

ORIGINAL ARTICLE

Diversity of iron oxidizers in wetland soils revealed by novel 16S rRNA primers targeting *Gallionella*-related bacteria

Juanjuan Wang¹, Gerard Muyzer^{1,2}, Paul LE Bodelier¹ and Hendrikus J Laanbroek¹

¹Department of Microbial Wetland Ecology, Center for Limnology, Netherlands Institute of Ecology (NIOO – KNAW), Maarssen, The Netherlands and ²Department of Biotechnology, Environmental Biotechnology Group, Delft University of Technology, Delft, The Netherlands

Neutrophilic iron-oxidizing bacteria (FeOB) are important catalysts of iron cycling in wetland environments. However, little is known about their diversity and distribution in various environments. The aim of this study was to develop a PCR-DGGE assay enabling the detection of neutrophilic iron oxidizers in wetland habitats. Gradient tubes were used to enrich FeOB. From these enrichments, a clone library was established on the basis of the almost complete 16S rRNA gene using the universal bacterial primers 27f and 1492r. This clone library consisted of mainly α - and β -*Proteobacteria*, among which two major clusters were closely related to *Gallionella* spp. Specific probes and primers were developed on the basis of this 16S rRNA gene clone library. The newly designed *Gallionella*-specific 16S rRNA gene primer set 122f/998r was applied to community DNA obtained from three contrasting wetland environments, followed by Denaturing Gradient Gel Electrophoresis (DGGE) analysis. A second 16S rRNA gene clone library was constructed using the PCR products from one of our sampling sites amplified with the newly developed primer set 122f/998r. The cloned 16S rRNA gene sequences all represented novel culturable iron oxidizers most closely related to *Gallionella* sp. On the basis of their nucleotide sequences, four groups could be identified that were comparable to the DGGE banding pattern obtained before with the same PCR products as used for the second clone library. Using these *Gallionella*-specific 16S rRNA gene-based primers, in combination with DGGE, first insights into the diversity and distribution of these bacteria in wetland soils were obtained.

The ISME Journal (2009) 3, 715–725; doi:10.1038/ismej.2009.7; published online 19 February 2009

Subject Category: microbial ecology and functional diversity of natural habitats

Keywords: iron oxidation; *Gallionella*; wetlands; rhizosphere

Introduction

Iron is the fourth most abundant element in the Earth's crust and the most prevalent redox active metal (Hauck *et al.*, 2001). It participates in many key biological processes, such as photosynthesis, N₂ fixation, methanogenesis, H₂ production and consumption, respiration, the trichloroacetic acid cycle, oxygen transport, gene regulation and DNA biosynthesis (Andrews *et al.*, 2003). Iron also acts as an electron donor and acceptor for microbial energy generation. It has been the focus of many studies because of its importance in biogeochemical redox

reactions in the environments (Straub *et al.*, 2001; Emerson and Weiss, 2004; Roden, 2004).

Microorganisms can accelerate iron oxidation through a variety of mechanisms, such as changing local pH (Fortin *et al.*, 1998), passive physicochemical sorption (Hallbeck and Pedersen, 1990; McLean *et al.*, 1996; Ferris *et al.*, 1999; Hanert, 2002), and enzymatic conversion of iron to oxidized forms (Ghiorse, 1984; Widdel *et al.*, 1993). Iron-oxidizing bacteria (FeOB) have been detected in a wide variety of environments and have been shown to increase the rate of iron oxidation by up to four orders of magnitude compared with the rate of strictly abiotic oxidation (Søgaard *et al.*, 2001).

Compared with acidophilic iron oxidizers that have been studied extensively (Leduc and Ferroni, 1994; Auernik *et al.*, 2008) much less is known about the eco-physiology of FeOB inhabiting environments of neutral pH. This is largely due to the fact that their competitive ability in relation to abiotic oxidation still remains controversial (Sobolev and

Correspondence: J Wang, Department of Microbial Wetland Ecology, Center for Limnology, Netherlands Institute of Ecology (NIOO – KNAW), PO Box 1299, Maarssen 3600 BG, The Netherlands.

E-mail: j.wang@nioo.knaw.nl

Received 2 November 2008; revised 5 January 2009; accepted 5 January 2009; published online 19 February 2009

Roden, 2002), and to the difficulties associated with obtaining and maintaining them in pure culture (Edwards *et al.*, 2003a). In the last decade, a number of studies (Hallbeck and Pedersen, 1995; Emerson and Moyer, 2002; Neubauer *et al.*, 2002; Sobolev and Roden, 2002) showed that neutrophilic FeOB could successfully compete with abiotic oxidation for Fe (II); in some batch culture experiments, they accounted for 50–60% of the total Fe (II) oxidation.

The typical characteristic of these neutrophilic FeOB is their lithoautotrophic metabolism using iron as the only energy source. They are found growing at oxic-anoxic interfaces (Lutters-Czekalla, 1990) in neutral pH environments and are classified as micro-aerophilic bacteria. Among them, *Gallionella ferruginea* is known as the characteristic iron-oxidizing bacterium (Engel and Hanert, 1967; Hanert, 1968; Hallbeck and Pedersen, 1991). The rhizosphere of wetland plants has been shown to be one of the niches for these organisms because of the fact that plant roots release excessive oxygen into the soil (Armstrong, 1964; Neubauer *et al.*, 2007) creating suboxic conditions required by the iron oxidizers. Research on neutrophilic FeOB has recently moved forward with the isolation of a number of morphologically distinct species from different environments (Emerson and Moyer, 1997; Hanert, 2002; Edwards *et al.*, 2003b, 2004; Sobolev and Roden, 2004; Weiss *et al.*, 2007). Most of them were classified as β - or γ -proteobacteria.

Despite the progress in isolation and the biogeochemical evidence supporting their potential role in Fe-cycling, knowledge concerning their diversity and environmental distribution is still lacking. In this study, FeOB were enriched and identified. A 16S rRNA gene clone library of the enrichments was constructed and specific probes and primers were designed subsequently. The distribution of FeOB in wetland environments was further investigated by the application of specific primers targeting the 16S rRNA gene of *Gallionella*-related bacteria, in combination with PCR-DGGE, cloning and phylogenetic analysis.

Materials and methods

Soil sampling

Three sampling sites were chosen for this study. The first location, Appels (A; Belgium, 51° 2' 0" N, 4° 4' 0" E), is situated along the Scheldt estuary and has been characterized extensively (Lin *et al.*, 2007). It is an intertidal freshwater marsh area typically colonized by *Typha latifolia*, *Bolboschoenus maritimus* and *Phragmites australis*. Waarde (W; The Netherlands, 51° 25' 2" N 4° 4' 8" E) is a salt marsh where iron oxidation is very common and *Spartina anglica* is the dominating plant species. The sediments at Appels and Waarde contain significant amounts of ascorbate extractable Fe at near neutral pH. The pore water Fe (II) concentration reached values as high as

400 μM (Hyacinthe and Van Cappellen, 2004; Lin *et al.*, 2007). The third location, Huis Den Doorn (H; The Netherlands, 52° 32' 60" N, 6° 7' 60" E), is an irregularly flooded inland wetland site in the center of the Netherlands, which has a remarkably high concentration of Fe (total Fe 702 $\mu\text{mol g}^{-1}$ dry soil and oxalate-extracted Fe 459 $\mu\text{mol g}^{-1}$ dry soil) (Loeb *et al.*, 2008). *Glyceria maxima* is the dominating plant species.

