

## ORIGINAL ARTICLE

# Molecular analysis of phosphate limitation in *Geobacteraceae* during the bioremediation of a uranium-contaminated aquifer

A Lucie N'Guessan<sup>1,2</sup>, Hila Elifantz<sup>1</sup>, Kelly P Nevin<sup>1</sup>, Paula J Mouser<sup>1</sup>, Barbara Methé<sup>3</sup>, Trevor L Woodard<sup>1</sup>, Kimberly Manley<sup>1</sup>, Kenneth H Williams<sup>4</sup>, Michael J Wilkins<sup>5</sup>, Joern T Larsen<sup>4</sup>, Philip E Long<sup>2</sup> and Derek R Lovley<sup>1</sup>

<sup>1</sup>Department of Microbiology, University of Massachusetts, Amherst, MA, USA; <sup>2</sup>Department of Microbiology, Pacific Northwest National Laboratory, Richland, WA, USA; <sup>3</sup>Microbial and Environmental Genomics Group, J Craig Venter Institute, Rockville, MD, USA; <sup>4</sup>Earth Sciences Division, Lawrence Berkeley National Laboratory, Berkeley, CA, USA and <sup>5</sup>Department of Earth and Planetary Science, University of California, Berkeley, CA, USA

**Nutrient limitation is an environmental stress that may reduce the effectiveness of bioremediation strategies, especially when the contaminants are organic compounds or when organic compounds are added to promote microbial activities such as metal reduction. Genes indicative of phosphate-limitation were identified by microarray analysis of chemostat cultures of *Geobacter sulfurreducens*. This analysis revealed that genes in the *pst-pho* operon, which is associated with a high-affinity phosphate uptake system in other microorganisms, had significantly higher transcript abundance under phosphate-limiting conditions, with the genes *pstB* and *phoU* upregulated the most. Quantitative PCR analysis of *pstB* and *phoU* transcript levels in *G. sulfurreducens* grown in chemostats demonstrated that the expression of these genes increased when phosphate was removed from the culture medium. Transcripts of *pstB* and *phoU* within the subsurface *Geobacter* species predominating during an *in situ* uranium-bioremediation field experiment were more abundant than in chemostat cultures of *G. sulfurreducens* that were not limited for phosphate. Addition of phosphate to incubations of subsurface sediments did not stimulate dissimilatory metal reduction. The added phosphate was rapidly adsorbed onto the sediments. The results demonstrate that *Geobacter* species can effectively reduce U(VI) even when experiencing suboptimal phosphate concentrations and that increasing phosphate availability with phosphate additions is difficult to achieve because of the high reactivity of this compound. This transcript-based approach developed for diagnosing phosphate limitation should be applicable to assessing the potential need for additional phosphate in other bioremediation processes.**

*The ISME Journal* (2010) 4, 253–266; doi:10.1038/ismej.2009.115; published online 10 December 2009

**Subject Category:** geomicrobiology and microbial contributions to geochemical cycles

**Keywords:** geobacteraceae; phosphate-limitation; uranium bioremediation; gene expression

## Introduction

The practice of groundwater bioremediation is often highly empirical (Lovley, 2003). For example, a common practice intended to stimulate the activity of subsurface microorganisms involved in bioremediation is to add nutrients, such as nitrogen or phosphate (Martani and Seto, 1991; Norris and

Dowd, 1994; Hincee *et al.*, 1995; Ronen *et al.*, 1996). The rationale is that phosphate is an essential nutrient for microbial metabolism and that phosphate availability may be low in, at least, some contaminated environments. However, without information on the physiological status of the subsurface microorganisms, it is not clear whether such phosphate amendments are warranted or even desirable. Addition of phosphate when the microorganisms are not limited for phosphate or adding more phosphate than is needed to overcome phosphate limitation adds unnecessary costs to bioremediation. Furthermore, adding phosphate beyond what is necessary to support optimal microbial

Correspondence: AL N'Guessan, Department of Microbiology, Pacific Northwest National Laboratory, PO Box 999, Richland, WA 99352, USA.

E-mail: lucie.nguessan@pnl.gov

Received 27 May 2009; revised 28 September 2009; accepted 2 October 2009; published online 10 December 2009

activity could have unintended negative consequences. For example, phosphate may form insoluble precipitates with trace metals and other cations that are also important nutrients, making these nutrients less available. Thus, information on the *in situ* phosphate requirements before initiating a bioremediation strategy and during the course of the bioremediation could make it possible to alleviate phosphate limitations impeding optimal bioremediation in the most resource- and cost-effective manner.

One of the most effective strategies for monitoring the metabolic state of microorganisms involved in bioremediation is to quantify *in situ* levels of gene expression (Lovley, 2002, 2003). Geochemical processes in pure cultures as well as in the environment can be linked to increased levels of mRNA for a particular gene (Holmes *et al.*, 2004b, 2005). For example, genes involved in naphthalene degradation (*nahA*) were also detected in sediments where naphthalene mineralization was taking place (Wilson *et al.*, 1999). Similarly, high levels of *merA* were measured when high rates of mercury volatilization occurred (Nazaret *et al.*, 1994). In addition, three reductive dehalogenase genes were monitored in samples where trichloroethene bioremediation was applied (Lee *et al.*, 2008).

*Geobacter* species are important agents in the bioremediation of organic contaminants (Lovley *et al.*, 1989; Rooney-Varga *et al.*, 1999; Lin *et al.*, 2005), as well as uranium (Anderson *et al.*, 2003; North *et al.*, 2004; Vrionis *et al.*, 2005) and vanadium (Ortiz-Bernad *et al.*, 2004b). Analysis of gene transcript abundance within the subsurface *Geobacter* community has been successful in diagnosing rates of metabolism in *Geobacter* species as well as nutrient limitations and stress responses (Holmes *et al.*, 2004b, 2005; O'Neil *et al.*, 2008; Mouser *et al.*, 2009a). Analysis of levels of transcripts for the nitrogen fixation gene, *nifD*, demonstrated that subsurface *Geobacter* species were limited for fixed nitrogen in petroleum-contaminated subsurface sediments and during *in situ* uranium bioremediation (Holmes *et al.*, 2004b, 2005). This may not be surprising because in petroleum-contaminated aquifers and in aquifers amended with organic electron donors to promote dissimilatory metal reduction there is likely to be an abundance of organic carbon, but limiting quantities of other nutrients. Although *Geobacter* species can overcome limitations for fixed nitrogen by fixation of abundant nitrogen gas (Bazylinski *et al.*, 2000; Coppi *et al.*, 2001; Holmes *et al.*, 2004a), phosphate sources are likely to be more limited. Until now it has not been possible to assess possible phosphate limitations of *Geobacter* species because of a lack of information on genes, which might be indicative of phosphate limitation.

In other microorganisms, such as *Bacillus subtilis*, *Clostridium acetobutylicum*, *Escherichia coli* or *Pseudomonas aeruginosa*, phosphate-limiting conditions

are associated with increased expression of genes in the *pst-pho* operon, which encode genes associated with a high-affinity phosphate uptake system (Nakata *et al.*, 1984; Nikata *et al.*, 1996; VanBogelen *et al.*, 1996; Aguenta *et al.*, 2002; Allenby *et al.*, 2004; Fischer *et al.*, 2006). All of the available *Geobacter* species genomes contain at least one copy of the *pst-pho* operon, suggesting that organisms in this genus share this high-affinity uptake system. In the natural environment, the availability of phosphorus often determines microbial growth and activity. For this reason, most bacteria have two phosphate transport systems, which differ in affinity and specificity for phosphate (Bardin and Finan, 1998). The low-affinity phosphate transport system (Pit) transports phosphate into the cells in the presence of high levels of extracellular phosphate (Gachter and Meyer, 1993; Voegelé *et al.*, 1997; Bardin and Finan, 1998; Van Dien and Keasling, 1999; Botero *et al.*, 2000; Harris *et al.*, 2001). It is usually constitutively expressed and is dependent on the proton motive force. The high-affinity phosphate specific transport system is induced during phosphate starvation by the Pho regulon and is an ABC transporter (Muda *et al.*, 1992; Nikata *et al.*, 1996; Aguenta *et al.*, 2002; Gebhard *et al.*, 2006). Therefore, in most environments with low phosphate availability, the high-affinity phosphate transport system is expected to be the primary pathway for phosphate introduction into cells.

The purpose of this study was to identify genes whose expression might be diagnostic of phosphate-limiting conditions in *Geobacter* species and to analyze the expression of these genes during *in situ* uranium bioremediation to determine if the subsurface *Geobacter* species are limited for phosphate under the current strategy of adding only acetate to promote this process. The results suggest that *Geobacter* species are phosphate-limited, but that addition of phosphate may not stimulate metal reduction because of the high capacity of the sediments for phosphate adsorption.

