlantic barrier, from eastern Atlantic stocks. However, neither migration nor allele distribution patterns from our new nuclear results fully supported this hypothesis and suggested instead the Mediterranean as a potential source of colonizers. Migration patterns estimated from microsatellites showed asymmetric gene flow among areas, with the most important historical migration likely flowing westward from the Mediterranean to the eastern and western Atlantic. Our results discard large historical connectivity between eastern and western Atlantic regions, which showed a low value of  $M$ . In addition, the Mediterranean origin of the western Atlantic populations can be also supported by 14 alleles shared (out of 250) between these two areas, whereas only two alleles were found in common between the eastern and western Atlantic stocks that can be indicative of long-term isolation between populations at both sides of the Atlantic. Interestingly, a

detailed re-evaluation of the COI network also points out the potential origin of the Brazilian haplotype cluster from some of the most frequent Mediterranean haplotypes. Therefore all current genetic evidences suggest divergence of the western Atlantic populations of *A. lixula* from the Mediterranean area, which likely happened after the Pleistocene colonization and demographic expansion in the Mediterranean Sea (93.8–205.2 kya) (Wangensteen et al. 2012). Nonetheless, further investigations are necessary to discard other unexplored genetic stocks and to confirm the Mediterranean origin of the western Atlantic lineages.

Additionally, subtler structure is also found in the Atlantic–Mediterranean area, with significant differentiation between the Macaronesian islands and the Mediterranean. The biogeographic break between Atlantic and Mediterranean leaves a strong signature in the genetic structure of many species of fish and invertebrates with different biological characteristics (Patarnello et al. 2007; Pascual et al. 2017), including sea urchins, sea stars, brittlestars and sea cucumbers (Borrero-Pérez et al. 2011; Pérez-Portela et al. 2010; Calderón et al. 2012; Taboada and Pérez-Portela 2016; Garcia-Cisneros et al. 2016, 2017). However, the Mediterranean Sea also has a number of internal oceanographic barriers that can restrict species dispersal. Among the better identified oceanographic barriers within the Mediterranean are: the Gibraltar Strait and the Almeria-Oran Front between the Atlantic and Mediterranean basins, the Ibiza Channel and Balearic Front dividing the north- and southwestern Mediterranean sub-basins, the Siculo-Tunisian Front between the western and eastern Mediterranean, and the Otranto Strait and Aegean Front delimiting the Adriatic and Aegean seas, respectively (e.g., Villamor et al. 2014; Riesgo et al. 2016; Garcia-Cisneros et al. 2016; and a review in Pascual et al. 2017).

Nevertheless, these oceanographic fronts do not have equal effect on all marine species. Pascual et al. (2017), reviewing published information for 70 species, found that the reduction of gene flow linked to the presence of the abovementioned oceanographic fronts is more important in species with long planktonic durations. This unexpected pattern is likely because these larvae move off-shore, along the continental shelf and slope, and are thus more affected by major oceanographic circulation and marine fronts than larvae that remain close to the coastline (Pascual et al. 2017). In our case, we detected genetic divergence between both sides of the Almeria-Oran front, as observed in other echinoderms (Calderón et al. 2012; Garcia-Cisneros et al. 2016, 2017), although the divergence detected in Alboran populations of *A. lixula* may actually be a transient process, as discussed below for the temporal analyses, rather than a permanent one. Nevertheless, we could not find any evidence of genetic divergence between the western and eastern Mediterranean sub-basins, nor was there any

significant isolation by distance effect in the Mediterranean, a pattern that contrasts with other echinoderms with large dispersal potential across the same geographical area (e.g. Garcia-Cisneros et al. 2016, 2017). This may suggest that *A. lixula* is not largely affected by discontinuities between the Mediterranean populations, representing a well-mixed genetic pool within this sea, and/or it reflects the recent evolutionary history within this basin, marked by a demographic expansion (Wangensteen et al. 2012), with not enough time to diverge within the Mediterranean basins.