Approximately 300 g of sieved soil (mesh size 1 mm) from each sampling site was put in a glass beaker (350 ml) and flooded with distilled water. One seedling (15 days old, two-leaf stage) of *Glyceria maxima* (for Appels and Huis den Doorn soil) or *Spartina anglica* (for Waarde soil) was transplanted into the glass beakers when the soil was completely saturated with water. Eight replicates for each soil type were prepared. The flood water level was kept constant at 2 cm above the soil surface by replenishing with sterile distilled water. After six weeks' growth at 20 °C (16 h/light and 8 h/dark) and 70% humidity, the plants were harvested. Soil from 1 cm below the soil surface (T), rhizosphere (R) and bulk (B) sections were sampled separately. A portion of each sample was used for enrichment of FeOB. Other aliquots were freeze-dried and stored for further molecular analysis.

Enrichment and isolation

Fresh soil samples were mixed at a ratio of 1:1 with Milli-Q water and used as inocula for enrichment of bacteria. Plant roots containing visible iron plaques were rinsed in sterile deionized water to remove adhering soil particles, cut into sections and used directly for inoculation.

Gradient tubes with opposing gradients (Emerson *et al.*, 1999) of oxygen and ferrous iron sulfide were used for the initial enrichment, isolation and subsequent maintenance of the culture. The system consisted of two layers of agarose in 16 mm screw-cap glass tubes. A bottom layer contained 1.25 ml FeS and Modified Wolfe Mineral Medium (Emerson and Floyd, 2005) at a ratio of 1:1, amended with 1% (wt/vol) agarose. The top layer consisted of 6.75 ml mineral medium supplemented with vitamins and minerals (<http://www.lgcstandards-atcc.org>, MD-VS, MD-TMS) and stabilized with 0.15% (wt/vol) agarose. Before autoclaving, the solutions for the upper layer were added together with NaHCO_3 to a concentration of 0.5 mM NaHCO_3 and flushed with CO_2 for 1 min to reach a final pH of 6.5.

The gradient tubes were always inoculated 24 h after preparation to facilitate the diffusion of Fe^{2+} into the top layer where bacterial growth occurs. The tubes were inoculated by pulling a micropipette vertically through the medium while the inoculum was expelled. The inoculum was serially diluted in sterile Modified Wolfe Mineral Medium. Abiotic controls were prepared without bacteria. Positive cultures were identified by observation of a discrete

band of Fe (III) oxide in contrast to the diffuse deposition of oxides in abiotic controls (Emerson and Moyer, 1997). The presence of bacteria was confirmed by both microscopy and molecular analysis.

DNA extraction and PCR

DNA was extracted using a modified DNA isolation procedure (Zhou *et al.*, 1996). The purification of DNA was carried out with DNA Clean and concentrator TM kit (Zymo Research, Orange, CA, USA). The quantity and quality of the extracted DNA was analyzed by spectrophotometry using NanoDrop ND-1000 TM (Nano-Drop Technologies, Wilmington, DE, USA) and by agarose gel electrophoresis. The genomic DNA was stored at -20°C for later use.

16S rRNA genes were amplified by PCR using different primer sets (see Table 1 for details).

Denaturing Gradient Gel Electrophoresis (DGGE)

Denaturing Gradient Gel Electrophoresis was performed with the Protean II system as described earlier (Muyzer *et al.*, 1993). An 8% polyacrylamide gel with a vertical gradient of 30–60% denaturants (100% denaturants constitute a mixture of 7 M urea and 40% (vol/vol) formamide) was used to analyze the 550 bp PCR products. The running conditions were 100 V at a constant temperature of 60°C in 23 l of $0.5 \times \text{TAE}$ buffer (20 mM Tris acetate, 0.5 mM EDTA, pH 8.0) for 18 h. The DGGE gels were visualized using a UV transilluminator after ethidium bromide staining. DNA fragments were excised from the gels, re-amplified, purified using the Gel DNA Recovery Kit (Zymoclean, Orange, CA, USA) and subjected to sequencing.

Clone library of 16S rRNA genes

Two clone libraries were constructed in this experiment using the TA cloning kit (Invitrogen, Carlsbad, CA, USA, version V). For analyzing the gradient tube enrichments, cloning of nearly complete 16S rRNA gene fragments was performed. For the application of *Gallionella*-specific primers on environmental soil samples, another clone library was constructed

with the PCR products obtained using the specific primers.

Ninety-six positive clones from each clone library were randomly picked for sequence analysis. The clones were inoculated into liquid LB medium amended with 20 mg ml^{-1} ampicillin and incubated on shakers at 37°C overnight. The cells were harvested and plasmid DNA was isolated using the Zippy plasmid miniprep kit (Zymo Research). The plasmid DNA was sequenced in both directions using M13F/M13R primers (Table 1) by MacroGen (<http://www.macrogen.co.kr>).

Comparative sequence analysis

Sequences of both excised DGGE bands and clones were manually checked and modified with Sequencher software (Sequencher 4.1.4) and then compared with existing sequences using the online database of BLAST (<http://www.ncbi.nlm.nih.gov/BLAST/Blast>). Sequences were aligned and phylogenetic trees were reconstructed in ARB (Ludwig *et al.*, 2004; Pruesse *et al.*, 2007) using the neighbor-joining algorithm (<http://www.arb-home.de/>). Probes for sequences that are closely related phylogenetically to defined iron-oxidizing species were designed using the probe design tool from the ARB package. The newly designed probes were aligned against sequences from the online sequence database (<http://rdp.cme.msu.edu/probematch/search.jsp>), and seven with less than 2 mismatches were chosen as specific probes.

On the basis of these probes, three primer sets were designed, which were 122F/998R, 122F/997R and 645F/996R. Primers 122F and 998R, specific to *Gallionella*-related FeOB, were selected for further analysis. The PCR conditions were optimized by running gradient PCRs at different annealing temperatures (50.3 , 51.9 , 53 , 54.3 , 55 and 58°C) and 55°C was found to yield products with the best quality and was used for this newly designed primer set. To test the specificity of the primers, plasmid DNA of clones of almost complete 16S rRNA gene sequences from non-related species (that is, MWE_C17, MWE_C40, MWE_N10, MWE_N19, MWE_C10 in Figure 4) were used as templates.