## Materials and methods

### *Growth of Geobacter sulfurreducens in continuous culture*

*Geobacter sulfurreducens* strain PCA<sup>T</sup> (ATCC 51573) was obtained from our laboratory culture collection. *G. sulfurreducens* was grown in electron donor limited chemostats at dilution rate of 0.05 h<sup>-1</sup> with acetate (5 mM) provided as the electron donor and fumarate (27.5 mM) as the electron acceptor at 30 °C, as previously described (Esteve-Nunez *et al.*, 2005). For phosphate-limited chemostats, *G. sulfurreducens* was first grown to steady-state under electron donor limiting conditions as described above. The medium contained 0.217 mM phosphate. The medium was then switched to one that was phosphate free. The chemostats were harvested when the

optical density (OD<sub>600</sub>) had dropped below 0.2. Analyses of acetate, fumarate, succinate, malate and protein were carried out as previously described (Esteve-Nunez *et al.*, 2005).

To track the number of mRNA transcripts for *pstB* and *phoU*, *G. sulfurreducens* was grown under donor-limiting conditions, as described above. The medium was switched after approximately 30 h to one that was phosphate free. When the OD<sub>600</sub> had dropped below 0.1, the medium was again switched to one that contained phosphate. Chemostats were sampled at different time points and the nucleic acids were extracted from the collected biomass as described below.

#### *Microarray analysis comparing acetate limitation and phosphate limitation*

Deoxyribonucleic acid microarray hybridization and data analyses were carried out as previously described (Methe *et al.*, 2005). The expression ratio between acetate and phosphate limiting for three biological replicates and three technical replicates was calculated.

#### *Site and plot description*

Detailed descriptions of the geology, hydrology and geochemistry of the site in Rifle, CO, USA have been presented elsewhere (Anderson *et al.*, 2003; Vrionis *et al.*, 2005; Yabusaki *et al.*, 2007). The field experimental results described here were obtained as part of the 2007 study of the U.S. Department of Energy's Integrated Field Research Challenge (IFRC) site at Rifle, Co, USA. In 2007, the 10 m × 10 m experimental plot was similar in design as the one from previous field test described in Anderson *et al.* (2003) and Vrionis *et al.* (2005). The experimental plot comprised of an injection gallery, four back-ground-monitoring wells and twelve downgradient monitoring wells. The injection gallery was made of 10 1 m spaced injection wells and was positioned approximately perpendicular to groundwater flow (see Mouser *et al.*, 2009b).

Acetate (5 mM-target concentration) was injected into the aquifer for a total period of 30 days from August to September 2007. The injection period was interrupted by a groundwater flush from day 10 to day 17 of the experimental period, to mimic a period of acetate limitation. This study focuses on D02, a well located in the first row of downgradient wells, where some of the lowest U(VI) concentrations (<0.13 μM) were achieved by the end of the experimental period. Furthermore, *Geobacteraceae* were the predominant organisms in this well for the majority of the experimental period.

#### *Groundwater sampling and geochemical analysis*

Before sample collection, 12 l of water were purged from each well at a rate of approximately 2 l min<sup>-1</sup>. About 15 l of groundwater from upgradient and

downgradient wells were filtered onto a series of two 293 mm diameter filters, a 1.2 and 0.2 μm Supor membrane filter (Pall Life Sciences, Port Washington, NY, USA) at different time intervals through the course of the field experiment. The filters were flash frozen in an ethanol-dry ice bath, shipped on dry ice to the laboratory and stored at -80 °C until analysis.

Acetate, sulfate, nitrate and bromide were analyzed on site using an ion chromatograph (ICS-1000, Dionex, Sunnyvale, CA, USA) equipped with an autosampler (AS40, Dionex). The samples were separated through an AS22 column (Dionex) with a carbonate/bicarbonate (4.5 mM/1.4 mM) eluent. Phosphate was analyzed by ICP-MS. These samples were run at a 1:5 dilution on a Perkin-Elmer Elan DRC II ICP-MS (Perkin-Elmer, Waltham, MA, USA) using a Micro Mist 0.2 ml min<sup>-1</sup> nebulizer. The method used a dynamic reaction cell with oxygen as a reaction gas to form the oxide PO<sup>+</sup> (phosphorus oxide) and analyzing the oxide at mass 47, using Ga 69 as an internal standard. No other elements were analyzed with PO. Randomly selected samples were spiked with 5 and 20 μg l<sup>-1</sup> P (Phosphorus), typical recoveries were 99 and 105%, respectively. QC samples and laboratory reagent blanks were also run, showing minimal drift during analysis and no detectable contamination from sample preparation/sampling vessels.

#### *Extraction of mRNA and DNA from G. sulfurreducens chemostat cultures and environmental samples*

Deoxyribonucleic acid was extracted from pure cultures, and Rifle groundwater filters using the Qbiogene soil extraction kit (Qbiogene/MP Biomedicals, Solon, OH, USA) as per the manufacturer's instructions. Nucleic acids were extracted from chemostat samples as previously described, using a trizol:chloroform extraction (Methe *et al.*, 2005). Extraction of nucleic acids from groundwater samples were done as described elsewhere (Holmes *et al.*, 2004b) with a few modifications, such as: tRNA was not added to the procedure of mRNA extraction from groundwater samples; linear acrylamide was excluded from the precipitation step; finally, additional DNase treatment was carried out at the end of the extraction using a DNA-free kit (Ambion, Austin, TX, USA) for all samples.

#### *Design of primers and amplification of phoU and pstB genes*

Primers targeting *Geobacteraceae phoU*, *pstB*, and *proC* genes were designed from nucleotide sequences from *G. sulfurreducens* (Methe *et al.*, 2003), *Geobacter metallireducens*, *D. palmitatis*, *P. carbinolicus* and *P. propionicus* genomes. Preliminary sequence data from *G. metallireducens*, *D. palmitatis*, *P. carbinolicus* and *P. propionicus* genomes was obtained from the DOE Joint Genome

**Table 1** List of *Geobacteraceae* primers used in this study.

Primer set	Sequence	Amplified <i>Geobacteraceae</i> gene
phoU-3F phoU-3R	5'-ATCGTYACSGACCTGGAGCG-3' 5'-CCAKRTACTTGGAGATGTAG-3'	<i>phoU</i> from <i>G. psychrophilus</i> , <i>G. bremensis</i> , <i>G. uraniireducens</i> , Rifle environmental DNA
phoU-2F phoU-1R	5'-GASATCGAYGARAAGTGCCT-3' 5'-CATGAAGATSACCATYTCGG-3'	<i>phoU</i> from <i>G. bemidjensis</i> , <i>G. uraniireducens</i> , <i>G. psychrophilus</i> , Rifle environmental DNA
phoU-290F phoU-650R	5'-ACCTGGAGCGGATCGGCGAC-3' 5'-TCGTGTGGCGGATGTCTTC-3'	<i>phoU</i> from Rifle environmental DNA
pstB-2F pstB-1R	5'-TTCCAGAARCCGAACCCSTT-3' 5'-CCGAAMCGTCCSGTAATGTA-3'	<i>pstB</i> from <i>G. psychrophilus</i> , <i>G. bemidjensis</i> , <i>G. uraniireducens</i> , Rifle environmental DNA
pstB-3F pstB-1R	5'-RTCRGCATGGTSTTCCAGAA-3' 5'-CCGAAMCGTCCSGTAATGTA-3'	<i>pstB</i> from <i>G. bremensis</i>
phoU-401F phoU-521R	5'-TGGTGAAGGAGTCGCTGGA-3' 5'-AAGGTGAGCAGTTCCCGTTG-3'	<i>phoU</i> from <i>G. sulfurreducens</i> (quantitative PCR (qPCR))
pstB-310F pstB-422R	5'-CCGTTCCCCAAATCGATCT-3' 5'-ATGGCGGCGTTAGTGAGG-3'	<i>pstB</i> from <i>G. sulfurreducens</i> (qPCR)
phoU-28F phoU-269R	5'-AATATCGCCAAACGGTCG-3' 5'-CGGCTGATGGTGGAGGGA-3'	<i>phoU</i> from Rifle environmental DNA (qPCR)
pstB-155F pstB-339R	5'-CTGAAGACCAACGCCATGAC-3' 5'-ACGATGATGATGGTGTAGCG-3'	<i>pstB</i> from Rifle environmental DNA (qPCR)
proC-75F proC-471R	Sequences presented elsewhere (Holmes <i>et al.</i> , 2005)	<i>proC</i> from environmental DNA
proC-412F proC-496R	Sequences presented elsewhere (Holmes <i>et al.</i> , in preparation)	<i>proC</i> from <i>G. sulfurreducens</i> (qPCR)
proC-49F proC-143R	5'-GCGTCCAGCAGTTCTTCA-3' 5'-GGCTACAACGCAACCGATG-3'	<i>proC</i> from Rifle environmental DNA (qPCR)

Institute (JGI) website [www.jgi.doe.gov](http://www.jgi.doe.gov). Table 1 lists all of the primers used in this study. All primers had an optimal annealing temperature of 60 °C. The following PCR parameters were used: an initial denaturation step at 95 °C for 5 min, followed by 35 cycles of 95 °C (45 s), 60 °C (1 min), 72 °C (1 min) and with a final extension step at 72 °C for 10 min (Holmes *et al.*, 2004b). The PCR products were resolved on 1% agarose gel and the bands of *phoU* and *pstB* at the expected size were excised and purified with a Gel Extraction Kit (Qiagen, Valencia, CA, USA). Clone libraries were constructed with a TOPO TA cloning kit (Invitrogen, Carlsbad, CA, USA) according to the manufacturer's instructions. Plasmid inserts (48) from each cDNA clone library were then sequenced with the M13F primer using an ABI BigDye Terminator v3.1 (Applied Biosystems, Foster City, CA, USA) and ABI DNA Analyzer 3730xl (Applied Biosystems). Sequences were compared with those compiled in GenBank with the BLAST suit of programs to verify the specificity of each primer.

