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OPEN Changes in litter quality induced by nutrient addition alter litter decomposition in an alpine meadow on the Qinghai-Tibet Plateau

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The effects of nitrogen (N) and phosphorus (P) addition on litter decomposition are poorly understood in Tibetan alpine meadows. Leaf litter was collected from plots within a factorial N \times P addition experiment and allowed to decompose over 708 days in an unfertilized plot to determine the effects of N and/or P addition on litter decomposition. Results showed that nutrient addition significantly affected initial P and P-related biochemical properties of litter from all four species. However, the responses of litter N and N-related biochemical properties to nutrient addition were quite speciesspecific. Litter C decomposition and N release were species-specific. However, N and P addition significantly affected litter P release. Ratios of Hemicellulose + Cellulose to N and P were significantly related to litter C decomposition; C:N ratio was a determinant of litter N release; and C:P and (Hemicellulose + Cellulose): P controlled litter P release. Overall, litter C decomposition was controlled by litter quality of different plant species, and strongly affected by P addition. Increasing N availability is likely to affect litter C decomposition more indirectly by shifting plant species composition than directly by improving litter quality, and may accelerate N and P cycles, but shift the ecosystem to P limitation.

Plant litter decomposition is a key process regulating carbon (C) storage and nutrient (e.g., N, nitrogen, and P, phosphorus) cycles in most terrestrial ecosystems^{1,2}. Many studies have shown that litter decomposition usually depends on the environmental conditions, litter quality (i.e., its chemical composition) and decomposer activity²⁻⁶. Under specific climatic conditions, litter quality, e.g., N and P concentrations, and ratios of C:N, C:P and lignin:N are generally thought to control litter C decomposition and nutrient release^{5,7–9}.

However, litter chemical composition and thus litter decomposition can be strongly affected by changes in nutrient availability induced by human activity and climate change. For instance, increasing atmospheric N deposition has been widely observed on regional to global scales due to increased fossil fuel consumption and fertilizer application 10-12. In addition, soil organic matter (including organic N and P) mineralization is expected to increase due to global warming^{13,14}. Such increased nutrient availability could directly affect litter decomposition by improving litter quality (e.g., a higher N and P content, and lower C:N and C:P ratios)^{15,16}, or indirectly by changing plant species composition^{17–19} and thus the litter's chemical composition.

Whilst it has been suggested that plant species exert major control on litter decomposition at the global scale²⁰, the magnitude of direct or indirect effects of nutrient enrichment may be quite different at the site or regional scale. For instance, in an annual herb-based microcosm ecosystem (with eight species), Manning, et al.²¹ found that the direct effects of N addition significantly affected litter decomposition, but the indirect effects did not. However, in two grasslands rich in species (19-22 species m⁻²) Aerts, et al.¹⁷ showed that bulk litter decomposition was not increased by long-term (12-year) nutrient additions, but determined by initial plant species

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composition and thus by litter quality. These studies suggest that original plant species composition significantly affects the response of litter decomposition to nutrient addition, and further studies at the species level are still needed.

The alpine meadows of the Qinghai-Tibet plateau, an ecologically fragile area, are deficient in available N and P²². While litter decomposition could be an important source of nutrients for plant growth, the low temperature caused by the high altitude of this region could be a major limitation for litter decomposition. However, during recent decades, the alpine meadows have experienced increasing N deposition^{12,23} and a warming climate^{24,25}, which could significantly increase nutrient availability and thus affect litter decomposition and nutrient cycles. However, data on the extent to which nutrient enrichment affects litter decomposition at species level are still lacking. Therefore, our objectives were to investigate how nutrient (N and/or P) addition affects quality, C decomposition and nutrient release in aboveground plant litter produced by the dominant species in the alpine meadows of the Qinghai-Tibet plateau.

Results

Initial quality of leaf litter. The addition of N and/or P had different effects on initial chemical content and the nutrient ratios of leaves, and the response to nutrient addition was specific to each species. We used generalised linear models (GLMs) to analyse the effect of N addition, P addition, and their interaction (Tables 1 and 2). N addition and P addition and the interaction between the two significantly affected C content in *Kobresia humilis*, *Tibetia himalaica* and *Gentiana straminea* (P < 0.05) but did not affect C in *Stipa aliena* (P > 0.05). Across all four species, N- or P addition affected total C in leaf litter. Total N content in *K. humilis*, *S. aliena* and *G. straminea* was increased (P < 0.05) by nutrient addition but this was not the case in *T. himalaica* (P > 0.05) (Tables 1 and 2), and N concentration was not affected by nutrient addition in any of the species (P > 0.05).

Generally, the initial litter P content in the four species was significantly (P < 0.05) higher after P and NP addition treatments than either after CK or N addition (Table 2). The GLM analysis showed that N- or P addition and N, P interaction significantly (P < 0.05) increased the initial P content in leaf litter of the four species (Table 1). The initial P content of K. humilis, T. himalaica and G. straminea after the NP treatment was significantly lower than that after the P addition treatment, whereas the P content of S. aliena after the NP treatment was significantly higher than that after P addition alone (Table 2). Although the lignin, cellulose, and hemicellulose content was occasionally significantly (P < 0.05) altered in T. himalaica and G. straminea, their content in K. humilis and S. aliena was not greatly affected by nutrient addition (Table 1). In the four species, nutrient addition considerably increased the initial P content and decreased lignin content in leaf litter (Table 1).

We used the ratios to give information about relative chemical composition. C:N ratios and (Hemicellulose + cellulose):N ratios were not changed by nutrient addition in any of the species (P > 0.05), and lignin:N was not affected in *T. himalaica*. Otherwise, nutrient addition affected ratios in all four species: significantly influencing the C:P ratios, N:P ratios, lignin:P ratios and (Hemicellulose + cellulose):P ratios (P < 0.05).

Leaf litter chemical composition and the ratios differed more between species than between nutrient addition treatments (Table 2). The maximum mean total C was in *T. himalaica*, and the variation was relatively small between the four species. Mean total N and P content were higher in *T. himalaica* than in the other three non-legume species. Lignin in *T. himalaica* and *G. straminea* were higher but hemicellulose and cellulose were lower than in *K. humilis* and *S. aliena*.