Table 1 16S rRNA gene PCR primers used in this study

Primer	Sequence (5'–3')	Application	Reference
341F-GC	CCTACGGGAGGCAGCAG	DGGE analysis	Muyzer <i>et al.</i> (1993)
907R	CCGTCAATTCMTTGTGATTT		Muyzer <i>et al.</i> (1993)
27F	GTGCTGCAGAGAGTTTGATCCTGGCTCAG	Clone library of gradient tube enrichments	Lane (1991)
1492R	CAGGGATCCTACGGGTACCTTGTACGACTT		Lane (1991)
M13F	GTAAAACGACGGCCAG	Sequencing DNA from clones	Huey and Hall (1989)
M13R	CAGGAAACAGCTATGAC		Huey and Hall (1989)
122F	ATATCGGAACATGTCCGG	Detecting <i>Gallionella</i> -like organisms in soil samples	This study
998R	CTCTGGAAACTTCTCTGAC		This study

Results

Enrichment of *FeOB*

Enrichment experiments started in winter time when no plant growth occurs in the field. Hence, plants were grown in growth chambers to enable differentiation between surface layers, rhizosphere and bulk soils. After 6 weeks of growth under controlled conditions, the soils in the beakers were fully rooted. Deposits of oxidized iron could be seen both around the roots and close to the soil surface (Figure 1a). Three days after inoculation of the soil samples into the gradient tubes, all the inoculated tubes showed growth. In each tube, one brownish iron oxide band mixed with bacterial biomass was formed approximately 1 cm below the surface. It was defined as growth because the shape and position of iron oxides were distinct from those in non-inoculated controls (as shown in Figure 1b). In the tubes inoculated with bulk or rhizosphere soil, similar iron oxide band patterns had been formed, whereas in tubes inoculated with fresh roots, the oxide band developed close to the head-space. Microscopic test and DNA isolation further confirmed bacterial growth associated with the iron oxide bands.

It was expected that pure cultures would be obtained after a few transfers, as described by Emerson and Moyer, (1997). However, in our case, a mixture of several species remained together till the last stage of isolation. Among them, one was closely related to *G. ferruginea* on the basis of the 16S rRNA gene information (96% sequence identity). With the increase in transfer times, the brown-colored iron bands developed much slower. It took approximately 24 h to see a bright orange iron band at the beginning of the enrichment, whereas after a few transfers, it took several days to form a thick white band. A brown color was observed only a week later. Typically, these bands were thinner and tended to move deeper toward the bottom of the tubes compared with the early gradient cultures.

DGGE analysis of enrichment cultures

The cultures obtained in the gradient tube during the first transfers were analyzed using PCR-DGGE on the basis of general bacterial primers. On average, 15 different DGGE bands were detected in the enrichments (Figure 2a), which were excised and sequenced. The sequences were quite diverse, including different species of α - and β -*Proteobacteria*, and *Bacteroidetes* as well as Firmicutes (Figure 3). After another four transfers, the total number of DGGE bands diminished in each of the samples (for example, Figure 2b).

Clone library

The enrichments from the gradient tubes (5–6 transfers of AT, HB and HT) were pooled and used

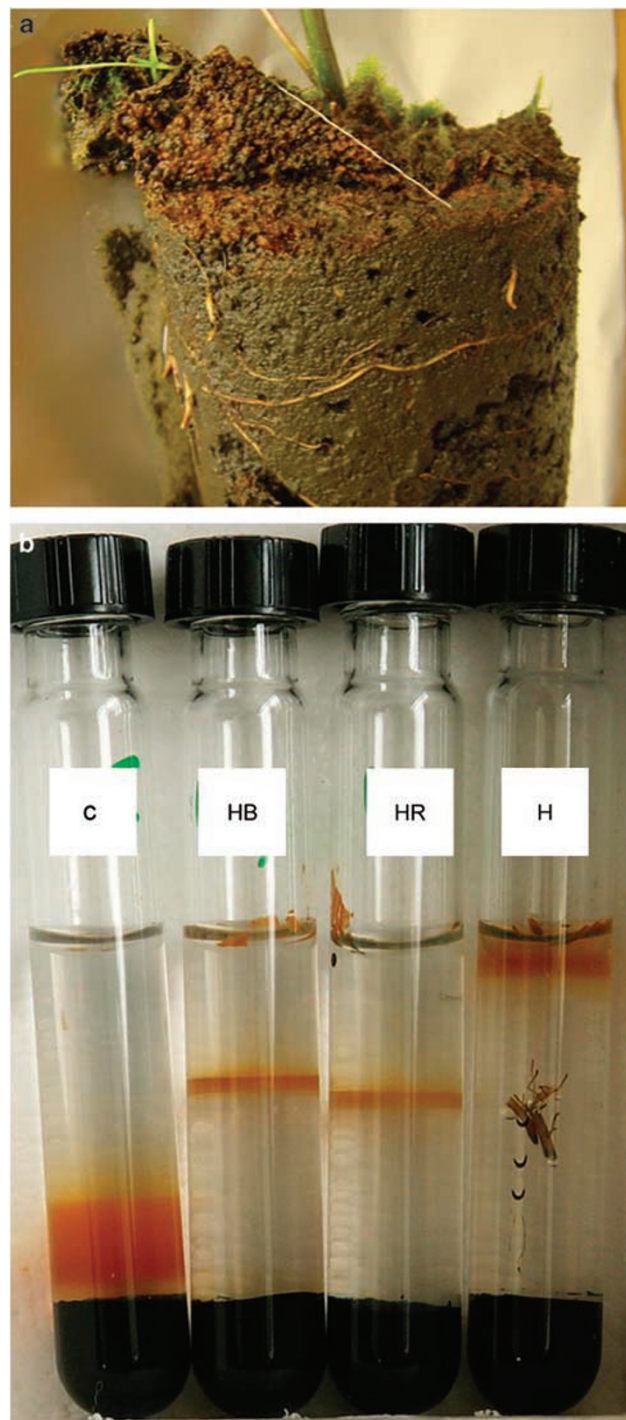


Figure 1 (a) Iron oxidation at the soil surface and around the roots of wetland plants. (b) Gradient tubes showing bacterial (HB, HR, H) and chemical iron oxidation (C). HB: Huis den Doorn bulk soil, HR: Huis den Doorn rhizosphere soil, H: Huis den Doorn root inoculum.

for DNA isolation and cloning. In total, 96 clones were screened. Figure 4 shows the phylogenetic affiliation of the different clone types (coded MWE) found in the enrichments. In the clone library, over 20% of the clones were closely related to *G. ferruginea* (Hallbeck *et al.*, 1993), which is a