Sequences obtained were aligned and quantitative PCR (qPCR) primers designed (Table 1) to quantify *pstB* and *phoU* as well as *proC* in environmental samples as described below.

#### Phylogenetic analysis

*phoU* and *pstB* gene sequences were compared with NCBI GenBank nucleotide database using the blastn algorithms (Altschul *et al.*, 1990). Amino acid and nucleotide sequences for *phoU* and *pstB* genes were initially aligned in MEGA version 4 (Tamura *et al.*, 2007) where phylogenetic trees were inferred. Distances and branching order were determined and compared using maximum parsimony and distance-based algorithms (Neighbor-joining and Kimura). Bootstrap values were obtained from 100 replicates.

#### Quantification of mRNA transcript levels using reverse transcriptase qPCR

The reverse transcriptase enhanced avian kit (Sigma, St Louis, MO, USA) was used to reverse complement RNA from either environmental samples or *G. sulfurreducens* cultures in chemostats according to manufacturer's instructions. A 1:10 dilution of the RNA template was made before the reaction to minimize template inhibition. Subsequently, the cDNA was also diluted (1:10) before the qPCR reaction. Each PCR reaction consisted of a

total volume of 25 µl and contained 1.5 µl of primers (stock concentration was 20 µM), 12.5 µl Power SYBR green PCR mix (PE Biosystems, Foster City, CA, USA) and 9.5 µl template. Standard curves were constructed as previously described (Holmes *et al.*, 2004b), and covered a range of 8 orders of magnitude. qPCR reactions were carried out on a 7500 ABI real time (Applied Biosystems, Foster City, CA, USA) according to the ABI guide. The thermal cycling parameter consisted of an activation step at 50 °C for 2 mins, a denaturation step of 95 °C for 10 mins, followed by 45 cycles at 95 °C for 15 secs and 60 °C for 1 min. Triplicate runs of qPCR were done and the fold change was calculated as the ratio between baseline and the change observed. To verify amplification and correct amplicon size, aliquots from real-time PCR were examined on a 1% agarose gel stained with ethidium bromide. The qPCR primers used are listed in Table 1. The housekeeping gene, *proC*, was used for the normalization of chemostat and environmental data.

#### Laboratory sediment incubations with phosphate amendment

Laboratory sediment incubations were conducted, as previously described (N'Guessan *et al.*, 2008), with sediments collected near the acetate-injection test plot. The anoxic sediment-groundwater slurries were amended with 10 mM acetate and 250 µM sodium phosphate (target groundwater concentration). The sediment and groundwater from each set of triplicates was sampled over time under anaerobic conditions. Incubations without phosphate amendments served as controls. Analyses of acetate, sulfate and U(VI) in the groundwater sample was carried out as described above. Iron and uranium concentrations in sediment were determined as previously described (N'Guessan *et al.*, 2008). Phosphate was analyzed by ion chromatography (ICS-1000) as described above.

## Results and discussion

#### Expression of *phoU* and *pstB* in response to phosphate limitation in chemostats

To identify *Geobacter* genes with increased transcript levels during phosphate limitation, chemostats of *G. sulfurreducens* were grown under steady-state conditions with acetate as the electron donor and fumarate as the electron acceptor. Acetate availability was the factor limiting growth. The phosphate concentration was 0.217 mM. Once steady-state conditions (protein, 0.045 ± 0.0001 mg ml<sup>-1</sup>; fumarate, 4.054 ± 0.753 mM; and acetate <50 µM) were established for four culture vessel medium turnovers, the medium being added to the chemostat was changed to one that was identical with the exception that phosphate was omitted. As phosphate in the culture vessel was diluted, the cell density started to decrease, indicative of phosphate

limitation. Cells were harvested for microarray analysis when the phosphate concentration had decreased to 0.086 mM. At this point cell protein had decreased (0.029 ± 0.004 mg ml<sup>-1</sup>) and (fumarate 6.607 ± 0.436 mM) and acetate (1.106 ± 0.300 mM) concentrations were much higher. These changes demonstrate that cell growth was phosphate-limited.

The switch from electron donor-limiting to phosphate-limiting conditions was associated with changes in expression of a variety of genes involved in transport and binding as well as energy metabolism. A total of 35 genes had transcript levels at least twofold higher under the phosphate-limiting conditions (Table 2), whereas 53 genes had transcript levels that were at least twofold lower under this same condition (Table 3).

Decreases in transcript abundance were all less than fivefold (Table 3). Most of the genes with lower transcript levels in phosphate-limited cells were those involved in metabolism and cell growth and included electron transport proteins, various transport and binding proteins, ribosomal proteins and amino acid biosynthesis. The decreased level of transcripts in these genes is consistent with the slower growth of the phosphate-limited cells and the fact that phosphorus is an essential nutrient for the synthesis of nucleic acids and ATP. A similar response was previously noted in phosphate-limited cells of *Sinorhizobium meliloti* (Bardin and Finan, 1998; Krol and Becker, 2004).

The gene with the highest increase in transcript levels in phosphate-limited cells was *pstB* (Table 2). *pstB* has been annotated as the ATP-binding protein of a phosphate ABC transporter (Methe *et al.*, 2003; Yan *et al.*, 2004). It is part of a predicted operon, which also includes *phoU* (GSU1095), *pstA* (GSU1097), *pstC* (GSU1098) and *pstS* (GSU1099). *pstA*, *pstC* and *pstS* were also upregulated and are annotated to have a role in phosphate uptake, whereas *phoU* is predicted to regulate the expression of the *pst* genes (Table 2), consistent with similar operon organization of related genes in other microorganisms (Nakata *et al.*, 1984; Nikata *et al.*, 1996; VanBogelen *et al.*, 1996; Bardin and Finan, 1998; Aguena *et al.*, 2002; Fischer *et al.*, 2006). The *pst-pho* operon was previously shown to be the most strongly up-regulated operon under phosphate-limiting conditions in *S. meliloti* (Krol and Becker, 2004) and *Acidithiobacillus ferrooxidans* (Vera *et al.*, 2003; Krol and Becker, 2004).

Other genes with a greater than fivefold increase in transcript levels in phosphate-limited cells included a putative metal ion efflux outer membrane family protein (GSU1330); an ABC transporter permease protein for an unknown substrate (GSU1340); a protein in the GH3 auxin-responsive promoter family but of unknown function (GSU1092); and two hypothetical proteins (GSU 2700 and GSU 1100). The majority of the remaining proteins with transcript levels that were more than twofold higher in phosphate-limited cells

**Table 2** Reporters whose expression was significantly increased when *Geobacter sulfurreducens* was grown with phosphate as limiting nutrient versus acetate as limiting nutrient

Locus ID	Common name	Main role	Subrole	Fold change	s.d.
GSU0074	elbB, enhancing lycopene biosynthesis protein 2	Biosynthesis of cofactors, prosthetic groups and carriers	Other	2.622	0.251
GSU2078	rodA, rod shape-determining protein RodA	Cell envelope	Biosynthesis of murein sacculus and peptidoglycan	2.243	0.223
GSU1984	gumC, polysaccharide chain length determinant protein, putative		Other	2.131	0.310
GSU1330	Metal ion efflux outer membrane protein family protein, putative	Cellular Processes	Detoxification	7.472	3.464
GSU1332	Heavy metal efflux pump, CzcA family	DNA metabolism		3.052	0.738
GSU0997	mutM, formamidopyrimidine-DNA glycosylase	Energy metabolism	DNA replication, recombination and repair	2.048	0.270
GSU1538	Methylamine utilization protein MauG, putative		Amino acids and amines	2.955	1.003
GSU2732	orf2-2, cytochrome c family protein	Hypothetical proteins	Electron transport	2.227	0.441
GSU0987	Conserved hypothetical protein		Conserved	3.087	0.408
GSU2558	Conserved domain protein		Domain	4.840	2.610
GSU2075	subtilisin	Protein fate	Degradation of proteins	2.355	0.388
GSU1102	DNA-binding response regulator	Regulatory functions	DNA interactions	6.496	1.525
GSU1095	phoU, phosphate transport system regulatory protein PhoU		Other	9.067	2.953
GSU0896	tldD protein			2.012	0.180
GSU1101	Sensory box histidine kinase	Transport and binding proteins	Protein interactions	5.652	0.999
GSU1096	pstB, phosphate ABC transporter, ATP-binding protein		Anions	12.935	4.976
GSU1097	pstA, phosphate ABC transporter, permease protein			6.650	2.484
GSU1098	pstC, phosphate ABC transporter, permease protein			6.427	1.601
GSU1099	pstS, phosphate ABC transporter, periplasmic phosphate-binding protein			4.974	2.073
GSU1340	ABC transporter, permease protein		Unknown substrate	11.121	3.052
GSU1331	Efflux transporter, RND family, MFP subunit		Metal ion	2.717	0.295
GSU1092	GH3 auxin-responsive promoter family protein	Unknown function	General	8.183	2.964
GSU1945	Fibronectin type III domain protein			3.058	0.801
GSU3435	Ankyrin repeat protein			2.849	0.776
GSU3464	gidA, glucose inhibited division protein A			2.543	0.551
GSU2427	Nitrite/sulfite reductase domain protein			2.185	0.303
GSU2780	Hypothetical protein			6.653	0.152
GSU1100	Hypothetical protein			6.017	1.412
GSU1944	Hypothetical protein			3.897	1.041
GSU1333	Hypothetical protein			3.582	0.525
GSU1994	Hypothetical protein			3.299	0.882
GSU0715	Hypothetical protein			2.813	1.104
GSU1943	Hypothetical protein			2.622	0.943
GSU0850	Hypothetical protein			2.415	0.084
GSU0790	Hypothetical protein			2.009	0.383

Results are sorted first by role category then in descending order by fold change. Fold changes are averages over all the biological and technical replicates.

are hypothetical proteins and proteins involved in transport (Table 2).

