

SCIENTIFIC REPORTS



OPEN

Peptide bonds affect the formation of haloacetamides, an emerging class of N-DBPs in drinking water: free amino acids versus oligopeptides

Received: 15 April 2015
Accepted: 26 August 2015
Published: 23 September 2015

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Haloacetamides (HAcAms), an emerging class of nitrogenous disinfection by-products (N-DBPs) of health concern, have been frequently identified in drinking waters. It has long been appreciated that free amino acids (AAs), accounting for a small fraction of the dissolved organic nitrogen (DON) pool, can form dichloroacetamide (DCAcAm) during chlorination. However, the information regarding the impacts of combined AAs, which contribute to the greatest identifiable DON portion in natural waters, is limited. In this study, we compared the formation of HAcAms from free AAs (tyrosine [Tyr] and alanine [Ala]) and combined AAs (Tyr-Ala, Ala-Tyr, Tyr-Tyr-Tyr, Ala-Ala-Ala), and found that HAcAm formation from the chlorination of AAs in combined forms (oligopeptides) significantly exhibited a different pattern with HAcAm formation from free AAs. Due to the presence of peptide bonds in tripeptides, Tyr-Tyr-Tyr and Ala-Ala-Ala produced trichloroacetamide (TCAcAm) in which free AAs was unable to form TCAcAm during chlorination. Moreover, peptide bond in tripeptides formed more tri-HAcAms than di-HAcAms in the presence of bromide. Therefore, the peptide bond may be an important indicator to predict the formation of specific N-DBPs in chlorination. The increased use of algal- and wastewater-impacted water as drinking water sources will increase health concerns over exposure to HAcAms in drinking water.

As a result of rapid population growth and rising water demand, drinking water source waters are facing threats of insufficiently treated wastewater effluents or algal blooms. These pollution sources are characterized by higher levels of dissolved organic nitrogen (DON) that can potentially react with certain disinfectants (e.g., chlorine) to form unwanted nitrogenous disinfection by-products (N-DBPs) in drinking water treatment plants (DWTPs)^{1–3}. Recently, interest in the formation of N-DBPs has increased because toxicological studies have demonstrated that N-DBPs are typically more genotoxic, cytotoxic, or carcinogenic than most carbonaceous disinfection by-products (C-DBPs) that have long been a major focus in previous studies^{1,4,5}. Haloacetamides (HAcAms), an emerging class of halogenated N-DBPs, are of particular concern because they were reported to be very cytotoxic and genotoxic in mammalian cell assays (for example, over 100 times more cytotoxic and 10 times more genotoxic than HAAs)⁶ and were frequently detected in drinking water^{2,7,8}.

The formation of N-DBPs from amino acids (AAs) upon chlorination is of interest, as AAs account for a significant fraction of DON in natural waters. In prior studies, free AAs were mostly selected as

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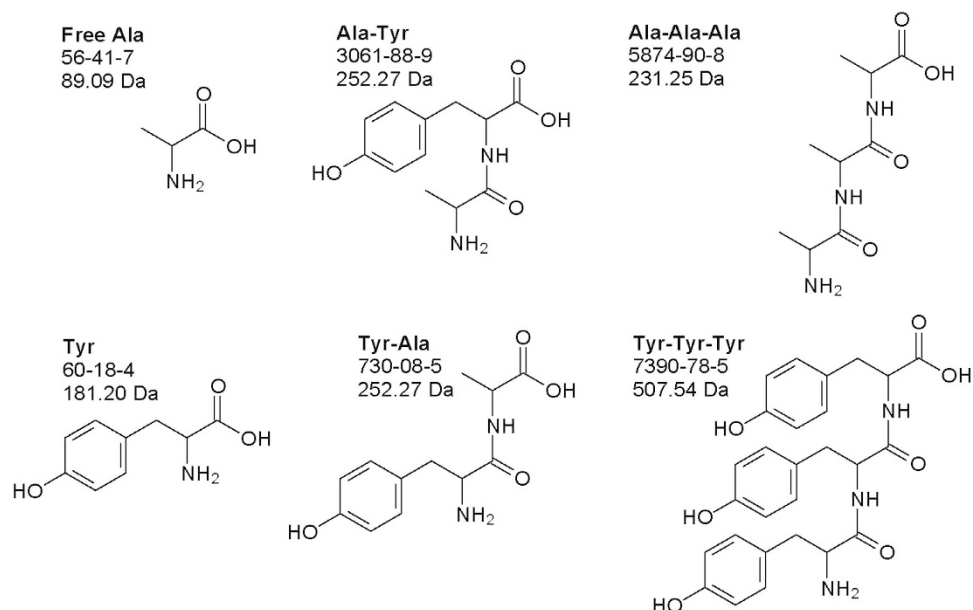