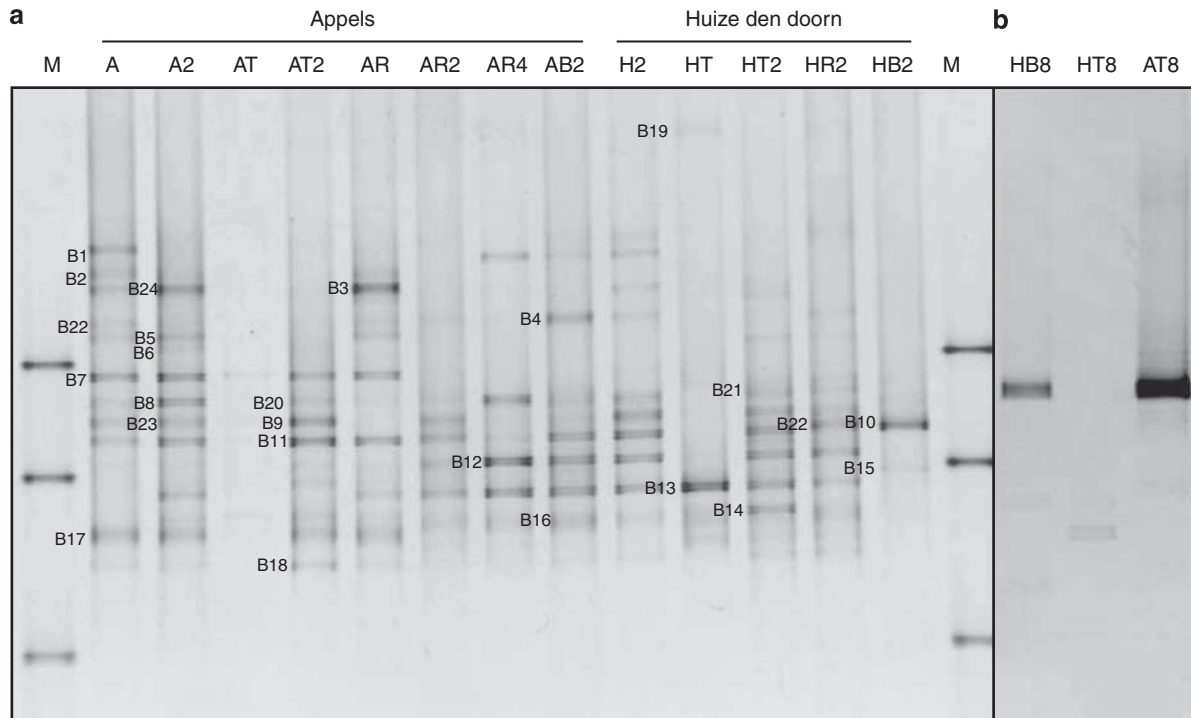


Figure 2 Denaturing gradient gel electrophoresis (DGGE) patterns showing diverse bacterial communities after the first (a) and later (b) transfers in gradient tube enrichments. A and H refer to the sampling sites Appels and Huis Den Doorn, respectively; B, T and M indicate bulk soil, surface soil and markers, respectively; numbers refer to numbers of transferring times. B1–B24 are cut and sequenced bands, and the sequence information can be found in Figure 3.

typical neutrophilic iron oxidizer (Hallberg and Ferris, 2004). They clustered either with *Siderooxydans lithotrophicus* LD-1 and *Siderooxydans lithoautotrophicus* ES-1, which are two recently identified iron oxidizers (Weiss *et al.*, 2007), or with uncultured *Gallionellaceae* (Winderl *et al.*, 2008).

Apart from the iron oxidizer group, in the clone library, we also found sequences that are related to other genera such as *Azospira*, *Rhodocyclaceae*, *Janthinobacterium*, *Thiobacillus* (Figure 4).

Design of specific primers and probes

Sequences from three clusters of interest (see arrows in Figure 4), including six representative clones on the bases of their assumed iron-oxidizing capacity, were chosen for specific probe or primer development. MWE_C19, MWE_N34, MWE_N10 and MWE_C13 fall into the *Gallionella* group. MWE_C19 and MWE_N34 share 95% similarity to *G. ferruginea* and 98%, similar to *Gallionella capsiferiformans* (Weiss *et al.*, 2007), whereas MWE_N10 and MWE_C13 are closely related to *Siderooxydans* species. MWE_C10, MWE_C12 are related to some uncultured β -*Proteobacterium* found in iron-rich areas (Juretschko *et al.*, 2002). In total, seven probes (that is, P1–P7) were developed and their specificity was checked using a probe-matching program (Table 2). Finally, three primer sets were designed on the basis of the probes, that is, 122F/998R, 122F/997R and 645F/997R. Primer set 122F/998R is

specific to two almost identical sequences that share 98% similarity to a recently described iron-oxidizing bacterium *Gallionella capsiferiformans* on the basis of the 16S rRNA gene. Clones MWE_N10 and MWE_C13, clustering with *Siderooxydans* spp., share the same forward primer 122F with clones MWE_C19 and MWE_N34, but have their own specific reverse primer 997R. The third primer set 645F/997R was designed for clones MWE_C10 and MWE_C12, which have uncultured *Rhodocyclaceae* spp. as their closest relative.

Detection of *Gallionella*-like bacteria in soil samples

Confirmed to be specific for one novel cluster of culturable iron oxidizers, the primer set 122f/998r designed for *Gallionella*-like organisms was then used to detect these organisms in environmental samples. After a nested-PCR using the universal Eubacterial primer set 3G/907R, the PCR products were checked on DGGE gel. The result showed that these organisms were present in most of the soil samples tested (Figure 5). The bands migrated to or around the position of the marker band that belongs to the clone on which the primers were designed (Marker c in Figure 5). Differences in band patterns can be seen among the samples. As to the samples from Huis den Doorn, no bands were detected from the bulk soil (HB) and rhizosphere (HR), whereas the most intensive band was obtained in the surface soil (HT). Appels soil sample (AB, AR and AT) appeared