Litter Closs. In all four species, the percentage of litter remained stable during the cold season (October to April) and decreased rapidly in the warm season (May to September) (Fig. 1). Overall, the decomposition of litter C was significantly affected by P addition in all the plant species except T. himalaica, but this was not the case for N addition. For example, compared with the CK treatment, P addition accelerated the loss of litter C in K. humilis and S. aliena, but decreased it in G. straminea. The interaction of N and P addition only showed significant effect in S. aliena (Fig. 1). At the end of the field incubation period, the percentage of C remaining was significantly different between the different plant species, but not between the nutrient addition treatments (Table 3). For instance, the average percentage of C remaining was ranked: K. humilis (41.6%) = S. aliena (39.5%) > T. himalaica (35.2%) = G. straminea (32.7%). The decomposition kinetics estimated using the two-pool exponent model showed that the fraction of initial total C was 64-78% in the active C pool with a decomposition rate constant of 0.63–1.80 yr⁻¹ and was 22–36% in the stable C pool with no decomposition during the experimental period (Fig. S1, Table 3). Statistics showed that the fraction of initial total C in the active/stable C pool, and the decomposition rate constant for active C were significantly affected by plant species, but not by nutrient addition. The average fraction in the active C pool was greater in K. humilis (76.9%) and S. aliena (73.0%) than in G. straminea (68.4%) and T. himalaica (66.5%). However, the decomposition rate constant for the active C pool was lower in K. humilis (0.66 yr⁻¹) and S. aliena (0.87 yr⁻¹) than in G. straminea (1.51yr⁻¹) and T. himalaica (1.58 yr⁻¹) (Table 3). The estimated fraction of the stable C pool in the initial total C was much lower than the observed percentage of C remaining at the end of the experiment in K. humilis and S. aliena (23.1-27.0% vs. 39.5-41.6%), but similar to the observed value in T. himalaica and G. straminea (32.7–35.2% vs. 31.6–33.5%), suggesting that material from the former two plant species will continue to decompose quickly, while the later two had reached a stable level and will decompose more slowly subsequently.

Litter N and P release. The pattern of N gain or loss in leaf litter during the experimental period was significantly affected by plant species and nutrient addition (Fig. 2a–d and Fig. S2). The percentage of N remaining was significantly affected by N addition and P addition in K. humilis, S. aliena and G. straminea, but not in T. himalaica, and by the interaction of N and P addition in all the plant species except S. aliena. In detail, the average percentage of N remaining at the end of the experiment was similar to the value of the stable level estimated by two-pool models, and the species could be ranked: K. humilis (92.2%) > S. aliena (62.0%) = G. straminea

		Total C	Total N	Phosphorus	Lignin	Cellulose	Hemicellulose
All species	s			_			
N		4.79*	0.87 ^{ns}	7.26**	2.19 ^{ns}	0.05 ^{ns}	0.09 ^{ns}
P		6.92*	0.00 ^{ns}	168.89***	4.94 [*]	0.01 ^{ns}	0.10 ^{ns}
$N \times P$		0.09 ^{ns}	0.41 ^{ns}	4.08*	0.02 ^{ns}	0.12 ^{ns}	0.06 ^{ns}
K. humilis					Į.	I.	
N		0.04 ^{ns}	17.22***	20.22***	0.94 ^{ns}	2.04 ^{ns}	3.78 ^{ns}
P		1.53 ^{ns}	9.01**	238.38***	1.17 ^{ns}	3.55 ^{ns}	1.85 ^{ns}
N×P		27.32***	48.54***	16.00***	2.17 ^{ns}	0.82 ^{ns}	2.08 ^{ns}
S. aliena			I.	L	I.	Į.	
N		0.21 ^{ns}	64.05***	16.76***	0.23 ^{ns}	0.001 ^{ns}	0.81 ^{ns}
P		0.02 ^{ns}	7.77 [*]	580.1***	0.01 ^{ns}	0.19 ^{ns}	0.74 ^{ns}
N×P		1.37 ^{ns}	7.12*	11.58**	0.33 ^{ns}	1.27 ^{ns}	0.09 ^{ns}
T. himalai	са		I.	L	I.	I.	
N		11.54**	0.08 ^{ns}	17.34***	0.00 ^{ns}	0.009 ^{ns}	0.72 ^{ns}
P		16.90***	2.37 ^{ns}	85.96***	2.18 ^{ns}	7.11*	1.11 ^{ns}
$N \times P$		25.03***	1.45 ^{ns}	5.67*	6.95*	2.47 ^{ns}	6.30*
G. stramin	iea			l.			
N		7.00*	1.14 ^{ns}	7.48*	13.90**	1.54 ^{ns}	0.96 ^{ns}
P		5.18*	10.29***	123.75***	20.50***	0.57 ^{ns}	4.94*
$N \times P$		0.53 ^{ns}	0.85 ^{ns}	6.42*	1.84 ^{ns}	3.99 ^{ns}	0.02 ^{ns}
	C: N	C:P	N:P	Lignin:N	Lignin:P	(H+C):N	(H+C):P
All species	s						
N	2.57 ^{ns}	0.903 ^{ns}	19.750***	4.386*	0.057 ^{ns}	1.362ns	0.152 ^{ns}
P	0.01 ^{ns}	183.129***	230.160***	1.201 ^{ns}	121.731***	0.052 ^{ns}	49.274***
$N \times P$	1.91 ^{ns}	0.260 ^{ns}	3.607 ^{ns}	2.430 ^{ns}	1.233 ^{ns}	0.910 ^{ns}	0.013 ^{ns}
K. humilis							
N	9.54**	3.023 ^{ns}	12.283**	7.966*	0.402 ^{ns}	7.424*	3.560 ^{ns}
P	2.37 ^{ns}	62.347***	36.673***	0.002 ^{ns}	46.155***	0.989 ^{ns}	63.446***
$N \times P$	45.20***	2.825 ^{ns}	1.957 ^{ns}	26.582***	4.283ns	15.653***	1.891 ^{ns}
S. aliena							
N	80.67***	0.030 ^{ns}	5.865*	26.670***	0.019 ^{ns}	59.760***	0.054 ^{ns}
P	6.03*	49.071***	37.005***	2.786 ^{ns}	35.454***	3.007 ^{ns}	49.550***
$N \times P$	3.87 ^{ns}	0.440 ^{ns}	0.706 ^{ns}	0.689 ^{ns}	0.145 ^{ns}	2.902 ^{ns}	0.592ns
T. himalai	са						
N	1.07 ^{ns}	11.093**	11.360**	0.053 ^{ns}	10.217**	0.116 ^{ns}	10.699**
P	6.26*	55.868***	38.603***	3.396 ^{ns}	41.808***	6.472*	48.703***
$N \times P$	0.11 ^{ns}	0.103 ^{ns}	0.115 ^{ns}	2.649 ^{ns}	0.327 ^{ns}	1.336 ^{ns}	0.028 ^{ns}
G. stramin	iea		•		•	•	
N	2.09 ^{ns}	2.229ns	5.009*	15.776***	0.493 ^{ns}	0.101 ^{ns}	2.035 ^{ns}
Р	13.69**	65.948***	78.167***	33.101***	71.664***	4.618*	43.447***
- 1	10.00						

Table 1. Summary of GLMs (F-value) for the effects of nitrogen addition (N), phosphorus addition (P), and N and P interaction (N \times P) on the chemical properties of leaf litter. $^*P \le 0.05, ^{**}P \le 0.01, ^{***}P \le 0.001, \text{ ns: not significant, } (H + C)$: Hemicellulose + cellulose.