#### *Changes in pstB and phoU transcripts in chemostat cultures*

The significant increase in transcript abundance of *phoU* and *pstB* under phosphate-limiting conditions suggested that their abundance might serve as good

indicators for phosphate limitation during the growth of *Geobacter* species in the subsurface. This was first evaluated under controlled conditions with *G. sulfurreducens* grown in chemostats. Transcript abundance of *phoU* and *pstB* was normalized to transcript abundance of *proC*, a housekeeping gene that is expressed at consistent levels under many growth conditions (Holmes *et al.*, 2005). As expected from the microarray study, introducing

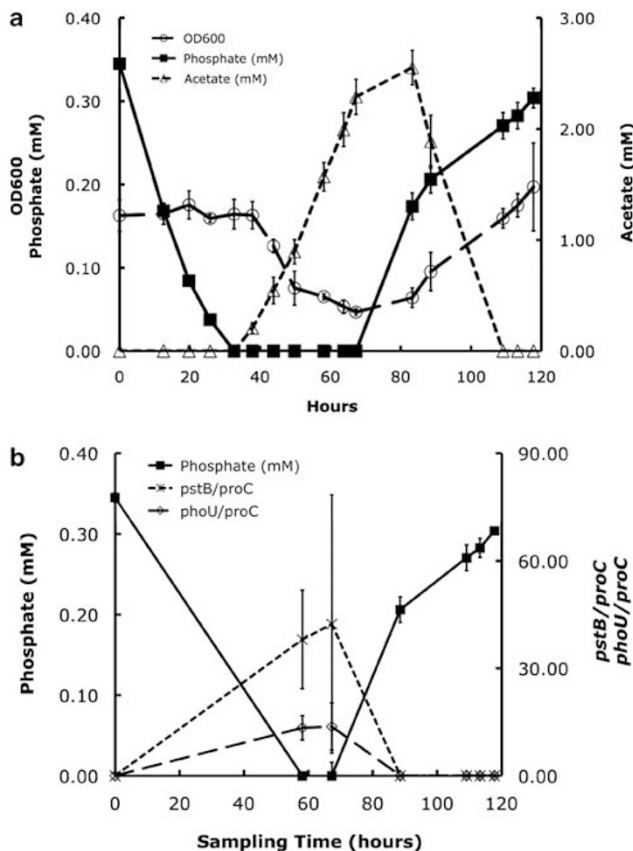
**Table 3** Reporters whose expression was significantly decreased when *Geobacter sulfurreducens* was grown with phosphate as limiting nutrient versus acetate as limiting nutrient

Locus ID	Common name	Main role	Subrole	Fold change	s.d.
GSU1906	leuA, 2-isopropylmalate synthase	Amino acid biosynthesis	Pyruvate family	-4.041	0.606
GSU1912	ilvD, dihydroxy-acid-dehydratase			-2.430	1.297
GSU1902	3-isopropylmalate dehydratase, small subunit			-2.357	0.286
GSU3068	murC, UDP-N-acetylmuramate—alanine ligase	Cell envelope	Biosynthesis of murein sacculus and peptidoglycan	-2.014	0.420
GSU3072	murF, UDP-N-acetylmuramoylalanyl-D-glutamyl-2, 6-diaminopimelate-D-alanyl-D-alanyl ligase, frameshifted			-2.014	0.420
GSU2707	ackA-1, acetate kinase	Central intermediate metabolism	Other	-2.463	0.377
GSU0096	recR, recombination protein RecR	DNA metabolism	DNA replication, recombination, and repair	-3.393	0.651
GSU0346	nuoI-1, NADH dehydrogenase I, I subunit	Energy metabolism	Electron transport	-4.103	0.903
GSU0347	nuoJ-1, NADH dehydrogenase I, J subunit			-3.119	0.596
GSU3259	Cytochrome c family protein			-3.083	0.639
GSU0339	nuoB, NADH dehydrogenase I, B subunit			-2.882	0.150
GSU0348	nuoK-1, NADH dehydrogenase I, K subunit			-2.803	0.853
GSU0341	nuoD, NADH dehydrogenase I, D subunit			-2.786	0.089
GSU0350	nuoM-1, NADH dehydrogenase I, M subunit			-2.474	0.079
GSU0345	nuoH-1, NADH dehydrogenase I, H subunit			-2.204	0.255
GSU0349	nuoL-1, NADH dehydrogenase I, L subunit			-2.083	0.481
GSU0097	Pyruvate ferredoxin/ferredoxin oxidoreductase		Fermentation	-2.872	0.348
GSU1738	Indolepyruvate ferredoxin/ferredoxin oxidoreductase, beta subunit			-2.433	0.354
GSU1737	Phenylacetate-CoA ligase		Other	-3.215	0.638
GSU1729	Phenylacetate-CoA ligase			-2.514	0.521
GSU1071	Conserved hypothetical protein	Hypothetical proteins	Conserved	-3.556	0.886
GSU0388	Conserved hypothetical protein			-3.505	0.482
GSU0317	Conserved hypothetical protein			-2.368	0.380
GSU3289	Conserved hypothetical protein (ferritin-like domain protein)			-2.307	0.318
GSU0955	group II intron, maturase	Other categories	Transposon functions	-2.303	0.329
GSU1772	ctpA-2, carboxy-terminal processing protease	Protein fate	Degradation of proteins	-2.466	0.530
GSU0276	Secretion protein, HlyD family		Protein and peptide	-2.101	0.221
GSU2835	MAP, methionine aminopeptidase, type I		Protein modification	-2.053	0.546
GSU2847	rplN, ribosomal protein L14	Protein synthesis	Ribosomal protein	-2.530	0.279
GSU2838	rplO, ribosomal protein L15			-2.413	0.355
GSU2865	rplJ, ribosomal protein L10			-2.351	0.256
GSU3235	rpmA, ribosomal protein L27			-2.278	0.875
GSU2852	rplV, ribosomal protein L22			-2.054	0.486
GSU3236	rplU, ribosomal protein L21			-2.046	0.177
GSU2840	rpsE, ribosomal protein S5			-2.028	0.512
GSU3465	trmE, tRNA modification GTPase TrmE		tRNA and rRNA base	-2.711	0.318
GSU0800	Amino acid ABC transporter, periplasmic amino acid-binding protein	Transport and binding proteins	Amino acids, peptides	-3.298	0.158
GSU1735	Branched-chain amino acid ABC transporter, periplasmic amino acid-binding protein, putative			-2.592	0.404
GSU1068	Sodium/solute symporter family protein		Unknown substrate	-4.411	0.395
GSU2352	Sodium/solute symporter family protein			-4.364	0.592
GSU1070	Sodium/solute symporter family protein			-4.016	0.146
GSU0490	Acetyl-CoA hydrolase/transferase family protein	Unknown function	Enzymes of unknown specificity	-4.131	0.653
GSU3454	Radical SAM domain protein			-2.211	0.755

**Table 3** (Continued)

Locus ID	Common name	Main role	Subrole	Fold change	s.d.
GSU1736	ACT domain protein		General	-2.500	0.539
GSU3092	YqeY family protein			-2.357	0.365
GSU0869	LysM domain/NLP/P60 family protein			-2.047	0.573
GSU1932	Hypothetical protein			-2.916	0.487
GSU0208	Hypothetical protein			-2.131	0.329
GSU2471	Hypothetical protein			-2.340	0.487

Results are sorted first by role category then in descending order by fold change. Fold changes are averages over all the biological and technical replicates.



**Figure 1** Phosphate limitation chemostats with *Geobacter sulfurreducens*. (a) Changes in cell density and acetate concentration in response to phosphate limitation. (b) Transcript abundance of *pstB* and *phoU* normalized to *proC* transcript abundance in various concentrations of phosphate. Results are means of triplicate chemostats.

medium without phosphate into the chemostat resulted in a decrease in cell density and an increase in acetate as growth became phosphate-limited (Figure 1a). When phosphate was reintroduced into the medium, *G. sulfurreducens* was able to grow again and as cell density increased, acetate concentrations decreased (Figure 1a).