The temporal genetic patterns among the two populations from the Atlantic–Mediterranean transition and the one from the north-western Mediterranean indicate that populations were more divergent, particularly Torremuelle (Alboran Sea), in 2009 than in 2014. Interannual variations in the hydrological features along the Iberian Mediterranean shores are well known (Pascual et al. 2002; Pinot et al. 2002; Bouffard et al. 2010; Balbín et al. 2014), and have been held responsible for temporal patterns of genetic variation in organisms such as the fish *Sardina pilchardus* (Olivar et al. 2003), the sea urchin *Paracentrotus lividus* (Calderón et al. 2012), or the crab *Liocarcinus depurator* (Pascual et al. 2016). In particular, in the Alboran area, there is a complex structure with two main anticyclonic gyres and a central cyclonic gyre (Sanchez-Vidal et al. 2004; Sánchez-Garrido et al. 2013). The relative intensity of these gyres changes over time, and it determines a temporally variable system of hydrological fronts in the area (Renault et al. 2012; Oguz et al. 2014). These features affect the interplay between Atlantic and Mediterranean waters, leading to variable patterns of distribution of water masses in the Alboran Sea. This can explain our finding of significant temporal genetic differences in Torremuelle and La Herradura located in the Alboran Sea, while the northern population of Colera, outside of this transitional area, remained more stable. Such temporal changes in genetic composition of southern Iberian populations relative to more northern populations were also detected for *Paracentrotus lividus* (Calderón et al. 2012). Torremuelle, in particular, lies in western Alboran Sea, in a relatively isolated spot just outside the frontal system generated by the western anticyclonic gyre (Sánchez-Garrido et al. 2013; Oguz et al. 2014). Thus, arrival of larvae to this locality is subject to stochastic and oceanographic changes among years, which may explain its higher genetic distance compared to other Mediterranean populations.

The effective population sizes ( $N_e$ ) detected examining temporal variation in genetic composition were small (from tens to a few hundred individuals), and similar to  $N_e$  estimates for *P. lividus* (Calderón et al. 2009). In this study, we did not specifically measure *A. lixula* abundances but information obtained from other studies showed densities that vary across space and time from low-density

populations (0.2–0.3 individuals/m<sup>2</sup>) to densely populated sites (over 1.0 individuals/m<sup>2</sup>) (Palacín et al. 1998; Hereu et al. 2012). It is common for invertebrates and fish to have effective population sizes 2–6 orders of magnitude smaller than census sizes (Turner et al. 2002; Hauser and Carvalho 2008; Plough 2016), which is often explained by large variance of reproductive success, whereby only few adults are able to produce successful progeny (sweepstake reproduction, Hedgecock 1994). Statistic methods to calculate effective population size based on genetic data at two time points are appropriate to estimate contemporary  $N_e$  that reflects the effective number of parental specimens from which the collected sample comes from (e.g. Casilagan et al. 2013). Thus, the stochastic events that can take place during the reproduction together with the long planktonic period and the settlement and recruitment phases of *A. lixula* can likely explain the small effective population sizes detected. It is noteworthy that the hydrologically more stable northern population of Colera had ca. 6–10 times larger effective population sizes than the two southern populations.

From the last few years, there is increasing evidence of the importance of *A. lixula* in the formation and maintenance of bare habitats (Bulleri et al. 1999; Gianguzza et al. 2011; Bonaviri et al. 2011). *Arbacia lixula* is a thermophilous species likely to be enhanced by warming temperatures (Francour et al. 1994; Gianguzza et al. 2011; Wangensteen et al. 2013a) and a generalist species with a catholic diet that qualifies it as omnivore (Wangensteen et al. 2011; Agnetta et al. 2013) not affected by a commercial fishing industry. Thus, although some populations of *A. lixula* at the Levant basin are in decline due to the ecosystem collapsing (Rilov 2016), under the current scenario of the ongoing tropicalization of the Mediterranean, *A. lixula* can be favoured, leading to important changes in ecosystem structure and functioning.

This study shows a main genetic break in *A. lixula* between both sides of the Atlantic, and smaller differentiation signals associated with the Atlanto-Mediterranean transition. However, no genetic structure was found within the Mediterranean populations, suggesting that either the species' dispersal abilities suffice to break the hydrological barrier separating the two Mediterranean sub-basins and/or the genetic homogeneity is the result of the recent evolutionary history of the species, although both hypotheses are not mutually exclusive. A picture of genetic homogeneity across the Mediterranean implies that the species may safely overcome occasional adverse local conditions and quickly replenish populations from neighbouring and distant locations. Future research, including whole-genome scans and the inclusion of populations from other areas (such as the Adriatic sea, Levant basin and/or the Atlantic African shores) will likely show a more nuanced picture of the

underlying genetic structure associated with adaptation (e.g. Carreras et al. 2017). Overall, however, the patterns found suggest that the spread potential of *A. lixula* in the Mediterranean is large and the ongoing expansion of this thermophilous species will not be restricted by the potential impact of postulated barriers to gene flow.

## Data archiving

Data sets are available from Mendeley at: <https://doi.org/10.17632/kzrgwmpztt.1> and <https://doi.org/10.17632/ksgcv7dd4y.1>

**Acknowledgements** This research was financially supported by the Spanish Government projects CTM2013-48163 and CTM2017-88080 (AEI, FEDER-UE) and by a “Juan de la Cierva” contract from the Spanish Government to RPP. We are indebted to Carlos Renato Rezende Ventura for supplying us with the Brazilian samples, and to Jacob González-Solís for providing the samples from Cape Verde.

## Compliance with ethical standards

**Conflict of interest** The authors declare that they have no conflict of interest.

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