(56.6%) > T. himalaica (48.5%) (Table 3). In K. humilis, the percentage of N remaining slightly decreased in the first few months and increased thereafter to 98-101% at the end of the experiment under the CK, P and NP treatments, but decreased to a stable value of 69% in the N treatment (Table 3). In S. aliena, and G. straminea, N release exhibited the same pattern, i.e., a sharp decrease during the first few months, and stable thereafter. While net N release was also found in T. himalaica, it took more than a year to reach a stable level. Interestingly, the percentage of N remaining at the end of the experiment was significantly lower in the N treatment (50.5%) than in the CK, P and NP treatments (62.3-71.5%) for S. aliena, but greater for G. straminea (69.1% vs. 48.9-56.2%) and similar for T. himalaica (51.6% vs. 46.9-48.3%) (Table 3).

Generally, the percentage of P remaining was significantly affected by N and P addition in all the plant species (Fig. 2e-h and Fig. S3). One exception was that N addition did not have a significant effect on P release in K. humilis. However, the significant interactive effect of N and P addition on P release was only found in G. straminea, and not in the other three species. The percentage of P remaining at the end of the experiment was greater in the CK and N treatments (61–91%) than in the P and NP treatments (33–56%).

Species	Treatment	Total C (%)	Total N (%)	Phosphorus (‰)	Lignin (%)	Cellulose (%)	Hemicellulose (%)	
	CK	45.2 ± 0.41a	0.78 ± 0.09b	0.59 ± 0.08c	$6.00 \pm 1.09a$	29.0 ± 2.55b	34.6 ± 1.13a	
	N	44.0 ± 0.34b	1.20 ± 0.09a	0.56 ± 0.04c	5.21 ± 0.52a	30.6 ± 1.00ab	32.9 ± 1.15b	
K. humilis	P	43.6 ± 0.90b	$0.93 \pm 0.07b$	1.77 ± 0.25a	5.17 ± 0.41a	30.9 ± 0.95ab	33.2 ± 1.26ab	
	NP	44.9 ± 0.38a	0.82 ± 0.09 b	1.25 ± 0.07b	5.34 ± 0.70a	31.3 ± 0.76a	33.0 ± 0.74b	
	Mean	44.4 ± 0.85AB	0.93 ± 0.19C	1.05 ± 0.53B	5.4 ± 0.75B	30.4 ± 1.63B	33.4 ± 1.21A	
	CK	44.1 ± 0.35a	1.13 ± 0.07c	0.55 ± 0.07c	3.91 ± 0.51a	$33.0 \pm 0.77a$	34.6 ± 0.82a	
	N	44.3 ± 0.98a	1.67 ± 0.15a	$0.58 \pm 0.05c$	$4.15 \pm 0.50a$	32.6 ± 0.62a	34.4 ± 0.64a	
S. aliena	P	44.4 ± 0.76a	1.13 ± 0.09c	1.33 ± 0.09b	$4.06 \pm 0.53a$	$32.7 \pm 0.83a$	34.4 ± 0.77a	
	NP	43.9 ± 0.28a	1.39 ± 0.12b	1.61 ± 0.12a	$4.04 \pm 0.54a$	$33.1 \pm 0.84a$	34.0 ± 1.19a	
	Mean	44.2 ± 0.64B	1.33 ± 0.25B	1.02 ± 0.48B	4.0 ± 0.49C	32.8 ± 0.74A	34.4±0.85A	
	CK	44.9 ± 0.30a	2.74 ± 0.15a	1.25 ± 0.18c	6.01 ± 0.36 ab	20.1 ± 0.96ab	17.4 ± 1.20b	
	N	45.2 ± 0.41a	2.84 ± 0.18a	1.08 ± 0.12c	6.71 ± 0.51a	$21.1 \pm 1.12a$	18.2 ± 0.91ab	
T. himalaica	P	$45.1 \pm 0.50a$	2.93 ± 0.20a	2.40 ± 0.38a	6.32 ± 0.90ab	19.4 ± 2.00ab	19.0 ± 1.23a	
	NP	43.5 ± 0.46b	$2.87 \pm 0.08a$	1.76 ± 0.05b	5.61 ± 0.49b	18.3 ± 1.68b	17.5 ± 0.65b	
	Mean	44.7 ± 0.81A	2.85 ± 0.16A	1.62 ± 0.57A	6.2 ± 0.69A	19.7 ± 1.74C	18.0 ± 1.15B	
	CK	45.4 ± 0.69a	1.33 ± 0.09b	$0.58 \pm 0.03c$	8.32 ± 0.86a	12.5 ± 0.98b	14.5 ± 2.49a	
	N	44.9 ± 0.93a	1.33 ± 0.11b	0.56 ± 0.04c	6.31 ± 0.65b	14.5 ± 1.55a	13.7 ± 1.93ab	
G. straminea	P	45.0 ± 0.31a	1.45 ± 0.13ab	1.93 ± 0.40a	6.00 ± 1.11bc	13.3 ± 1.58ab	12.5 ± 1.33ab	
	NP	44.0 ± 0.47b	1.56 ± 0.15a	1.41 ± 0.17b	5.06 ± 0.85c	12.8 ± 1.37ab	11.4 ± 2.70b	
	Mean	44.8 ± 0.79A	1.42 ± 0.15B	1.12 ± 0.63B	6.4 ± 1.46A	13.3 ± 1.50D	13.0 ± 2.34C	
Species	Treatment	C: N Ratio	C:P	N:P	Lignin:N	Lignin:P	(H+C):N	(H+C):P
	CK	58.9 ± 6.17a	772.6 ± 88.61a	13.3 ± 2.46b	$7.8 \pm 1.25a$	$102.8 \pm 22.77a$	83.0 ± 11.91a	1090.1 ± 154.40a
	N	36.7 ± 2.40c	$785.8 \pm 60.13a$	21.5 ± 2.48a	4.4 ± 0.63c	92.7 ± 7.55a	53.0 ± 3.58c	1134.5 ± 67.48a
K. humilis	P	47.2 ± 3.02b	250.7 ± 38.00c	5.4 ± 1.07c	5.6 ± 0.37bc	29.9 ± 6.79b	69.5 ± 6.72b	368.1 ± 48.90c
	NP	55.4 ± 6.50a	359.2 ± 20.48b	6.5 ± 0.75c	6.6 ± 1.25ab	42.6 ± 5.97b	79.2 ± 9.45ab	513.5 ± 31.61b
	Mean	49.5 ± 9.91A	542.1 ± 252.11A	11.68 ± 6.82B	6.1 ± 1.56A	67.0 ± 34.16AB	71.2 ± 14.21A	776.6 ± 358.31A
	CK	39.0 ± 2.18a	811.0 ± 91.98a	20.9 ± 3.13b	$3.5 \pm 0.51a$	$72.0 \pm 13.28a$	59.8 ± 3.48a	$1242.3 \pm 132.3a$
	N	26.8 ± 2.25c	$772.8 \pm 54.21a$	28.9 ± 1.48a	2.5 ± 0.24b	72.1 ± 6.16a	40.5 ± 3.55c	1169.3 ± 101.19b
S. aliena	P	39.5 ± 2.71a	335.7 ± 25.10b	8.5 ± 0.96b	3.6 ± 0.32a	30.7 ± 4.65b	59.8 ± 4.65a	507.4 ± 31.41c
	NP	31.7 ± 2.80b	274.0 ± 21.28b	8.7 ± 0.83b	2.0 ± 0.31b	25.1 ± 2.92b	48.5 ± 5.36b	418.5 ± 37.42c
	Mean	34.2 ± 5.92B	548.4 ± 256.40A	16.8 ± 9.03A	3.11 ± 0.56C	50.0 ± 23.84BC	52.1 ± 9.29B	834.4±391.53A
	CK	16.4 ± 0.78a	364.8 ± 51.48b	22.2 ± 3.16b	$2.2\pm0.17ab$	49.1 ± 9.79b	$13.7 \pm 0.44a$	304.5 ± 42.07b
	N	15.9 ± 0.92ab	424.6 ± 45.44a	26.6 ± 2.57a	2.4 ± 0.09a	62.8 ± 5.66a	13.8 ± 0.70a	369.5 ± 49.42a
T. himalaica	P	$15.4 \pm 1.00 ab$	191.1 ± 26.29d	12.5 ± 2.33d	$2.2 \pm 0.43 ab$	26.4 ± 1.87c	$13.2 \pm 1.34 ab$	162.3 ± 18.21c
	NP	15.2 ± 0.36b	247.9 ± 8.25c	16.3 ± 0.84c	2.0 ± 0.22b	31.9 ± 2.67c	12.5 ± 0.85b	204.0 ± 14.60c
	Mean	15.7 ± 0.89C	$307.1 \pm 100.63B$	19.42 ± 5.97A	2.2 ± 0.28D	42.5 ± 15.69C	13.3 ± 0.99D	260.1 ± 89.53B
	CK	34.4 ± 2.26a	784.7 ± 34.85a	$22.9 \pm 1.72a$	$6.3 \pm 0.76a$	$144.1 \pm 20.08a$	20.6 ± 3.82a	$468.3 \pm 75.72a$
	N	33.8 ± 2.49a	803.7 ± 43.32a	23.8 ± 1.04a	4.7 ± 0.42 b	112.7 ± 7.63b	21.2 ± 2.69a	506.4 ± 74.11a
G. straminea	P	31.2 ± 2.58ab	241.1 ± 46.39b	7.8 ± 1.80c	$4.2 \pm 1.00 bc$	31.4 ± 4.08c	18.0 ± 3.27ab	137.0 ± 25.37b
	NP	28.5 ± 2.92b	315.7 ± 32.68b	11.3 ± 2.10b	3.2 ± 0.45c	36.6 ± 9.14c	15.5 ± 1.93b	175.0 ± 40.31b
	Mean	32.0 ± 3.38B	536.3 ± 268.58A	16.44 ± 7.38A	4.6 ± 1.31B	81.2 ± 50.95A	18.8 ± 3.61C	321.7 ± 179.20B