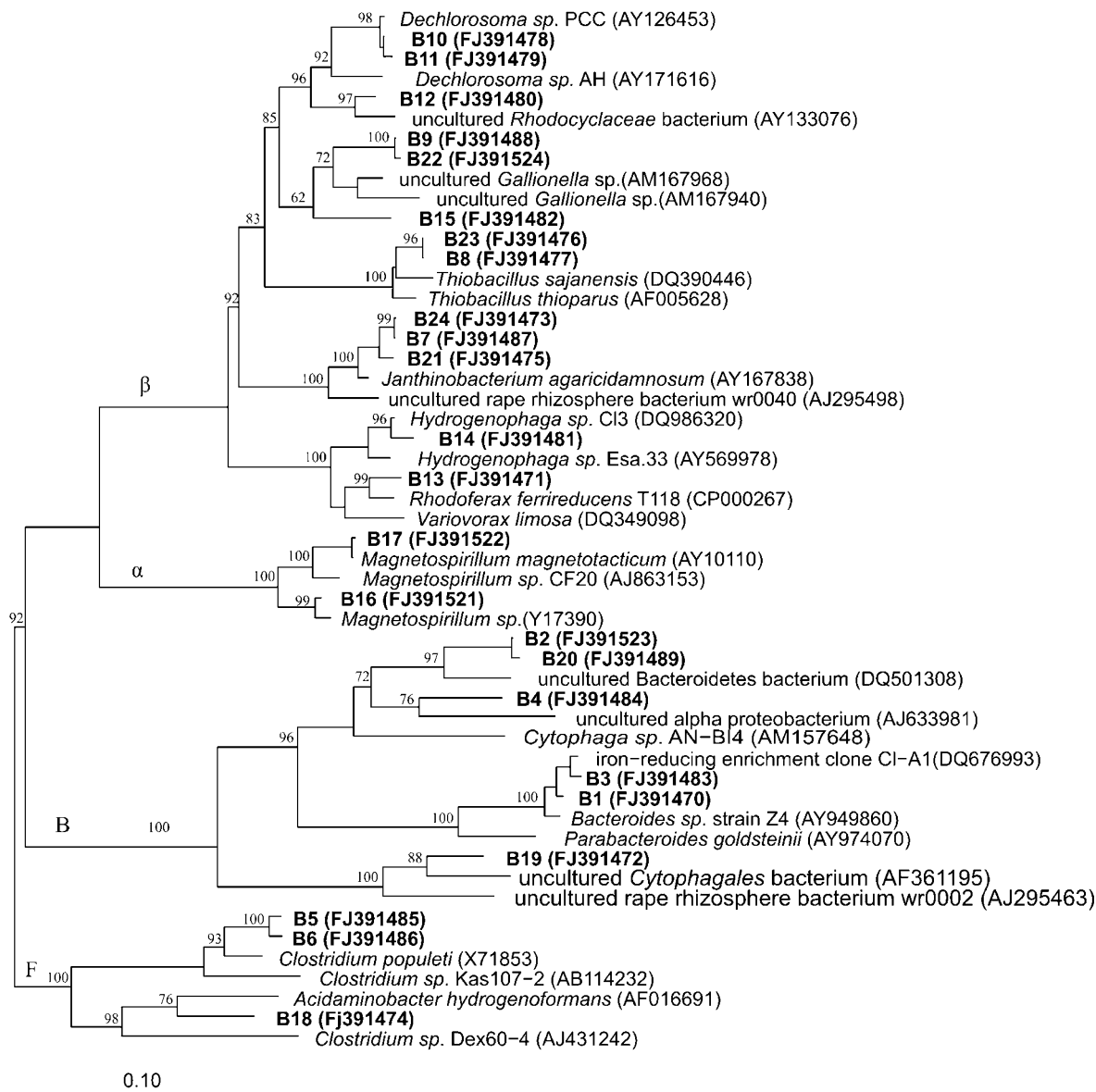


Figure 3 Neighbor-joining tree of sequences obtained from gradient tube enrichments shown in Figure 2a. Sequences determined in this study are printed in bold type. Bootstrap values are shown near the nodes (based on 1000 replicate trees). The bar indicates 10% sequence difference. *Thermotoga maritima* was used as an outgroup, but was pruned from the tree. α = Alphaproteobacteria, β = Betaproteobacteria, B = Bacteroidetes, F = Firmicutes.

to be more diverse with respect to the presence of *Gallionella*-like bacteria than the other two soils.

To validate the PCR-DGGE method, a clone library was constructed using the PCR products of the *Gallionella*-specific primers (122F and 998R). Soil DNA was taken from one of the sampling sites, Huis den Doorn. Within 96 sequenced clones retrieved from these soil samples, four clusters could be identified for their differences in nucleotide sequences. Twelve sequences were chosen as representatives to be included in the phylogenetic tree on the basis of the clones from the gradient tube enrichments (Figure 4). Compared with the other clones, Hc8 is less closely related to *G. ferruginea* and other

clones. Clones Hc37, Hc11 and Hc16 are identical to some uncultured *Gallionella* spp., whereas clones Hc9, Hc25, Hc30 and Hc1 share more similar sequences. Clones Hc18, Hc46 and Hc38 shared 97% similarity with *G. ferruginea*. By comparing the obtained sequences, the most discriminative variable region could be identified, which corresponds to *Escherichia coli* 16S rRNA range 455–479. These 12 representative sequences were checked again on DGGE (Figure 6). According to their migrating positions in the gradient gel, different groups could be distinguished. This was compared with the result obtained with the nested PCR-DGGE method (Lane HT in Figure 5). As expected, more

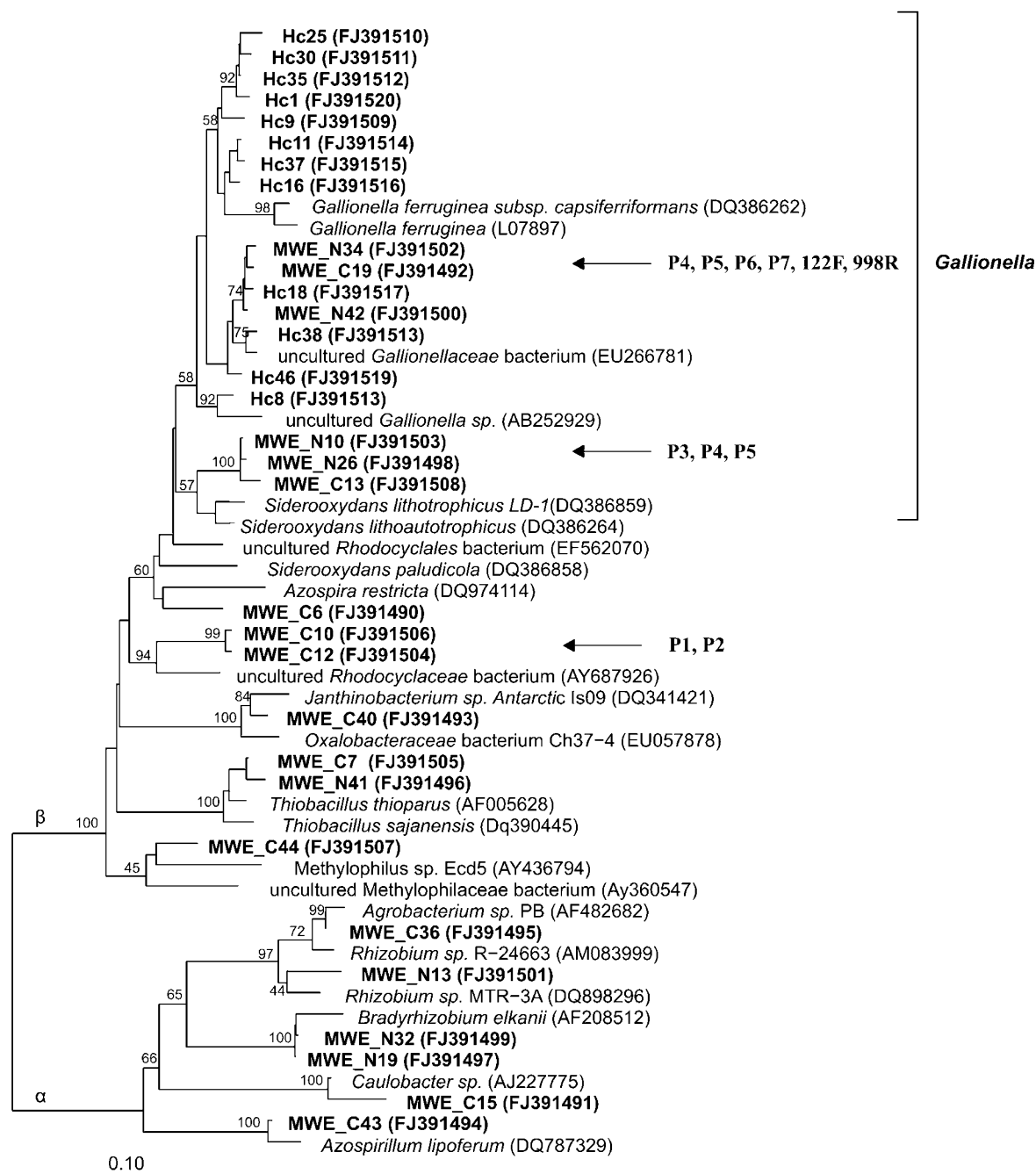


Figure 4 Phylogenetic tree of 16S rRNA-based clones derived from gradient tube enrichments using the 27f/1492r primer set (coded MWE), as well as from iron-rich Huis den Doorn soil samples using primer 122F/998R developed in this study (coded Hc). The sequences in bold were selected clones. Bootstrap values are shown near the knots (based on 1000 replicate trees). The bar indicates 10% sequence difference. *Thermotoga maritima* was used as an outgroup, but was pruned from the tree. α = *Alphaproteobacteria*, β = *Betaproteobacteria*. Arrows indicate the clones based on which specific probes and primers developed.

bands were obtained after cloning, but most of them ended up at the same positions.