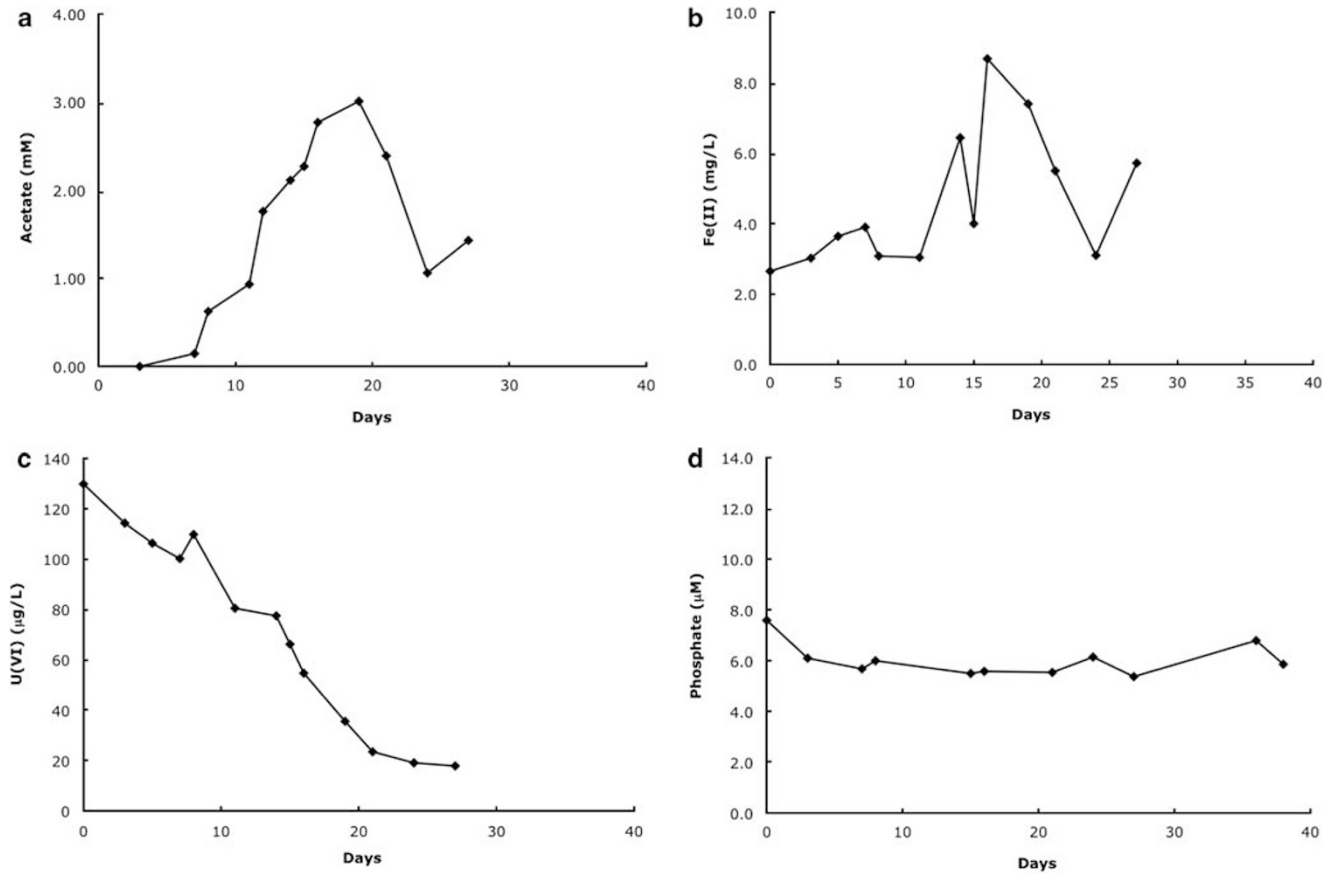
Transcripts of *phoU* and *pstB* were not detected during steady-state growth before imposing phosphate limitation (Figure 1b). Levels of *phoU* and *pstB* transcripts increased dramatically under phos-

phate-limiting conditions. Large s.d. were observed under phosphate limiting conditions when cell density was at its lowest and may have varied between triplicate chemostats. There were more *pstB* transcripts than *phoU* transcripts, consistent with the results from the microarray study. When the chemostat medium reservoir was switched back to phosphate-containing medium, the number of transcripts for both genes rapidly decreased back to baseline levels. These results further suggested that *phoU* and *pstB* expression levels are a sensitive indicator of phosphate availability.

#### Geochemical changes during in situ uranium bioremediation

As expected from previous field studies (Anderson *et al.*, 2003; Vrionis *et al.*, 2005), the addition of acetate to the subsurface resulted in increased acetate concentrations in the downgradient monitoring well (Figure 2a). Acetate concentrations declined temporarily between on day 10 and then began to increase as acetate inputs were stopped between on day 10 and then began to increase as acetate additions resumed by day 17. Because of the residual acetate in the injection well, the groundwater flush only resulted in a slight decrease in acetate concentration in D02. Increased acetate concentrations were associated with an increase in dissolved Fe(II) (Figure 2b) and a decrease in dissolved U(VI) (Figure 2c), thus suggesting that dissimilatory metal reduction had been stimulated. This was associated with a marked increase in the abundance of *Geobacter* species in the groundwater (Elifantz *et al.*, 2008; Mouser *et al.*, 2009b). *Geobacter* species are known Fe(III)- and U(VI)-reducing microorganisms and are considered to be responsible for most of the dissimilatory metal reduction in these acetate-amended subsurface sediments (Lovley, 1991; Finneran *et al.*, 2002; Holmes *et al.*, 2002, 2005; Anderson *et al.*, 2003; Istok *et al.*, 2004; Vrionis *et al.*, 2005). Within a period of 20 days of acetate addition, the uranium concentration in D02 was reduced by more than 80%, starting at an initial background concentration of about  $130 \mu\text{g l}^{-1}$  and diminishing to values well below the EPA maximum contaminant level ( $30 \mu\text{g l}^{-1}$ ) by the end of the study period (Figure 2c).





**Figure 2** Geochemical changes associated with the addition of acetate to the subsurface. Groundwater concentrations of acetate (a), Fe(II) (b), U(VI) (c) and phosphate (d) following the addition of acetate starting on day 0. Acetate inputs were interrupted from day 10 to day 17 and then resumed.

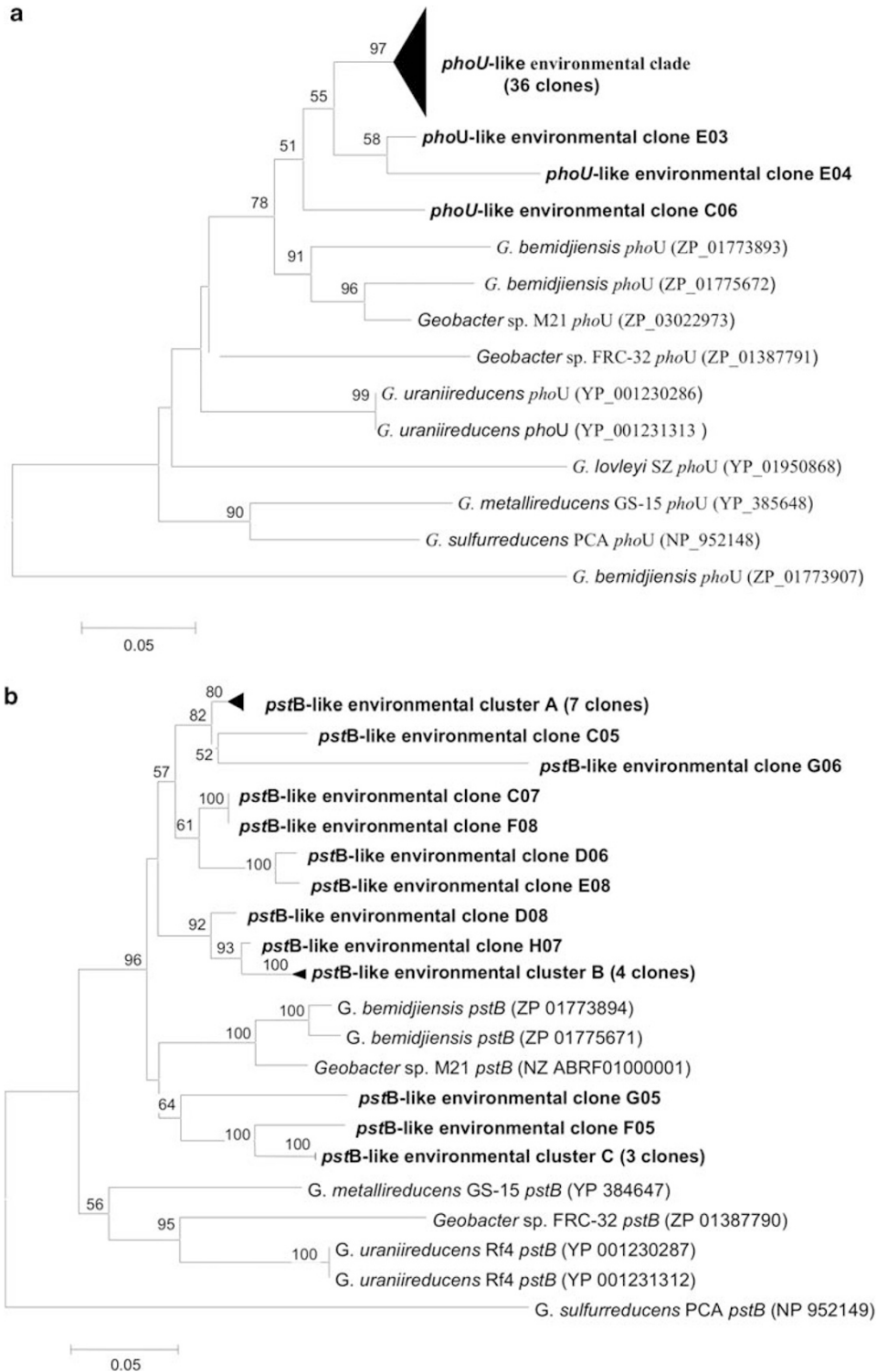
Phosphate concentrations remained relatively constant during the field experiment (Figure 2d). It was originally hypothesized that phosphate concentrations in the groundwater would increase during biostimulation as a result of the release of iron-bound phosphate species upon reduction of the iron oxides (Johnson and Loeppert, 2006). The constant concentration of phosphate in the groundwater suggests that dissolved phosphate concentrations are in equilibrium with solid phase phosphate. Therefore, as the microorganisms use phosphate, the loss of this nutrient from pore water is replaced by re-equilibration with solid phases.