Table 2. Mean biochemical composition of leaf litter after different nutrient addition in the four species. Different lowercase letters in a column of the same plant species indicate significant difference at $P \le 0.05$ in treatment; different capital letters in the same column indicate significant difference in species. CK: Control, N: nitrogen addition, P: phosphorus addition, NP: combined N and P addition.

Relationships between biochemical composition and litter decomposition. Biochemical components were significantly related to the decomposition of leaf C, and nutrient release (Tables 4 and 5). Linear regression showed that the C:N ratio, lignin content, cellulose + hemicellulose, and ratios of cellulose + hemicellulose to N and P were significantly related to the fractions of active and stable C pools in the initial total C pool and the decomposition rate constant of the active C pool (Table 4). The ratios of total C, lignin, and cellulose + hemicellulose to N were significantly related to the percentage of N in the stable level. Similarly, those organic components to P ratios were significantly related to the percentage of P in the stable level (Table 4). Nevertheless, multi-step regression analysis showed that the decomposition rate constant of the active C pool was mainly controlled by the ratios of cellulose + hemicellulose to nutrient (i.e., N and P) (Table 5). The fraction in the stable pool was controlled by (Hemicellulose + Cellulose):N ratio with respect to the C remaining, by C:N ratio with respect to the N remaining, and by C:P and (Hemicellulose + Cellulose):P ratios with respect to the P remaining (Table 5).

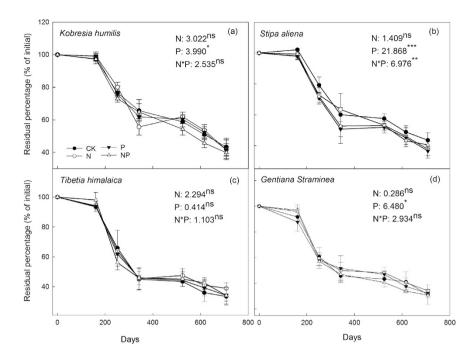


Figure 1. Percentage of C remaining in leaf litter under different nutrient addition treatments in four species: (a) Kobresia humilis, (b) Stipa aliena, (c) Tibetia himalaica, and (d) Gentiana straminea. CK: control, N: nitrogen addition, P: phosphorus addition, NP: combined N and P addition. A repeated measures ANOVA was used to analysis the effect of N and P addition, and their interaction on litter C decomposition.

		Observ	ed after 7	08 days	Estimated by a two-pool model			
Plant species	Treatment	f(C)	f(N)	f(P)	K_C	C_S	N _S	P_S
	CK	43.3a	100.6a	90.3a	0.64	23.9	105.3	90.4
K. humilis	N	41.6a	68.9b	90.8a	0.63	22.6	68.1	96.3
	P	41.6a	98.1a	41.7b	0.65	23.7	95.4	41.1
	NP	39.8a	101.2a	49.4b	0.70	22.3	100.0	47.0
	Mean	41.6 A	92.2 A	68.1 A	0.66 A	23.1 A	92.2 C	68.7 A
	CK	43.5a	63.6a	85.8a	0.68	24.4	65.7	86.2
	N	37.6a	50.5b	86.8a	0.78	23.8	49.7	84.3
S. aliena	P	36.5a	71.5a	47.8b	1.00	28.8	72.1	48.6
	NP	40.4a	62.3a	48.6b	1.02	31.1	57.0	46.8
	Mean	39.5A	62.0 B	67.2 A	0.87 A	27.0 AB	61.1 AB	66.5 A
	CK	33.5a	47.3a	61.1ab	1.48	30.3	48.3	51.9
	N	39.0a	51.6a	68.6a	1.57	35.9	51.3	60.7
T. himalaica	P	34.3a	46.9a	42.2b	1.47	31.9	37.7	35.9
	NP	34.1a	48.3a	56.3ab	1.80	35.8	50.9	60.1
	Mean	35.2 B	48.5 C	57.0 A	1.58 B	33.5 B	47.1 A	52.1 A
	CK	33.9a	56.2b	75.7a	1.62	33.6	67.3	73.7
	N	34.2a	69.1a	81.2a	1.43	33.0	67.8	75.3
G. straminea	P	32.1a	52.3b	33.2b	1.50	31.8	63.9	31.1
	NP	30.8a	48.9b	34.8b	1.48	28.0	56.9	39.4
	Mean	32.7 C	56.6 B	56.2 A	1.51 B	31.6 B	64.0 B	54.9 A

Table 3. Percentage of initial C, N and P remaining in leaf litter after 708 days, and values at stable level estimated by a two-pool model. f(C), f(N) and f(P) were the percentage of initial litter C, N and P remaining at the end of the experiment (%). K_C was the decomposition rate constant of the active C pool (yr⁻¹). C_S , N_S and P_S were the fraction of litter C, N and P in the stable pools, respectively (%).