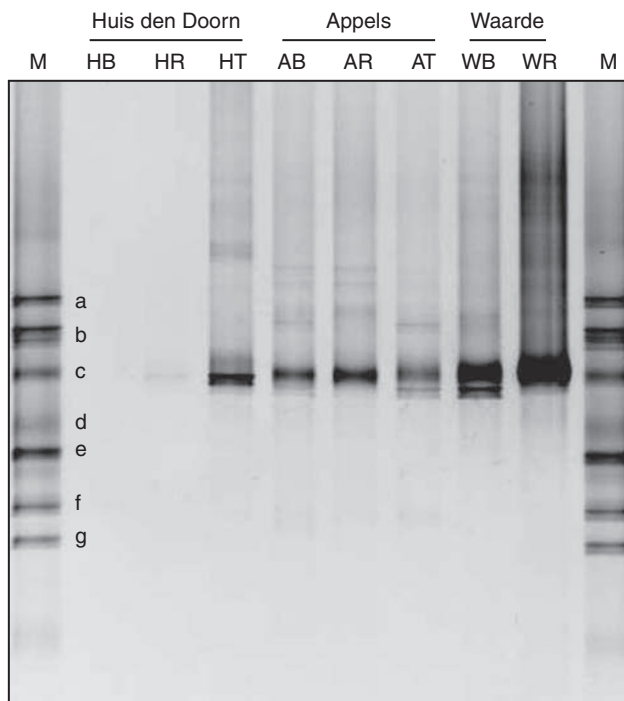
Discussion

Microbial iron oxidation under neutral conditions has been reported earlier (Hallbeck and Pedersen, 1990; Straub *et al.*, 2001; Sobolev and Roden, 2002), but

progress in this research area is hampered by the difficulty of obtaining the responsible organisms in pure culture. The development of the gradient tube technique (Emerson and Floyd, 2005) made it possible to obtain these bacteria in pure cultures. In this study, it was observed that when fresh roots were used as inoculation, they formed iron oxide bands close to the headspace where the oxygen concentration is relatively high. This is probably due to the depletion of

Table 2 Specific 16S rRNA-targeted oligonucleotide probes designed according to the clone library from the mixed enrichment cultures of the gradient tubes

Probe name	Probe Sequence (5'–3')	Position	Clone specificity
P1_996	CAGCAGGATTCCAGGCAT	996–1013	MWE_C10, MWE_C12
P2_645	GCCACACTCGAGTCTTGC	645–662	MWE_C10, MWE_C12
P3_997	TCTTCTGGATTCTCGGCA	997–1014	MWE_N10, MWE_C13
P4_460	CACTCCACGTATAGGTG	460–477	MWE_N10, MWE_C13
P5_122	CCGGACATGTTCCGATAT	122–140	MWE_C19, MWE_N34
P6_645	GCTACACTCTAGCCTTCC	645–662	MWE_C19, MWE_N34
P7_998	CTCTGGAAACTTCCTGAC	998–1015	MWE_C19, MWE_N34

**Figure 5** Denaturing gradient gel electrophoresis (DGGE) patterns showing DNA of *Gallionella*-like iron oxidizers from different wetland environments being amplified by using a nested approach (*Gallionella*-specific primers and universal DGGE primers 341GC/907R, respectively). A and H refer to the sampling sites Appels and Huis Den Doorn, respectively; B, T and M indicate bulk soil, surface soil and markers, respectively. Marker sequences are clones from the library of the gradient tube enrichments; from a to g: MWE_C40, MWE_N41, MWE_N34, MWE_N13, MWE_N19, MWE_C13 and MWE_N26 (see also Figure 4). Marker c indicates the clone from which the *Gallionella*-specific primers are developed.

oxygen by root respiration, which results in the upward movement of micro-aerophilic bacteria.

The success of isolation depends in large part on experience. In our study, the repeated dilution and inoculation of gradient tubes reduced the bacterial diversity significantly (as can be seen comparing Figures 2a and b). With the increase of the time between transfers, the iron oxide bands in the gradient tubes developed much slower and were thinner, which might be the consequence of a decrease in the number of many non-dominant

FeOB, as well as of those microorganisms that might benefit the growth of the remaining iron-oxidizing species. A pure iron-oxidizing isolate was obtained in this study, which shared the same 16S rRNA sequences as the clone used for designing the *Gallionella*-specific primers.

Despite its advantages of accuracy and efficiency, molecular tools are hardly applied in the study of FeOB in environments of approximately neutral pH, mainly because of the lack of sequence information to design specific primers or probes for conservative regions of the 16S RNA genes (Konstantinidis and Tiedje, 2007) as well as for functional genes (Parro *et al.*, 2007). We developed 16S rRNA gene-based probes and primers, which can be used directly on environmental samples to study the occurrence of these microbes and their temporal and spatial distribution. In combination with laboratory cultivation, this opens up the ecological studies on this environmentally relevant group of bacteria.

By using a nested PCR-DGGE approach, we were able to detect *Gallionella*-like organisms in different compartments of freshwater as well as brackish wetland environments. As shown in Figure 5, the bands ended up at slightly different positions in DGGE gels, showing that the diversity of *Gallionella*-related iron oxidizers is higher than recently known. The slightly higher diversity revealed by cloning compared with a nested PCR-DGGE approach showed that cloning has decreased the detection limit as compared with DGGE. In fact, in this study, the species detected by cloning, but missed by DGGE, that is, Hc8, Hc9 and Hc25, count for only 5% of the total clones obtained. Hc8 is the only one out of the 96 clones that are distantly related to the other sequences. The low density of this species might increase the difficulty in detecting it by DGGE.

It was reported that in wetland environments iron oxidation mainly occurs close to the plant roots where sufficient oxygen is available owing to its release (Neubauer *et al.*, 2007). We have indeed detected *Gallionella*-related organisms not only in the rhizosphere but also in the bulk soil. Detection based on the 16S rRNA gene indicates the presence but not the activity of microbial cells. Hence, cells detected in the rhizosphere may still be more active than those found in the bulk soil.