Figure 1. Chemical structures of selected free and combined AAs in the study.

model compounds to investigate the DBP formation mechanism^{1,9}. However, free AAs make up only an insignificant fraction (<6%) of the DON pool; in contrast, combined AAs contribute to the greatest identifiable portion, especially in algal- and wastewater-impacted water^{10–12}. Therefore, it is essential to examine the formation of N-DBPs from combined AAs. Combined amino acids (e.g., oligopeptides and proteins) are ubiquitous in surface waters and typically derive from viral lysis or autolysis of bacteria, microbial secretion of extracellular enzymes, atmospheric deposition, or anthropogenic inputs as pollutants^{11,13,14}.

It has been appreciated that part of the free AAs may serve as HAcAm precursors^{1,15,16}. For example, free tyrosine (Tyr) could react with chlorine to form dichloroacetamide (DCAcAm) and trichloroacetamide (TCAcAm)¹⁷. However, alanine (Ala) cannot form any HAcAm but might serve as a chloroform precursor^{15,18}. Unfortunately, it was still unclear whether the formation of HAcAms from the chlorination of oligopeptides and free AAs behaves significantly differently due to the presence of peptide bonds in the oligopeptides. The objective of this study was to compare the formation of HAcAms between the chlorination of free AAs and low-molecular mass combined AAs (oligopeptides), and thus evaluate the impacts of peptide bonds on HAcAm formation. Two free AAs, Tyr (HAcAm precursor) and Ala (Non-HAcAm precursor), and four oligopeptides, Tyr-Ala, Ala-Tyr, Tyr-Tyr-Tyr, and Ala-Ala-Ala (Fig. 1) were selected as precursor compounds in this study, because they share similar molecular structures except the presence or absence of peptide bonds on HAcAm formation.

Methods

Materials. Chloroacetamide (CAcAm) (98.5%), DCAcAm (98.5%), and TCAcAm (99%) standards were obtained from Alfa Aesar (Karlsruhe, Germany). Bromochloro- (BCAcAm), dibromo- (DBAcAm), bromodichloro- (BDCAcAm), dibromochloro- (DBCAcAm), and tribromoacetamide (TBAcAm) standards were all purchased from Orchid Cellmark (New Westminster, BC, Canada). Bromoacetamide (BACAm), two haloacetonitriles (HANs) (dichloroacetonitrile [DCAN] and trichloroacetonitrile [TCAN]), and the model compounds (Tyr [$\geq 99\%$], Ala [$\geq 99\%$], Tyr-Ala [$> 98\%$], Ala-Tyr [$> 98\%$], Ala-Ala [$> 98\%$], Tyr-Tyr-Tyr [$> 98\%$], and Ala-Ala-Ala [$> 98\%$]) were purchased from Sigma-Aldrich (Oakville, ON, Canada). A sodium hypochlorite solution (reagent grade [$> 98\%$], active chlorine $> 5\%$, Sinopharm Chemical Reagent Co., Ltd., China) was used to prepare free chlorine stock solutions. The ultrapure water was produced with a Millipore Milli-Q Gradient water purification system (Billerica, MA, USA). All the other chemical reagents were at least of analytical grade, and obtained from Sinopharm Chemical Reagent Co., Ltd (Shanghai, China) unless otherwise noted.