*Diversity of the phoU and pstB genes amplified from the Rifle IFRC site*

To evaluate the diversity of phosphate metabolism genes found in the groundwater at the Rifle IFRC site, *phoU* and *pstB* genes were amplified from genomic DNA extracted from the groundwater on day 7 when dissimilatory metal reduction was very active and *Geobacteraceae* accounted for over 95% of the 16S rRNA gene sequences recovered in libraries constructed from groundwater samples (data not shown). Most (36 of 39) of the *phoU* sequences recovered from the groundwater sample

were in a tight gene cluster sharing 96% similarity with each other (Figure 3a). The other three environmental sequences were closely related and shared 87–93% similarity with the other environmental sequences (Figure 3a). The low diversity of *phoU* sequences was reflected in rarefaction curves (data not shown). At and below 98% similarity level, rarefaction curves suggested that the entire diversity of the *phoU* genes were covered in the clone library. The Shannon diversity index at the 98% similarity level was 1.1. This low diversity of *phoU* sequences is consistent with the low diversity of *Geobacter* species in this sample. The environmental samples clustered with the *phoU* from *Geobacter bemidjensis* and *Geobacter* sp. M21, an isolate from the Rifle site, and were 85–93% similar to these genes (Figure 3a).

The *pstB* genes recovered from groundwater were more diverse (Shannon diversity index 2.1) than *phoU*, but, like *phoU*, were closely related (77–87% similarity) to the *pstB* sequences of Subsurface Clade 1 microorganisms whose genomes are available (Figure 3b). Clusters A, B and C represented 29, 12.5 and 16.7% of the sequences analyzed, respectively (Figure 3b). Sequences within these clusters were at least 98% similar to each other, suggesting that this gene is highly conserved. As with the *phoU*



**Figure 3** Phylogenetic trees comparing *Geobacteraceae phoU* genes (a) and *pstB* genes (b) isolated from the Rifle IFC site with genes from *Geobacteraceae* available in pure culture. For simplicity, branches with more than 98% similarity were collapsed into a cluster. Branch points were supported by the Kimura algorithm and neighbor-joining distance analysis. Bootstrap values were calculated for 100 replicates and values above 50 are shown. The scale bars represent the expected number of changes per nucleotide.

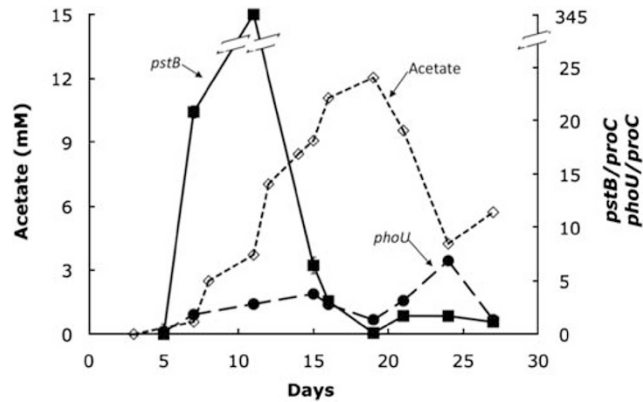
sequences, the *pstB* sequences recovered from the subsurface were most closely related to *pstB* sequences in pure cultures that have been recovered from the Rifle site and other subsurface environments.

#### Abundance of *phoU* and *pstB* transcripts in subsurface *Geobacter*

Transcripts of *phoU* and *pstB* in the subsurface *Geobacter* community were not detectable when

acetate amendments were initiated but increased in abundance when acetate availability increased, thereby stimulating the growth and activity of *Geobacter* species (Figure 4). With the exception of

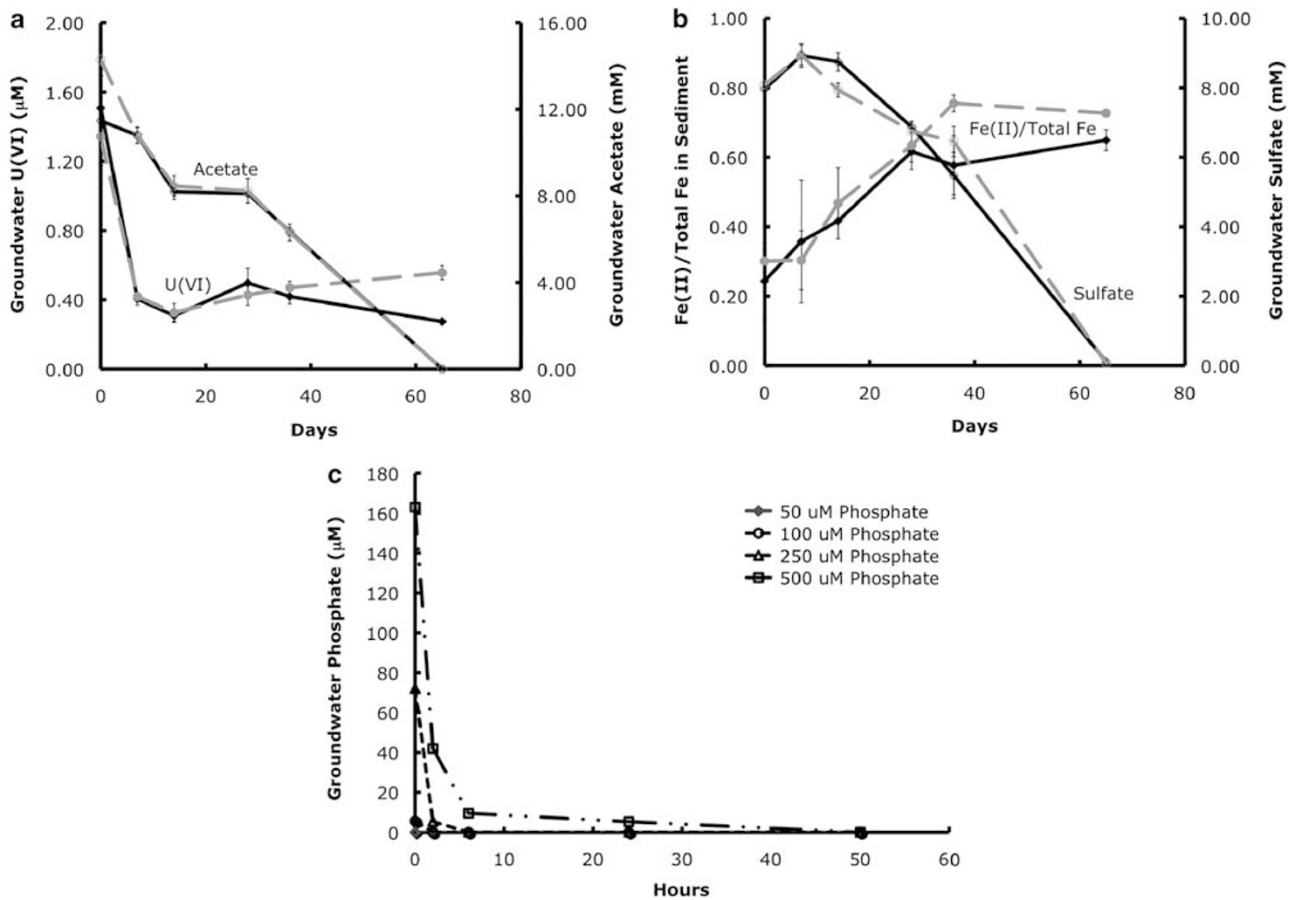
a very high abundance of *pstB* transcripts on day 11, the levels of *phoU* transcripts and *pstB* remained within a relatively restricted range, once acetate in the groundwater reached millimolar concentrations (Figure 4). With the exception of the one apparent anomaly in *pstB* abundance on day 11, the relative abundance of *phoU* and *pstB* was higher than that observed under phosphate-replete conditions in chemostats, but somewhat lower than the abundance of these transcripts under phosphate-limiting conditions. These results suggested that phosphate concentrations were less than those required for optimal growth during *in situ* uranium bioremediation or that the form of phosphorus was not as readily available as what might be provided in laboratory studies.