Discussion

Litter quality and litter C decomposition. In line with previous findings²⁰, plant species significantly affected the decomposition of litter C (Fig. 1 and Table 3); this could be attributed to differences in litter quality. In general, initial lignin content and lignin:N ratio are considered to be the major factors controlling the

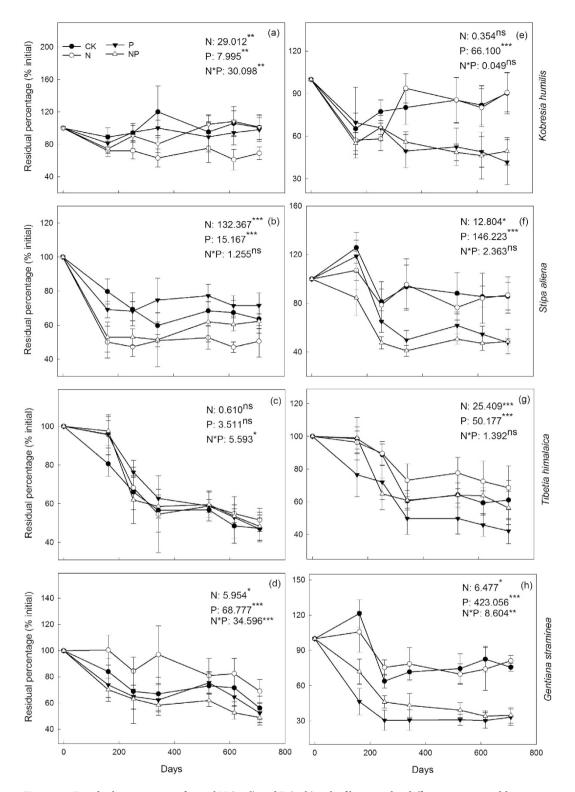


Figure 2. Residual percentages of initial N (\mathbf{a} - \mathbf{d}) and P (\mathbf{e} - \mathbf{h}) in leaf litter under different nutrient addition treatments in the four plant species. A repeated measures ANOVA was used to analysis the effect of N and P addition, and their interaction on residual nutrient. CK: control, N: nitrogen addition, P: phosphorus addition, NP: combined N and P addition.

decomposability of litter, especially in woody plant species 7,20,26 . In a path analysis, Zhang, *et al.* 3 found that the C:N ratio and total nutrient content (the sum of individual nutrient concentrations) of the litter were the two factors that had the greatest effect on litter mass loss across a wide range of plant species and tissues. However, in the alpine meadow studied here, we found that (Hemicellulose + Cellulose):N ratio was the mostly direct factor on the participation of litter C components and the decomposition rate constant of active C (Table 5).

Biochemical composition	K _C	Cs	Ns	Ps
C:N ratio	0.57***	0.50**	0.90***	0.06
C:P ratio	0.12	0.10	0.04	0.86***
Lignin	0.34*	0.27*	0.00	0.00
Lignin:N ratio	0.21	0.20	0.82***	0.07
Lignin:P ratio	0.00	0.00	0.04	0.59***
Hemicellulose + Cellulose	0.80***	0.52**	0.19	0.11
(Hemicellulose+Cellulose):N	0.85***	0.65***	0.68***	0.08
(Hemicellulose + Cellulose):P	0.55***	0.44**	0.09	0.76***

Table 4. Correlation coefficients (\mathbb{R}^2) between biochemical composition and C decomposition, residual N and P at stable level across the four species. K_C was the decomposition constant of active litter C pools; C_S , N_S and P_S were the fraction of litter C, N and P in the stable pool.

Model	F	P	Adjusted R ²	df
$K_{C} = -0.012 \times (\text{Hemicellulose} + \text{Cellulose}) : \text{N} - 0.34 \times (\text{Hemicellulose} + \text{Cellulose}) : \text{P} \times 10^{-3} + 1.82 \times 10^{-3} + 1.8$	63.88	0.000	0.89	15
$Cs = -0.15 \times (Hemicellulose + Cellulose):N + 34.6$	26.07	0.000	0.63	15
$Ns = 1.36 \times C:N + 21.3$	130.70	0.000	0.90	15
$Ps = 0.054 \times C:P + 0.019 \times (Hemicellulose + Cellulose):P + 24.1$	57.22	0.000	0.88	15

Table 5. Multi-step regression analysis of C decomposition, residual N and P at stable level with biochemical compositions across the four species. K_C is the decomposition rate constant of the active C pool. Cs, Ns, Ps are the percentage of initial C, N and P in the stable pool, respectively.

Despite its strong influence on initial N content and the ratios of C, lignin, and hemicellulose/cellulose to N (especially in *K. humilis* and *S. aliena*), N addition did not significantly affect the decomposition of litter C associated with any of the four plant species (Table 1 and Fig. 1). This result suggests a neutral direct effect of increasing atmospheric N deposition on litter C decomposition in this region. However, in a meta-analysis of the effects of N addition on litter decomposition, Knorr, *et al.*²⁷ found that decomposition was inhibited by N addition for low-quality (lignin content > 20%) litters, but accelerated for high-quality (lignin content < 10%) litters. Nevertheless, we did not find a significantly positive effect of N addition on litter C decomposition, even though all four plant species in our study (4–8% lignin content) produced high-quality litters (Table 2). On the one hand, the increased N supply could increase the proportion of litter humification rather than decomposition, e.g., by suppressing the activity of lignolytic enzymes^{28,29}, or increasing the microbial C use efficiency^{30,31}. On the other hand, the quality of leaf litter in our study was within the ideal (non-limiting) levels (i.e., lignin content <10% and lignin:N < 10) for decomposition as summarised by Prescott⁴, suggesting that other nutrient content (e.g., P), or climatic conditions (e.g., temperature) could be major limiting factors for litter decomposition. For example, the mean annual temperature at this site ($-1.6\,^{\circ}$ C) was less than the threshold value ($10\,^{\circ}$ C), and could inhibit the overall decay process⁴.