Experimental procedure. Chlorination tests were performed in 40-mL brown glass ampoule bottles at a controlled room temperature ($23.0 \pm 0.2^\circ\text{C}$) and under a headspace-free and light-free condition. In a typical run, an appropriate dose of chlorine was added to each model precursor solution (0.05 mM) to achieve the same molar ratio of chlorine (Cl_2) to model precursor nitrogen atom (Cl_2/N in model precursor = 20) at the beginning of the chlorination reaction. Solution pH was maintained in buffer solution (10 mM), which were prepared from phosphate and carbonate salts. If needed, NaOH and HCl were used to adjust the pH to a desirable level. To examine the speciation of HAcAms, an appropriate

dose of bromide (potassium bromide) was added to each model precursor solution (0.05 mM), to achieve the same molar ratio of bromide to model precursor nitrogen atom (bromide/N in model precursor = 2) at the beginning of the chlorination reaction. The Cl₂/N in model precursor ratio of 20 and bromide/N in model precursor ratio of 2 were selected in order to apply more realistic process conditions^{7,10–13,19,20}. To quench the chlorination reaction at designated times, the disinfectant residual was quenched with a stoichiometric amount of ascorbic acid. The quenched solution was analyzed as soon as possible after collection. Detailed information on the experimental procedure is available elsewhere¹⁵.

Analysis. In the analysis of 9 HAcAms, a simultaneous determination method for HAcAms, combining solid-phase extraction (SPE) enrichment, high-performance liquid chromatography (HPLC) separation, and triple quadrupole MS (tqMS) with atmospheric pressure chemical ionization (APCI), using selective reaction monitoring (SRM) in the positive mode, was developed.

The SPE performance of neutral (HLB), cation-exchanging (MCX, WCX), and anion-exchanging (MAX, WAX) OASIS polymers supplied by Waters (Milford, MA, USA) has been studied recently⁸. The neutral solutes HLB had the highest SPE performance (highest recoveries) for the nine HAcAms and was selected as the SPE sorbent for this method.

After SPE enrichment, an HPLC (e2695) from Waters (Milford, MA), employing a Hypersil GOLD C18 packed column (100 × 2.1 mm i.d., 5 μm) with a Hypersil GOLD precolumn (10 × 2.1 mm i.d., 5 μm) (Thermo Scientific; Waltham, MA) was used for separation. The 9 HAcAms were separated by LC in 9.0 min.

After the HPLC separation, a tqMS (TSQ Quantum Access MAX) from Thermo Scientific (Waltham, MA) was used to detect the 9 HAcAms by positive APCI combined with the SRM mode. The optimal operating parameters were as follows: discharge current at 4.0 μA, vaporizer temperature at 350 °C, sheath gas pressure at 40 psi, capillary temperature at 250 °C, and collision pressure at 1.5 m Torr. Transition ions, collision energy, and tube lens offset were optimized for individual analytes, as shown in Supplementary Information (SI) Table S1. The intraday and interday instrument precision were calculated by the relative standard deviations (RSDs) at three concentration levels (0.1, 1, 10 μg/L) for each HAcAm within the linear ranges. The intraday and interday RSDs (n = 5) for each HAcAm were generally lower than 10%. The details of the HAcAm and other N-DBP analyses are presented elsewhere²¹ and are summarized in the SI. The HAcAm yield was the molar ratio of the formed HAcAm to the initial concentration of selected free or combined AAs (equation 1). At the Cl₂/N in model precursor (molar ratio) of 20, the AAs could be consumed completely at a short period (<60 min)^{17,18,22}, thereby the initial AA molar concentration can be regarded as the consumed AA molar concentration.

$$\text{HAcAm yields} = \frac{\text{Formed HAcAm molar concentration}}{\text{Initial AA molar concentration}} \times 100\% \quad (1)$$