**Figure 4** Transcript abundance of *pstB* and *phoU* normalized to *proC* in D02 compared with acetate concentration during biostimulation of a uranium-contaminated aquifer in Rifle, CO, USA. Error bars represent the propagation of error of triplicate quantitative PCR (qPCR) reactions.

*Addition of phosphate in sediment incubations*

The finding that *Geobacter* species might be phosphate-limited suggested that the addition of phosphate to the subsurface might enhance the growth and activity of the subsurface *Geobacter* species and might stimulate rates of U(VI) reduction. Previous studies have demonstrated that the results



**Figure 5** Changes in geochemical parameters in sediment incubations with no nutrient amendment (solid lines) or amended with 250 μM phosphate (dashed lines). (a) Groundwater U(VI) and acetate concentrations; (b) fraction of Fe(II) in sediment; and (c) removal of phosphate from groundwater. Results are means of duplicate incubations.

from anoxic incubations of subsurface sediment and groundwater can provide an accurate indication of the impact of subsurface amendments (Finneran *et al.*, 2002; Holmes *et al.*, 2002; Ortiz-Bernad *et al.*, 2004a; N'Guessan *et al.*, 2008). Therefore, in order to evaluate whether adding phosphate might stimulate dissimilatory metal reduction at the Rifle site, sediment and associated groundwater, incubated under anoxic conditions, were amended with acetate to simulate *in situ* uranium bioremediation. As expected from previous studies (Finneran *et al.*, 2002; Ortiz-Bernad *et al.*, 2004a; N'Guessan *et al.*, 2008), consumption of the added acetate was associated with the reduction of Fe(III) to Fe(II) and the removal of U(VI) from the groundwater (Figure 5). Adding phosphate, which is calculated to increase the groundwater phosphate concentration to 250 µM, had no impact on the rate of acetate consumption, Fe(III) reduction or U(VI) removal (Figure 5). When phosphate concentrations were increased to potentially provide 500 µM phosphate, there was still no impact (data not shown).

One reason that the added phosphate failed to stimulate rates of acetate metabolism metal reduction may have been the high phosphate adsorption capacity of the sediment. Added phosphate was rapidly removed from the groundwater (Figure 5c). It has been shown that aquifer sediments generally have low phosphorus content and high phosphate sorption capacity (Patrick and Khalid, 1974; Mc Callister and Logan, 1978).

## Implications

These findings suggest that phosphate availability may limit the growth of *Geobacter* species during subsurface bioremediation and that the phosphate limitation cannot readily be alleviated with the direct addition of phosphate because of the high adsorptive capacity of sediments for phosphate. Despite this phosphate limitation U(VI) is effectively reduced, suggesting that the optimal growth of *Geobacter* species in the subsurface is not necessary for the success of this bioremediation strategy.

Phosphate is commonly added empirically as an amendment to stimulate subsurface bioremediation without information on whether the microorganisms involved in bioremediation are, in fact, phosphate-limited. An approach similar to the one described here may be useful for evaluating whether phosphate additions are warranted. For example, a strategy for promoting reductive dechlorination of chlorinated contaminants in subsurface environments is to add electron donors (Major *et al.*, 2002; Christ *et al.*, 2005), which, as with *in situ* uranium bioremediation, may result in phosphate limitations. However, the potentially complex geochemistry of phosphate must also be taken into account because as shown here, simple addition of phosphate may not materially increase the dissolved

phosphate available to the microorganisms. With no amendment of phosphate, phosphate limitation does not appear to be an impediment to achieving the goal of U(VI) reduction in field experiments performed to date. Over longer times of bioremediation (years vs months) it is possible that adsorbed phosphorus could be consumed, creating a more serious phosphate limitation. Future research might usefully focus on the time required to deplete bioavailable phosphate in sediments and on approaches to releasing sediment phosphate to make it more bioavailable.

## Acknowledgements

This research was supported by the Office of Science (BER), U. S. Department of Energy, Cooperative Agreement No. DE-FC02-02ER63446 and Grant No. DE-FG02-07ER64377, as well as LBNL-DOE Contract No. DE-AC02-05CH11231.

## References

- Aguena M, Yagil M, Spira B. (2002). Transcriptional analysis of the *pst* operon of *Escherichia coli*. *Mol Genet Genomics* **268**: 518–524.
- Allenby NEE, O'Connor N, Pragai Z, Carter NM, Miethke M, Engelmann S *et al.* (2004). Post-transcriptional regulation of the *Bacillus subtilis* *pst* operon encoding a phosphate-specific ABC transporter. *Microbiology* **150**: 2619–2628.
- Altschul SF, Gish W, Miller W, Myers EW, Lipman DJ. (1990). Basic local alignment search tool. *J Mol Biol* **215**: 403–410.
- Anderson RT, Vrionis HA, Ortiz-Bernad I, Resch CT, Long PE, Dayvault R *et al.* (2003). Stimulating the *in situ* activity of *Geobacter* species to remove uranium from the groundwater of a uranium-contaminated aquifer. *Appl Environ Microbiol* **69**: 5884–5891.
- Bardin SD, Finan TM. (1998). Regulation of phosphate assimilation in *Rhizobium* (*Sinorhizobium*) *meliloti*. *Genetics* **148**: 1689–1700.
- Bazylnski DA, Dean AJ, Schuler D, Phillips EJP, Lovley DR. (2000). N<sub>2</sub>-dependent growth and nitrogenase activity in the metal-metabolizing bacteria, *Geobacter* and *Magnetospirillum* species. *Environ Microbiol* **2**: 266–273.
- Botero LM, Al-Niemi TS, McDermott TR. (2000). Characterization of two inducible phosphate transport systems in *Rhizobium tropici*. *Appl Environ Microbiol* **66**: 15–22.
- Christ JA, Ramsburg CA, Abriola LM, Pennell KD, Löffler FE. (2005). Coupling aggressive mass removal with microbial reductive dechlorination for remediation of DNAPL source zones: A review and assessment. *Environ Health Perspect* **113**: 465–477.
- Coppi MV, Leang C, Sandler SJ, Lovley DR. (2001). Development of a genetic system for *geobacter* sulfur-reducers. *Appl Environ Microbiol* **67**: 3180–3187.
- Elifantz H, N'Guessan AL, Mouser PJ, Williams KH, Wilkins MJ, Ward JE *et al.* (2008). *American Society for Microbiology 108th General Meeting*, Boston, MA.
- Esteve-Nunez A, Rothermich M, Sharma M, Lovley D. (2005). Growth of *Geobacter sulfurreducens* under nutrient-limiting conditions in continuous culture. *Environ Microbiol* **7**: 641–648.