Unlike N addition, P addition significantly accelerated the decomposition of non-legume litter (i.e., litter of K. humilis, S. aliena, G. straminea) but not the legume litter (i.e. of T. himalaica) (Fig. 1). This was probably due to the improving stoichiometry of C and nutrients. For example, the C:P ratios of non-legume litter decreased from 773-810 in the control and N addition treatments to 241-359 in the P and NP treatments (Table 2). A wealth of evidence shows that there is a strong stoichiometry of C:N:P in soil humus (e.g., 1000:83.3:20, Himes³²) and microbial biomass (e.g., 1000:116.7:16.7; Creamer et al., 2014), and the stoichiometry controls litter C decomposition⁹. The initial C:N:P ratio of the three non-legume litters was 1000:(17.0-29.1):(1.2-1.3) under the no-fertilizer control. If one third of the initial C was finally converted to stable humus-C, the initial nutrient content could only support a ratio of 1000:(50.9-87.2):(3.7-3.9) in humus when N supply was very close to the demand of humus formation; however, the P supply was only able to fulfil a quarter of the demand. These results suggest that P was more limited than N with respect to litter C decomposition for the three non-legume plant species. At the end of the experiment, the percentage of C remaining was not affected by P addition (Table 3), suggesting a counterbalance between the positive effect of P addition to litter C decomposition and humification. Nevertheless, for S. aliena (a dominant species accounting for 31.5% of the community aboveground biomass), P addition tended to accelerate the decomposition of active C, but to increase the proportion of stable C (Table 3). In addition, the N:P ratios were generally greater in leaf litter under the no P addition treatment (20.5-28.8) than the critical value of 16 reported by Koerselman and Arthur³³, suggesting that P availability limited plant growth in this region. Therefore, long-term P addition could promote litter C sequestration in this region both by increasing plant biomass and thus C input, and by increasing litter C humification.

Nutrient release. Unlike the patterns of litter C loss, net N release and immobilization in leaf litter were strongly controlled by initial C:N ratio (Fig. 2a–d and Table 5). For instance, the N-fixing legume *T. himalaica*, with a low C:N ratio (15.2–16.4, Table 2), exhibited continuous N release during the experimental period, because the N requirements of microbial decomposers were being met. The non-legume plants *S. aliena* and *G. straminea*, with a median C:N ratio (26.8–39.5, Table 2), exhibited rapid N release during the first few months but no net N

Depth	pН	TC (g·kg ⁻¹)	TOC (g·kg ⁻¹)	TN (g·kg ⁻¹)	TP (g⋅kg ⁻¹)	AP (mg·kg ⁻¹)
0-5 cm	7.46 (0.04)	89.0 (1.33)	72.4 (0.16)	8.7 (0.12)	0.84 (0.006)	6.2 (0.009)
5-10 cm	7.61 (0.03)	66.9 (0.72)	62.4 (0.25)	6.9 (0.08)	0.69 (0.006)	5.2 (0.008)
10-20 cm	7.96 (0.02)	54.8 (0.46)	42.8 (0.16)	5.9 (0.07)	0.56 (0.018)	3.2 (0.008)
20-40 cm	8.36 (0.01)	40.4 (0.28)	33.0 (0.19)	3.7 (0.07)	0.44 (0.006)	2.2 (0.011)
40-60 cm	8.57 (0.02)	30.6 (0.29)	23.6 (0.11)	1.9 (0.06)	0.30 (0.011)	0.2 (0.006)

Table 6. Summary of soil characteristics at different depths at the study site. Means of pH, total carbon content (TC), total organic carbon (TOC), total nitrogen concentration (TN), total phosphorous content (TP), and available phosphorous (AP) are shown within parentheses.

immobilisation or release thereafter. However, the non-legume plant K. humilis with a high C:N ratio (47.2–58.9, Table 2) showed little or no net N immobilisation/release. Our results are in line with those of Parton, $et\ al.^8$ that net N release occurred when the average C:N ratio of the leaf litter was <40 (a range of 31 to 48). As the C:N ratios of major plant species were generally less than 40 in the alpine meadows of the Qinghai-Tibet plateau (Table 2, and supplemental Table 1 in Duan, $et\ al.^{34}$), the future increasing N deposition in this region may not change the pattern but may change the magnitude of initial N release.

The pattern of P release was strongly controlled by C:P and (Hemicellulose + Cellulose):P ratios (Table 5), and significantly regulated by P addition (Fig. 2). When the initial C:P ratios were within the critical values for P release (700–900) observed by Moore, *et al.*³⁵, e.g., litter of non-legume plant without P addition, we only found limited P immobilisation at the beginning of the experiment and little P loss thereafter. However, when the initial C:P ratio was lower than the critical values, e.g., litter of a legume plant with/without P addition and non-legume plant with P addition, P was immediately lost from the litter of all plant species and the magnitude of P loss was much greater.

Implications

Our study showed that the variation in litter C decomposition was much greater between plant species than between different nutrient additions (Table 3), suggesting that any management without a shift in plant species composition (and thus litter quality) is unlikely to affect the litter C decomposition significantly. Many studies have shown that increasing nutrient (especially N) addition could increase plant biomass, but decrease plant species diversity and cause shifts in plant community composition 15,16,36,37. In the same ecosystem near to our experimental site, Yang, et al. 38 found that N and P additions significantly increased the dominance of graminoids (including S. aliena), but decreased the dominance of forbs (including T. himalaica and G. straminea) in the community after 4 years of fertilizer application. Taken together, our results and those demonstrating plant species changes resulting from nutrient addition 38 suggested that the indirect effect of nutrient addition (i.e., changing the plant community composition) was more important than the direct effect (i.e., changing the litter quality) on litter C decomposition on the Qinghai-Tibet plateau. While future increasing N deposition in this region may increase the overall plant biomass and thus mean more highly N-enriched litter C input, the effect of P limitation on plant growth and microbial C use efficiency will restrict C sequestration in litter. Therefore, to maximise sequestering C and N in the soil-plant system of alpine meadows, a certain amount of P should be added.

Material and Methods

Experimental site. This study was conducted at the Haibei Alpine Meadow Ecosystem Research Station (latitude: 37° 37′ N, longitude: 101° 12′ E; elevation: 3200 m above sea level) in Qinghai Province; this area is characterised by a continental monsoon climate with severe, long winters and short, cool summers. The mean annual temperature and precipitation are $-1.6\,^{\circ}$ C and 500 mm, respectively; 80% of precipitation falls in the growing seasons from May to September. The soil is classified as Mat-GryicCambisol (according to Chinese Soil Taxonomy), with pH 7.46, total organic C 72.4 g kg $^{-1}$, total N 8.7 g kg $^{-1}$, total P 0.84 g kg $^{-1}$ and available P 6.2 mg kg $^{-1}$ in the 0–5 cm soil layer (Table 6). The plant species in the alpine meadows are divided into four functional groups: (I) sedges: *K. humilis, Carex scabrirostris*, etc; (II) graminoids: *S. aliena, Elymus nutans*; *Poa pratensis*, etc; (III) leguminous forbs: *T. himalaica*, *Medicago archiducis-nicolai*, etc; and (IV) non-leguminous forbs: *G. straminea*, *Potentilla nivea*, *Saussurea superba*, etc. The corresponding dominant species in each functional group were *K. humilis*, *S. aliena*, *T. himalaica* and *G. straminea*, accounting for 3.1%, 31.5%, 5.0% and 5.0% of the community aboveground biomass, respectively.