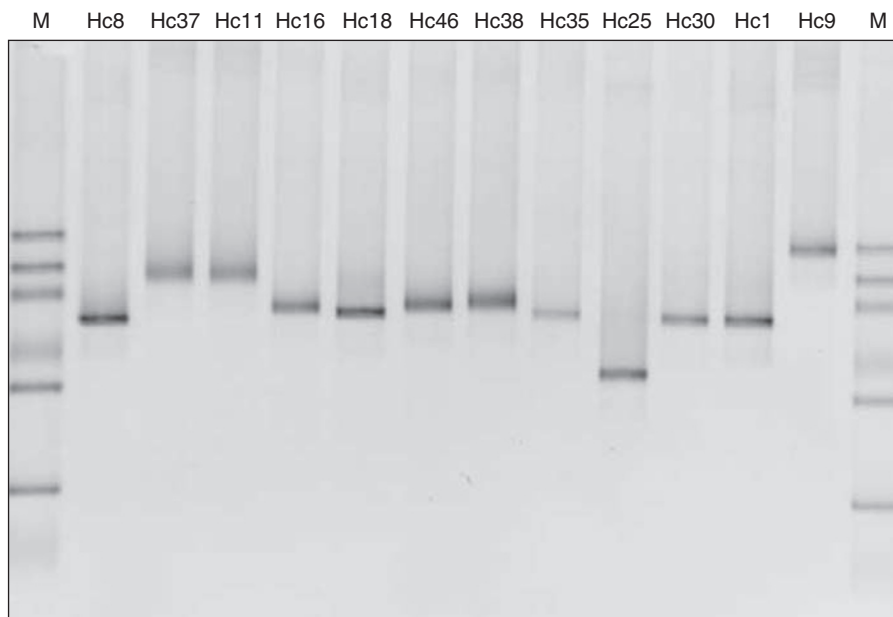


Figure 6 Denaturing gradient gel electrophoresis (DGGE) patterns showing cloned sequences amplified with the *Gallionella*-specific primers 122F and 998R. Corresponding clones can be found in Figure 4. Hc refer to the codes of the clones, and M is the marker (same as Figure 5).

Iron oxidizers are phylogenetically so diverse that they fall in evolutionarily distinct groups (Roden *et al.*, 2004). Beside the facts that not many culturable FeOB are available, the known sequences are so diverse that it is almost not possible to design group-specific probes or primers for them. With our specific primers, four different groups were selectively picked up from environmental wetland samples. The comparison of sequences showed that they vary in a specific region. Therefore, primers can be developed to detect *Gallionella*-related sequences separately from each other using real time-PCR.

To our knowledge, this is the first study using specific primers for detecting neutrophilic FeOB in environmental samples. Although bacterial iron oxidation has been claimed to be important especially in wetland environments (Neubauer *et al.*, 2007), studies on this subject are still limited. The lack of efficient molecular tools to study these microbes seriously hampers progress in this field. This fact will make the detection of their distributions in the environments a challenge. Isolation of pure cultures and sequencing of their genomes will help to develop and define primers on the basis of functional genes, which will also facilitate mRNA studies on these microbes.

Acknowledgements

We thank Dr Marzia Milletto for providing Huis Den Doorn soil samples. Dr Jeanine Geelhoed is acknowledged for sharing her experience with FeOB. This work was financially supported by the Netherlands Darwin Center for Biogeology (Grant number 142.16.1031).

References

- Andrews SC, Robinson AK, Rodriguez-Quinones F. (2003). Bacterial iron homeostasis. *FEMS Microbiol Rev* **27**: 215–237.
- Armstrong W. (1964). Oxygen diffusion from roots of some British bog plants. *Nature* **204**: 801–802.
- Auernik KS, Maezato Y, Blum PH, Kelly RM. (2008). The genome sequence of the metal-mobilizing, extremely thermoacidophilic archaeon *Metallosphaera sedula* provides insights into bioleaching-associated metabolism. *Appl Environ Microbiol* **74**: 682–692.
- Edwards KJ, Bach W, McCollom TM, Rogers DR. (2004). Neutrophilic iron-oxidizing bacteria in the ocean: their habitats, diversity, and roles in mineral deposition, rock alteration, and biomass production in the deep-sea. *Geomicrobiol J* **21**: 393–404.
- Edwards KJ, Bach W, Rogers DR. (2003a). Geomicrobiology of the ocean crust: a role for chemoautotrophic Fe-bacteria. *Biol Bull* **204**: 180–185.
- Edwards KJ, Rogers DR, Wirsen CO, McCollom TM. (2003b). Isolation and characterization of novel psychrophilic, neutrophilic, Fe-oxidizing, chemolithoautotrophic α - and γ -*Proteobacteria* from the deep sea. *Appl Environ Microbiol* **69**: 2906–2913.
- Emerson D, Floyd MM. (2005). Enrichment and isolation of iron-oxidizing bacteria at neutral pH. *Methods Enzymol* **397**: 112–123.
- Emerson D, Moyer C. (1997). Isolation and characterization of novel iron-oxidizing bacteria that grow at circumneutral pH. *Appl Environ Microbiol* **63**: 4784–4792.
- Emerson D, Moyer CL. (2002). Neutrophilic Fe-oxidizing bacteria are abundant at the Loihi seamount hydrothermal vents and play a major role in Fe oxide deposition. *Appl Environ Microbiol* **68**: 3085–3093.
- Emerson D, Weiss JV. (2004). Bacterial iron oxidation in circumneutral freshwater habitats: findings from the field and the laboratory. *Geomicrobiol J* **21**: 405–414.