- Finneran KT, Anderson RT, Nevin KP, Lovley DR. (2002). Potential for bioremediation of uranium-contaminated aquifers with microbial U(VI) reduction. *Soil Sediment Contam* **11**: 339–357.
- Fischer RJ, Oehmcke S, Meyer U, Mix M, Schwarz K, Fiedler T *et al* (2006). Transcription of the *pst* operon of *Clostridium acetobutylicum* is dependent on phosphate concentration and pH. *J Bacteriol* **188**: 5469–5478.
- Gachter R, Meyer JS. (1993). The role of microorganisms in mobilization and fixation of phosphorus in sediments. *Hydrobiologia* **253**: 103–121.
- Gebhard S, Tran SL, Cook GM. (2006). The Phn system of *Mycobacterium smegmatis*: a second high-affinity ABC-transporter for phosphate. *Microbiology-Sgm* **152**: 3453–3465.
- Harris RM, Webb DC, Howitt SM, Cox GB. (2001). Characterization of PitA and PitB from *Escherichia coli*. *J Bacteriol* **183**: 5008–5014.
- Hinchee RE, Kittel JA, Reisinger HJ. (1995). Battelle Press: Columbus, OH (United States): United States, 547 pp.
- Holmes DE, Finneran KT, O'Neil RA, Lovley DR. (2002). Enrichment of members of the family Geobacteraceae associated with stimulation of dissimilatory metal reduction in uranium-contaminated aquifer sediments. *Appl Environ Microbiol* **68**: 2300–2306.
- Holmes DE, Nevin KP, Lovley DR. (2004a). Comparison of 16S rRNA, *nifD*, *recA*, *gyrB*, *rpoB* and *fusA* genes within the family Geobacteraceae fam. nov. *Int J Syst Evol Microbiol* **54**: 1591–1599.
- Holmes DE, Nevin KP, Lovley DR. (2004b). *In situ* expression of *nifD* in Geobacteraceae in subsurface sediments. *Appl Environ Microbiol* **70**: 7251–7259.
- Holmes DE, Nevin KP, O'Neil RA, Ward JE, Adams LA, Woodard TL *et al*. (2005). Potential for quantifying expression of the Geobacteraceae citrate synthase gene to assess the activity of Geobacteraceae in the subsurface and on current-harvesting electrodes. *Appl Environ Microbiol* **71**: 6870–6877.
- Istok JD, Senko JM, Krumholz LR, Watson D, Bogle MA, Peacock A *et al*. (2004). *In situ* bioreduction of technetium and uranium in a nitrate-contaminated aquifer. *Environ Sci Technol* **38**: 468–475.
- Johnson SE, Loeppert RH. (2006). Role of organic acids in phosphate mobilization from iron oxide. *Soil Sci Soc Am J* **70**: 222–234.
- Krol E, Becker A. (2004). Global transcriptional analysis of the phosphate starvation response in *Sinorhizobium meliloti* strains 1021 and 2011. *Mol Genet Genomics* **272**: 1–17.
- Lee PKH, Macbeth TW, Sorenson KS, Deeb RA, Alvarez-Cohen L. (2008). Quantifying genes and transcripts to assess the *in situ* physiology of “Dehalococoides” spp. in a trichloroethene-contaminated groundwater site. *Appl Environ Microbiol* **74**: 2728–2739.
- Lin B, Braster M, van Breukelen BM, van Verseveld HW, Westerhoff HV, Roling WFM. (2005). Geobacteraceae community composition is related to hydrochemistry and biodegradation in an iron-reducing aquifer polluted by a neighboring landfill. *Appl Environ Microbiol* **71**: 5983–5991.
- Lovley DR. (1991). Dissimilatory Fe(III) and Mn(IV) reduction. *Microbiol Rev* **55**: 259–287.
- Lovley DR. (2002). Analysis of the genetic potential and gene expression of microbial communities involved in the *in situ* bioremediation of uranium and harvesting electrical energy from organic matter. *OMICS* **6**: 331–339.
- Lovley DR. (2003). Cleaning up with genomics: applying molecular biology to bioremediation. *Nat Rev Microbiol* **1**: 35–44.
- Lovley DR, Baedeker MJ, Lonergan DJ, Cozzarelli IM, Phillips EJP, Siegel DL. (1989). Oxidation of aromatic contaminants coupled to microbial iron reduction. *Nature* **339**: 297–300.
- Major DW, McMaster ML, Cox EE, Edwards EA, Dworatzek SM, Hendrickson ER *et al*. (2002). Field demonstration of successful bioaugmentation to achieve dechlorination of tetrachloroethene to ethene. *Environ Sci Technol* **36**: 5106–5116.
- Martani E, Seto M. (1991). Effects of mineral nutrients on the degradation of 2,4-dichlorophenol at low concentrations by microbial community and bacterial isolates. *J Pesticide Sci* **16**: 1–7.
- Mc Callister DL, Logan TJ. (1978). Phosphate adsorption-desorption characteristics of soils and bottom sediments in the Maumee river basin of Ohio. *J Environ Qual* **7**: 87–92.
- Methe BA, Nelson KE, Eisen JA, Paulsen IT, Nelson W, Heidelberg JF *et al*. (2003). Genome of *Geobacter sulfurreducens*: Metal reduction in subsurface environments. *Science* **302**: 1967–1969.
- Methe BA, Webster J, Nevin K, Butler J, Lovley DR. (2005). DNA microarray analysis of nitrogen fixation and Fe(III) reduction in *Geobacter sulfurreducens*. *Appl Environ Microbiol* **71**: 2530–2538.
- Mouser PJ, Holmes DE, Perpetua LA, DiDonato R, Postier B, Liu A *et al*. (2009a). Quantifying expression of *Geobacter* spp. oxidative stress genes in pure culture and during *in situ* uranium bioremediation. *ISME J* **3**: 454–465.
- Mouser PJ, N'Guessan AL, Elifantz H, Holmes DE, Williams KH, Wilkins MJ *et al*. (2009b). Influence of heterogeneous ammonium availability on bacterial community structure and the expression of nitrogen fixation and ammonium transporter genes during *in situ* bioremediation of uranium-contaminated groundwater. *Environ Sci Technol* **43**: 4386–4392.
- Muda M, Rao NN, Torriani A. (1992). Role of *phoU* in phosphate transport and alkaline phosphatase regulation. *J Bacteriol* **174**: 8057–8064.
- N'Guessan AL, Vrionis HA, Resch CT, Long PE, Lovley DR. (2008). Sustained removal of uranium from contaminated groundwater following stimulation of dissimilatory metal reduction. *Environ Sci Technol* **42**: 2999–3004.
- Nakata A, Amemura M, Shinagawa H. (1984). Regulation of the phosphate regulon in *Escherichia coli* K-12—Regulation of the negative regulatory gene *phoU* and identification of the gene product. *J Bacteriol* **159**: 979–985.
- Nazaret S, Jeffrey WH, Saouter E, Von Haven R, Barkay T. (1994). *merA* gene expression in aquatic environments measured by mRNA production and Hg(II) volatilization. *Appl Environ Microbiol* **60**: 4059–4065.
- Nikata T, Sakai Y, Shibata K, Kato J, Kuroda A, Ohtake H. (1996). Molecular analysis of the phosphate specific transport (*pst*) operon of *Pseudomonas aeruginosa*. *Mol Gen Genet* **250**: 692–698.
- Norris RD, Dowd KD. *In situ* bioremediation of petroleum hydrocarbon-contaminated soil and groundwater in a low permeability aquifer. In: Flathman PE, Jerger DE, Exner JH (eds). *Bioremediation: Field Experience*. Boca Raton, Fla: Lewis Publishers, 1994, pp 457–474.

- North NN, Dollhopf SL, Petrie L, Istok JD, Balkwill DL, Kostka JE. (2004). Change in bacterial community structure during *in situ* biostimulation of subsurface sediment co-contaminated with uranium and nitrate. *Appl Environ Microbiol* **70**: 4911–4920.
- O'Neil RA, Holmes DE, Coppi MV, Adams LA, Larrahondo MJ, Ward JE *et al.* (2008). Gene transcript analysis of assimilatory iron limitation in Geobacteraceae during groundwater bioremediation. *Environ Microbiol* **10**: 1218–1230.
- Ortiz-Bernad I, Anderson RT, Vrionis HA, Lovley DR. (2004a). Resistance of solid-phase U(VI) to microbial reduction during *in situ* bioremediation of uranium-contaminated groundwater. *Appl Environ Microbiol* **70**: 7558–7560.
- Ortiz-Bernad I, Anderson RT, Vrionis HA, Lovley DR. (2004b). Vanadium respiration by *Geobacter metallireducens*: novel strategy for *in situ* removal of vanadium from groundwater. *Appl Environ Microbiol* **70**: 3091–3095.
- Patrick Jr WH, Khalid RA. (1974). Phosphate release and sorption by soils and sediments: effect of aerobic and anaerobic conditions. *Science* **186**: 53–55.
- Ronen Z, Bollag JM, Hsu CH, Young JC. (1996). Feasibility of bioremediation of a ground water polluted with alkyipyridines. *Ground Water* **34**: 194–199.
- Rooney-Varga JN, Anderson RT, Fraga JL, Ringelberg D, Lovley DR. (1999). Microbial communities associated with anaerobic benzene degradation in a petroleum-contaminated aquifer. *Appl Environ Microbiol* **65**: 3056–3063.
- Tamura K, Dudley J, Nei M, Kumar S. (2007). MEGA4: Molecular Evolutionary Genetics Analysis (MEGA) software version 4.0. *Mol Biol Evol* **24**: 1596–1599.
- Van Dien SJ, Keasling JD. (1999). Control of polyphosphate metabolism in genetically engineered *Escherichia coli*. *Enzyme Microb Technol* **24**: 21–25.
- VanBogelen RA, Olson ER, Wanner BL, Neidhardt FC. (1996). Global analysis of proteins synthesized during phosphorus restriction in *Escherichia coli*. *J Bacteriol* **178**: 4344–4366.
- Vera M, Guilian N, Jerez CA. (2003). Proteomic and genomic analysis of the phosphate starvation response of *Acidithiobacillus ferrooxidans*. *Hydrometallurgy* **71**: 125–132.
- Voegelé RT, Bardin S, Finan TM. (1997). Characterization of the *Rhizobium* (*Sinorhizobium*) *meliloti* high- and low-affinity phosphate uptake systems. *J Bacteriol* **179**: 7226–7232.
- Vrionis HA, Anderson RT, Ortiz-Bernad I, O'Neill KR, Resch CT, Peacock AD *et al.* (2005). Microbiological and geochemical heterogeneity in an *in situ* uranium bioremediation field site. *Appl Environ Microbiol* **71**: 6308–6318.
- Wilson MS, Bakermans C, Madsen EL. (1999). *In situ*, real-time catabolic gene expression: extraction and characterization of naphthalene dioxygenase mRNA transcripts from groundwater. *Appl Environ Microbiol* **65**: 80–87.
- Yabusaki SB, Fang Y, Long PE, Resch CT, Peacock AD, Komlos J *et al.* (2007). Uranium removal from groundwater via *in situ* biostimulation: field-scale modeling of transport and biological processes. *J Contam Hydrol* **93**: 216–235.
- Yan B, Methe BA, Lovley DR, Krushkal J. (2004). Computational prediction of conserved operons and phylogenetic footprinting of transcription regulatory elements in the metal-reducing bacterial family Geobacteraceae. *J Theor Biol* **230**: 133–144.