Fertilisation design. A $110 \,\mathrm{m} \times 75 \,\mathrm{m}$ area of flat land with relatively uniform vegetation was fenced in May 2009. A two-way factorial design with N and P addition was implemented involving four treatments: (1) CK, control; (2) N, nitrogen addition ($10 \,\mathrm{g} \,\mathrm{Nm}^{-2} \mathrm{yr}^{-1}$ as urea); (3) P, phosphorus addition ($10 \,\mathrm{g} \,\mathrm{Pm}^{-2} \mathrm{yr}^{-1}$ as triple superphosphate); and (4) NP, combined N and P addition ($10 \,\mathrm{g} \,\mathrm{Nm}^{-2} \mathrm{yr}^{-1} + 10 \,\mathrm{g} \,\mathrm{P} \,\mathrm{m}^{-2} \mathrm{yr}^{-1}$). All nutrients were applied twice per year (50% on June 22^{nd} and 50% on July 15^{th}) after sunset, when the soil moisture was high, in both 2009 and 2010. In total, 20 plots (measuring $6 \,\mathrm{m} \times 6 \,\mathrm{m}$ with a 2-m wide buffer zone), comprising 4 treatments with 5 replicates each, were used in a completely randomised design at the study site.

Litter-bag incubation experiment. In early October of 2010, freshly senesced leaf litters were sampled from the four dominant species (K. humilis, S. aliena, T. himalaica, and G. straminea) grown under each of the different treatments. The litters were air-dried until they reached a constant weight, then placed in polyethylene litter bags ($10 \text{ cm} \times 15 \text{ cm}$, with an upper layer of $1 \text{ mm} \times 1 \text{ mm}$ mesh and a lower layer of $0.5 \text{ mm} \times 0.5 \text{ mm}$ mesh).

Each litter bag was filled with 4g of air-dried litter, clipped to fragments measuring 5 cm in length. The litter bags were placed in a plot that had not received any fertilizer treatment, and the bags were staked on the soil among the senescent forbs for 708 days from 10 November 2010 to 20 October 2012. The litter bags were collected twice during the growing season (April 20 and July 20 in 2011 and 2012) and once outside the growing season (October 20 of 2011 and 2012) in each year. In total, 480 litter bags (4 treatments applied to 4 species across 5 replicates with 6 collection and measurement points) were prepared and deployed onto the treatment plots. Another 80 litter bags (4 treatments applied to 4 species across 5 replicates) were prepared using the same initial drying treatment and directly brought to the laboratory for the measurement of weight after oven drying and determination of initial chemistry.

Litter chemical analysis. The initial chemical composition of all plant litter was determined for subsamples. The retrieved litter samples were oven-dried at 65 °C for 48 h and weighed. Next, carbon (C) and N content of each leaf litter sample was assayed using an elemental analyser (2400 II CHNS/O Elemental Analyzer; Perkin-Elmer, Boston, MA, USA). Total P was determined using the molybdate colorimetric method after ascorbic acid reduction³⁹. Lignin, cellulose, and hemicellulose were determined using the sulphuric acid procedure⁴⁰. The C:N ratio and initial P, lignin, cellulose, and hemicellulose content were used to represent the leaf litter quality.

Data analysis. Litter decomposition rate varied across the seasons, e.g., faster in the warm season (April to September) and slower or even zero in the cool season (October to March). To reduce the bias in simulation of litter decomposition, we used a standard year which was modified on the basis of cumulative degree days (CDD). CDD was calculated by summing >0 °C degree days from the date litter was placed in the field to the date of sampling. Daily temperature <0 °C was set to 0. A standard year was defined to be 1723 cumulative degree days at the experimental site, which was the average value of a period lasting 30 years (1981–2010). Therefore, the seven sampling dates were on the standard year of 0, 0.05, 0.58, 1.06, 1.09, 1.64 and 2.14, respectively. The percentage of litter C, N and P remaining was estimated by fitting a two-pool model (Fig. S1–S3):

$$C_t = C_s + (100 - C_s)e^{-K_C t} (1)$$

$$N_t = N_s + N_a e^{-K_N t} (2)$$

$$P_t = P_s + P_a e^{-K_P t} (3)$$

where C_t , N_t and P_t are the percentage of litter C, N and P remaining at time t (standard years), respectively. C_s , N_s and P_s are the fractions in the stable pools, and $100-C_s$, N_a and P_a are the fractions in the active pools for litter C, N and P, respectively. K_C , K_N and K_P are the decomposition rate constants of litter C, N and P in the active pools.

GLMs with multi-comparisons were used to determine the effects of plant species and nutrient addition (N and P addition and their interactions) on all the considered initial litter quality, and percentage of litter C, N and P remaining at the end of the experiment. Repeated measures ANOVA was used to analyse the effect of N, P and N, P interaction on litter C decomposition and nutrient release; the analyses were performed with R 3.1.1. Linear regression was used to test and verify the correlation between the litter quality parameters (total C, lignin, cellulose and hemicellulose, and their ratios to N and P) and the litter C decomposition or nutrient release. Statistical analyses were performed using SPSS 16.0.