- Emerson D, Weiss JV, Megonigal JP. (1999). Iron-oxidizing bacteria are associated with ferric hydroxide precipitates (Fe-plaque) on the roots of wetland plants. *Appl Environ Microbiol* **65**: 2758–2761.
- Engel H, Hanert H. (1967). Isolierung von *Gallionella ferruginea* ehrenberg. *Naturwissenschaften* **54**: 147.
- Ferris FG, Konhauser KO, Lyven B, Pedersen K. (1999). Accumulation of metals by bacteriogenic iron oxides in a subterranean environment. *Geomicrobiol J* **16**: 181–192.
- Fortin D, Ferris FG, Scott SD. (1998). Formation of Fe-silicates and Fe-oxides on bacterial surfaces in samples collected near hydrothermal vents on the Southern Explorer Ridge in the northeast Pacific ocean. *American Mineralogist* **83**: 1399–1408.
- Ghiorse WC. (1984). Biology of iron- and manganese-depositing bacteria. *Annul rev microbiol* **38**: 515.
- Hallbeck L, Pedersen K. (1990). Culture parameters regulating stalk formation and growth-rate of *Gallionella-ferruginea*. *J Gen Microbiol* **136**: 1675–1680.
- Hallbeck L, Pedersen K. (1991). Autotrophic and mixotrophic growth of *Gallionella ferruginea*. *J Gen Microbiol* **137**: 2657–2661.
- Hallbeck L, Pedersen K. (1995). Benefits associated with the stalk of *Gallionella ferruginea*, evaluated by comparison of stalk-forming and a non-stalk forming strain and biofilm studies *in-situ*. *Micro Ecol* **30**: 257–268.
- Hallbeck L, Stahl F, Pedersen K. (1993). Phylogeny and phenotypic characterization of the stalk-forming and iron-oxidizing bacterium *Gallionella ferruginea*. *J Gen Microbiol* **139**: 1531–1535.
- Hallberg R, Ferris FG. (2004). Biomineralization by *Gallionella*. *Geomicrobiol J* **21**: 325–330.
- Hanert H. (1968). Investigations on isolation physiology and morphology of *Gallionella ferruginea* ehrenberg. *Archiv Fur Mikrobiologie* **60**: 348–376.
- Hanert HH. (2002). Bacterial and chemical iron oxide deposition in a shallow bay on Palaea Kameni, Santorini, Greece: microscopy, electron probe microanalysis, and photometry of *in situ* experiments. *Geomicrobiol J* **19**: 317–342.
- Hauck S, Benz M, Brune A, Schink B. (2001). Ferrrous iron oxidation by denitrifying bacteria in profundal sediments of a deep lake (Lake Constance). *FEMS Microbiol Ecol* **37**: 127–134.
- Huey B, Hall J. (1989). Hypervariable DNA fingerprinting in *Escherichia Coli*: minisatellite probe from bacteriophage-M13. *J Bacteriol* **171**: 2528–2532.
- Hyacinthe C, Van Cappellen P. (2004). An authigenic iron phosphate phase in estuarine sediments: composition, formation and chemical reactivity. *Mar Chem* **91**: 227–251.
- Juretschko S, Loy A, Lehner A, Wagner M. (2002). The microbial community composition of a nitrifying-denitrifying activated sludge from an industrial sewage treatment plant analyzed by the full-cycle rRNA approach. *Syst Appl Microbiol* **25**: 84–99.
- Konstantinidis KT, Tiedje JM. (2007). Prokaryotic taxonomy and phylogeny in the genomic era: advancements and challenges ahead. *Curr Opin in Microbiol* **10**: 504–509.
- Lane DJ. (1991). 16S/23S rRNA sequencing. In: E Stackebrandt, M Goodfellow (eds). *Nucleic Acid Techniques in Bacterial Systematics*. Wiley: New York, pp 115–175.
- Leduc LG, Ferroni GD. (1994). The chemolithotrophic bacterium *Thiobacillus ferrooxidans*. *FEMS Microbiol* **14**: 103–119.
- Lin B, Hyacinthe C, Bonneville S, Braster M, Van Cappellen P, Røling WFM. (2007). Phylogenetic and physiological diversity of dissimilatory ferric iron reducers in sediments of the polluted Scheldt estuary, Northwest Europe. *Environ Microbiol* **9**: 1956–1968.
- Loeb R, Lamers LP, Roelofs JG. (2008). Prediction of phosphorus mobilisation in inundated floodplain soils. *Environ Pollut* **156**: 325–331.
- Ludwig W, Strunk O, Westram R, Richter L, Meier H, Yadhukumar A *et al*. (2004). ARB: a software environment for sequence data. *Nucleic Acids Res* **32**: 1363–1371.
- Lutters-Czekalla S. (1990). Lithoautotrophic growth of the iron bacterium *Gallionella ferruginea* with thiosulfate or sulfide as energy source. *Arch Microbiol* **154**: 417–421.
- McLean RJC, Fortin D, Brown DA. (1996). Microbial metal-binding mechanisms and their relation to nuclear waste disposal. *Can J Microbiol* **42**: 392–400.
- Muyzer G, Dewaal EC, Uitterlinden AG. (1993). Profiling of complex microbial-populations by Denaturing Gradient Gel-Electrophoresis analysis of polymerase chain reaction-amplified genes-coding for 16S ribosomal-RNA. *Appl Environ Microbiol* **59**: 695–700.
- Neubauer SC, Emerson D, Megonigal JP. (2002). Life at the energetic edge: kinetics of circumneutral iron oxidation by lithotrophic iron-oxidizing bacteria isolated from the wetland-plant rhizosphere. *Appl Environ Microbiol* **68**: 3988–3995.
- Neubauer SC, Toledo-Duran GE, Emerson D, Megonigal JP. (2007). Returning to their roots: iron-oxidizing bacteria enhance short-term plaque formation in the wetland-plant rhizosphere. *Geomicrobiol J* **24**: 65–73.
- Parro V, Moreno-Paz M, Gonzalez-Toril E. (2007). Analysis of environmental transcriptomes by DNA microarrays. *Environ Microbiol* **9**: 453–464.
- Pruesse E, Quast C, Knittel K, Fuchs BM, Ludwig WG, Peplies J *et al*. (2007). SILVA: a comprehensive online resource for quality checked and aligned ribosomal RNA sequence data compatible with ARB. *Nucleic Acids Res* **35**: 7188–7196.
- Roden EE. (2004). Analysis of long-term bacterial vs chemical Fe(III) oxide reduction kinetics. *Geochimica Et Cosmochimica Acta* **68**: 3205–3216.
- Roden EE, Sobolev D, Glazer B, Luther GW. (2004). Potential for microscale bacterial Fe redox cycling at the aerobic-anaerobic interface. *Geomicrobiol J* **21**: 379–391.
- Sobolev D, Roden EE. (2002). Evidence for rapid microscale bacterial redox cycling of iron in circumneutral environments. *Antonie Van Leeuwenhoek International Journal of General and Molecular Microbiology* **81**: 587–597.
- Sobolev D, Roden EE. (2004). Characterization of a neutrophilic, chemolithoautotrophic Fe(II)-oxidizing beta-proteobacterium from freshwater wetland sediments. *Geomicrobiol J* **21**: 1–10.
- Søgaard EG, Aruna R, Abraham-Peskir J, Koch CB. (2001). Conditions for biological precipitation of iron by *Gallionella ferruginea* in a slightly polluted ground water. *Appl Geochem* **16**: 1129–1137.
- Straub KL, Benz M, Schink B. (2001). Iron metabolism in anoxic environments at near neutral pH. *FEMS Microbiol Ecol* **34**: 181–186.
- Weiss JV, Rentz JA, Plaia T, Neubauer SC, Merrill-Floyd M, Lilburn T *et al*. (2007). Characterization of

- neutrophilic Fe(II)-oxidizing bacteria isolated from the rhizosphere of wetland plants and description of *Ferritrophicum radicolica* gen. nov sp nov., and *Sideroxydans paludicola* sp nov. *Geomicrobiol J* **24**: 559–570.
- Widdel F, Schnell S, Heising S, Ehrenreich A, Assmus B, Schink B. (1993). Ferrous iron oxidation by anoxygenic phototrophic bacteria. *Nature* **362**: 834–836.
- Winderl C, Anneser B, Griebler C, Meckenstock RU, Lueders T. (2008). Depth-resolved quantification of anaerobic toluene degraders and aquifer microbial community patterns in distinct redox zones of a tar oil contaminant plume. *Appl Environ Microbiol* **74**: 792–801.
- Zhou JZ, Bruns MA, Tiedje JM. (1996). DNA recovery from soils of diverse composition. *Appl Environ Microbiol* **62**: 316–322.