Results and Discussion

Impacts of peptide bonds in dipeptides on the HAcAm formation during chlorination. Figure 2 shows the time- and pH-dependent formation of DCACAm and TCACAm during chlorination of free AAs and combined AAs at a Cl₂/N in model precursor (molar ratio) of 20. As seen from Fig. 2A, the concentrations of DCACAm formed from free Tyr, Tyr+Ala, and two dipeptides (Tyr-Ala and Ala-Tyr) initially increased and then decreased with the contact time from 1 to 72 h, and peaked at 0.170%, 0.026%, 0.024%, and 0.005% at 24 h, respectively. The decrease of HAcAm yields after 24 h was probably because the residual chlorine accelerated the decomposition rate of HAcAms²². The mixed 'Tyr+Ala' formed less DCACAm than free Tyr, which implied the presence of Ala (Non-HAcAm precursor)¹⁸ in the water solution suppressed the formation of DCACAm from Tyr (Non-HAcAm precursor)¹⁷ upon chlorination, probably due to the difference of chlorine demand for Ala and Tyr.

In contrast, TCACAm was not detected during chlorination of Tyr, Ala, Tyr-Ala, or Ala-Tyr (Fig. 2B). This result is in agreement with the previous study which found that HAcAm precursors in natural waters more easily form DCACAm than TCACAm²³. Ala-Ala was similar with free Ala, which cannot form DCACAm and TCACAm. As shown in Fig. 2C, DCACAm yields continuously grew with the increasing pH from 6.5 to 8.5 for Tyr-Ala and Ala-Tyr, whereas DCAN yields generally dropped with the pH increase. The formation and degradation patterns of DCACAm may be ascribed to the hydrolysis of DCAN and DCACAm. DCAN is relatively stable at pH 6.5, but can hydrolyze to form DCACAm as the alkalinity increases (Equation S1)²⁴. Even though DCACAm can hydrolyze to form DCAA, the hydrolysis rate of DCACAm was generally less than the formation rate of DCACAm from DCAN hydrolysis^{15,21}.

Previous studies have found that free Tyr may form DCACAm through initial substitution (reaction A in Fig. 3), decarboxylation, elimination and further substitution reactions (reaction D in Fig. 3), as well as hydrolysis reaction (reaction F in Fig. 3)¹⁷. In this study, free Tyr in the mixed 'Tyr+Ala' solution yielded DCACAm at similar concentrations as Tyr-Ala, which formed more DCACAm than Ala-Tyr, probably because the protection of the amino group in Ala-Tyr inhibited the formation of organic chloramines by initial substitution (reaction C in Fig. 3) as the first step for the formation of N-DBPs^{25,26}.