References

- 1. Aerts, R. & Chapin III, F. S. The mineral nutrition of wild plants revisited: A re-evaluation of processes and patterns. *Adv Ecol Res* **30**, 1–67 (1999)
- 2. Berg, B. & McClaugherty, C. Plant litter decomposition, humus formation, carbon sequestration (third edition) (Springer 2014).
- 3. Zhang, D., Hui, D., Luo, Y. & Zhou, G. Rates of litter decomposition in terrestrial ecosystems: global patterns and controlling factors. *Journal of Plant Ecology* 1, 85–93 (2008).
- 4. Prescott, C. E. Litter decomposition: what controls it and how can we alter it to sequester more carbon in forest soils? *Biogeochemistry* **101**, 133–149 (2010).
- 5. Aerts, R. Climate, Leaf Litter Chemistry and Leaf Litter Decomposition in Terrestrial Ecosystems: A Triangular Relationship. *Oikos* **79**, 439–449 (1997).
- Keeler, B. L., Hobbie, S. E. & Kellogg, L. E. Effects of Long-Term Nitrogen Addition on Microbial Enzyme Activity in Eight Forested and Grassland Sites: Implications for Litter and Soil Organic Matter Decomposition. *Ecosystems* 12, 1–15 (2009).
- Walela, C. et al. The initial lignin:nitrogen ratio of litter from above and below ground sources strongly and negatively influenced decay rates of slowly decomposing litter carbon pools. Soil Biol Biochem 77, 268–275 (2014).
- 8. Parton, W. et al. Global-scale similarities in nitrogen release patterns during long-term decomposition. Science 315, 361–364 (2007).
- 9. Manzoni, S., Trofymow, J. A., Jackson, R. B. & Porporato, A. Stoichiometric controls on carbon, nitrogen, and phosphorus dynamics in decomposing litter. *Ecological Monographs* 80, 89–106 (2010).
- Galloway, J. N. et al. Transformation of the nitrogen cycle: Recent trends, questions, and potential solutions. Science 320, 889–892 (2008).
- 11. Dentener, F. et al. Nitrogen and sulfur deposition on regional and global scales: A multimodel evaluation. Global Biogeochem Cy 20, n/a-n/a (2006).
- 12. Liu, X. *et al.* Enhanced nitrogen deposition over China. *Nature* **494**, 459–462 (2013).
- 13. Rustad, L. *et al.* A meta-analysis of the response of soil respiration, net nitrogen mineralization, and aboveground plant growth to experimental ecosystem warming. *Oecologia* **126**, 543–562 (2001).
- Bai, E. et al. A meta-analysis of experimental warming effects on terrestrial nitrogen pools and dynamics. New Phytol 199, 441–451 (2013).
- Li, Y., Niu, S. & Yu, G. Aggravated phosphorus limitation on biomass production under increasing nitrogen loading: a meta-analysis. Global Change Biology 22, 934–943 (2016).

- 16. Xia, J. & Wan, S. Global response patterns of terrestrial plant species to nitrogen addition. New Phytol 179, 428-439 (2008).
- 17. Aerts, R., De Caluwe, H. & Beltman, B. Plant community mediated vs. nutritinal controls in litter decomposition rates in grasslands. *Ecology* **84**, 3198–3208 (2003).
- 18. Suding, K. N. *et al.* Functional- and abundance-based mechanisms explain diversity loss due to N fertilization. *P Natl Acad Sci USA* **102**, 4387–4392 (2005).
- 19. DuprÈ, C. *et al.* Changes in species richness and composition in European acidic grasslands over the past 70 years: the contribution of cumulative atmospheric nitrogen deposition. *Global Change* Biol **16**, 344–357 (2010).
- 20. Cornwell, W. K. et al. Plant species traits are the predominant control on litter decomposition rates within biomes worldwide. *Ecology Letters* 11, 1065–1071 (2008).
- 21. Manning, P. et al. Direct and indirect effects of nitrogen deposition on litter decomposition. Soil Biol Biochem 40, 688-698 (2008).
- 22. Zhao, X.-Q. & Zhou, X.-m. Ecological basis of alpine meadow ecosystem management in Tibet: Haibei alpine meadow ecosystem research station. *Ambio* 28, 642–647 (1999).
- 23. Lü, C. & Tian, H. Spatial and temporal patterns of nitrogen deposition in China: Synthesis of observational data. *Journal of Geophysical Research: Atmospheres* 112, D22S05 (2007).
- Li, L., Yang, S., Wang, Z., Zhu, X. & Tang, H. Evidence of Warming and Wetting Climate over the Qinghai-Tibet Plateau. Arctic, Antarctic, and Alpine Research 42, 449–457 (2010).
- 25. Chen, B. Climatic warming in the Tibetan Plateau during recent decades. Int J Climatol 20, 1729-1742 (2000).
- 26. Taylor, B., Prescott, C., Parsons, W. & Parkinson, D. Substrate control of litter decomposition in four Rocky Mountain coniferous forests. *Canadian Journal of Botany* **69**, 2242–2250 (1991).
- 27. Knorr, M., Frey, S. D. & Curtis, P. S. Nitrogen additions and litter decomposition: a meta-analysis. Ecology 86, 3252-3257 (2005).
- 28. Frey, S. D., Knorr, M., Parrent, J. L. & Simpson, R. T. Chronic nitrogen enrichment affects the structure and function of the soil microbial community in temperate hardwood and pine forests. Forest Ecology and Management 196, 159–171 (2004).
- 29. Carreiro, M., Sinsabaugh, R., Repert, D. & Parkhurst, D. Microbial enzyme shifts explain litter decay responses to simulated nitrogen deposition. *Ecology* 81, 2359–2365 (2000).
- Ågren, G. I., Bosatta, E. & Magill, A. H. Combining theory and experiment to understand effects of inorganic nitrogen on litter decomposition. Oecologia 128, 94–98 (2001).
- 31. Manzoni, S., Jackson, R. B., Trofymow, J. A. & Porporato, A. The global stoichiometry of litter nitrogen mineralization. *Science* 321, 684–686 (2008).
- 32. Himes, F. Nitrogen, sulfur, and phosphorus and the sequestering of carbon. Vol. 315 (CRC Press, Boca Raton, FL, 1998).
- 33. Koerselman, W. & Arthur, F. M. M. The vegetation N:P ratio: a new tool to detect the nature of nutrient limitation. *Journal of Applied Ecology* 33, 1441–1450 (1996).
- 34. Duan, J. et al. Non-additive effect of species diversity and temperature sensitivity of mixed litter decomposition in the alpine meadow on Tibetan Plateau. Soil Biol Biochem 57, 841–847 (2013).
- 35. Moore, R. T., Trofymow, A. J., Prescott, E. C., Fyles, J. & Titus, D. B. Patterns of Carbon, Nitrogen and Phosphorus Dynamics in Decomposing Foliar Litter in Canadian Forests. *Ecosystems* 9, 46–62 (2006).
- 36. Magnani, F. et al. The human footprint in the carbon cycle of temperate and boreal forests. Nature 447, 849-851 (2007).
- 37. Fornara, D. A. & Tilman, D. Soil carbon sequestration in prairie grasslands increased by chronic nitrogen addition. *Ecology* **93**, 2030–2036 (2012).
- 38. Yang, X.-X., Ren, F., Zhou, H.-K. & He, J.-S. Responses of plant community biomass to nitrogen and phosphorus additions in an alpine meadow on the Qinghai-Xizang Plateau. *Chinese Journal of Plant Ecology* **38**, 159–166 (2014).
- 39. Murphy, J. & Riley, J. P. A modified sigle solution method for the determ- ination of phosphate in natural waters. *Anal Chim Acta* 27, 31–36 (1962).
- 40. Goering, H. K. & Van Soest, P. J. Forage fiber analyses (apparatus, reagents, prcedures, and some applications). US Department of Agriculture, Washington, DC (1970).

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Author Contributions

W.Z. designed and carried out the experiment, collected the data and drafted the manuscript. J.W. drew the figures and improved the language. Z.Z. performed the statistical analysis. F.R. helped to collect the samples. L.C. modified the language. J.-S.H. offered valuable opinions about the discussion.

Additional Information

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