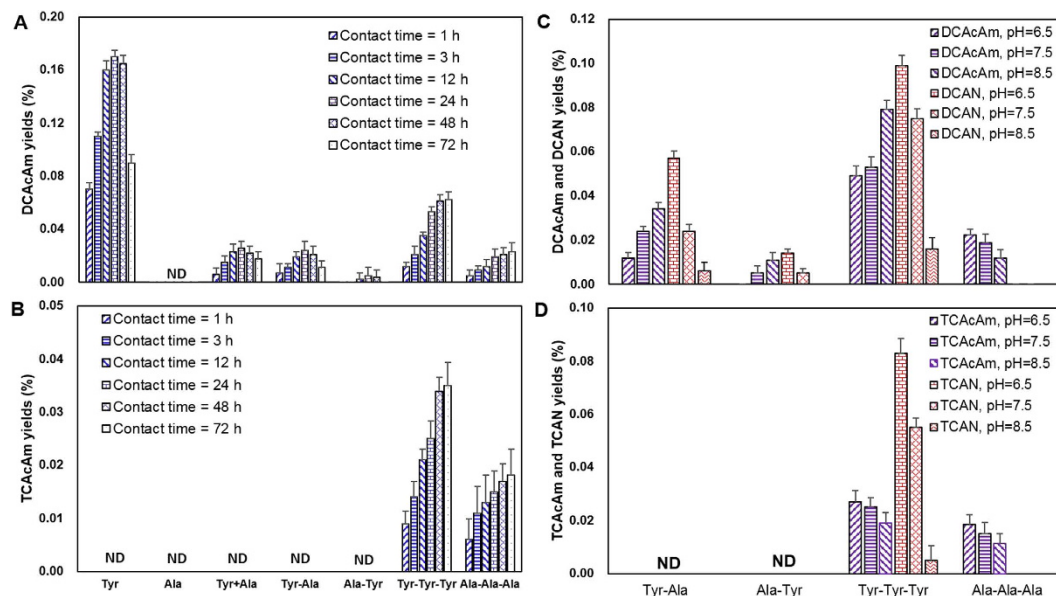


Figure 2. Formation of HACams during chlorination of the selected AAs at different contact times (DCACAm [A] and TCACAm [B]) and pH levels (DCACAm and DCAN [C] and TCACAm and TCAN [D]). AAs molar concentration = 0.05mM, Cl_2/N in model precursor (molar ratio) = 20, pH=7.5, except as noted. ‘Tyr+Ala’ represent the mixed solution of free Tyr and free Ala ($[Tyr] = [Ala] = 0.05$ mM). The bars represent the standard deviation of replicate measurements ($n = 3$).

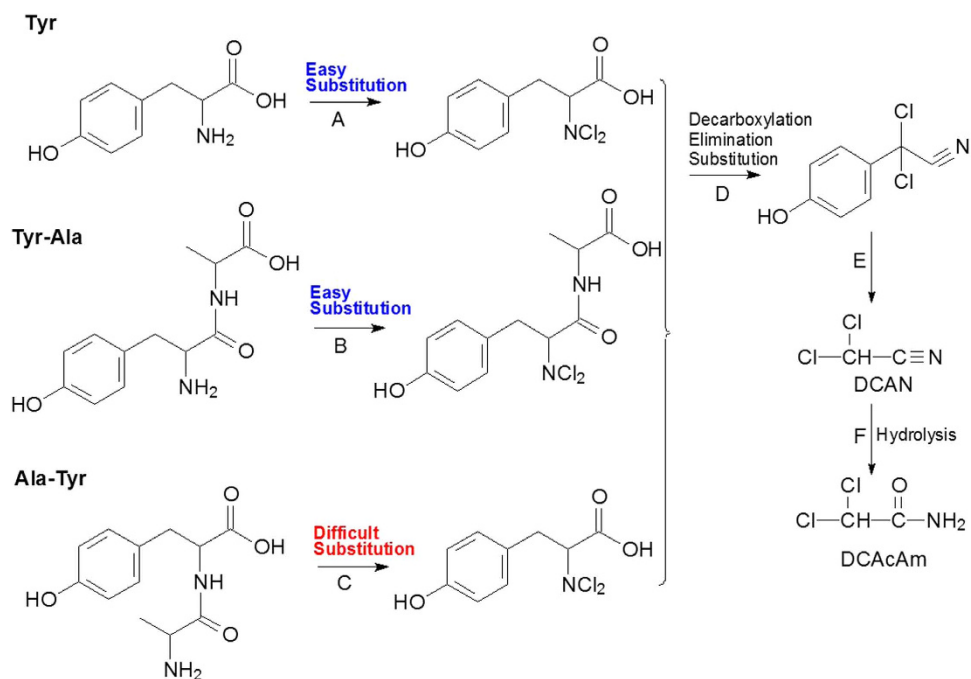


Figure 3. Proposed formation pathway of HACams from free Tyr, Tyr-Ala, and Ala-Tyr.

Impacts of peptide bonds in tripeptides on the HACAm formation during chlorination.

Figure 2 also presents the formation of DCACAm and TCACAm from the chlorination of two tripeptides (Tyr-Tyr-Tyr and Ala-Ala-Ala). Of note, Ala-Ala-Ala substantially transformed to DCACAm, whereas it is known that free Ala and Ala-Ala cannot form DCACAm above the detection limit during chlorination in the study, which was also found in the previous study¹⁵. Moreover, unlike free AAs (Tyr and Ala) and dipeptides (Ala-Ala, Tyr-Ala and Ala-Tyr), Tyr-Tyr-Tyr and Ala-Ala-Ala both produced TCACAm. The concentrations of DCACAm and TCACAm from the chlorination of Tyr-Tyr-Tyr and Ala-Ala-Ala increased over the entire study time from 1 to 72 h (Fig. 2A,B).

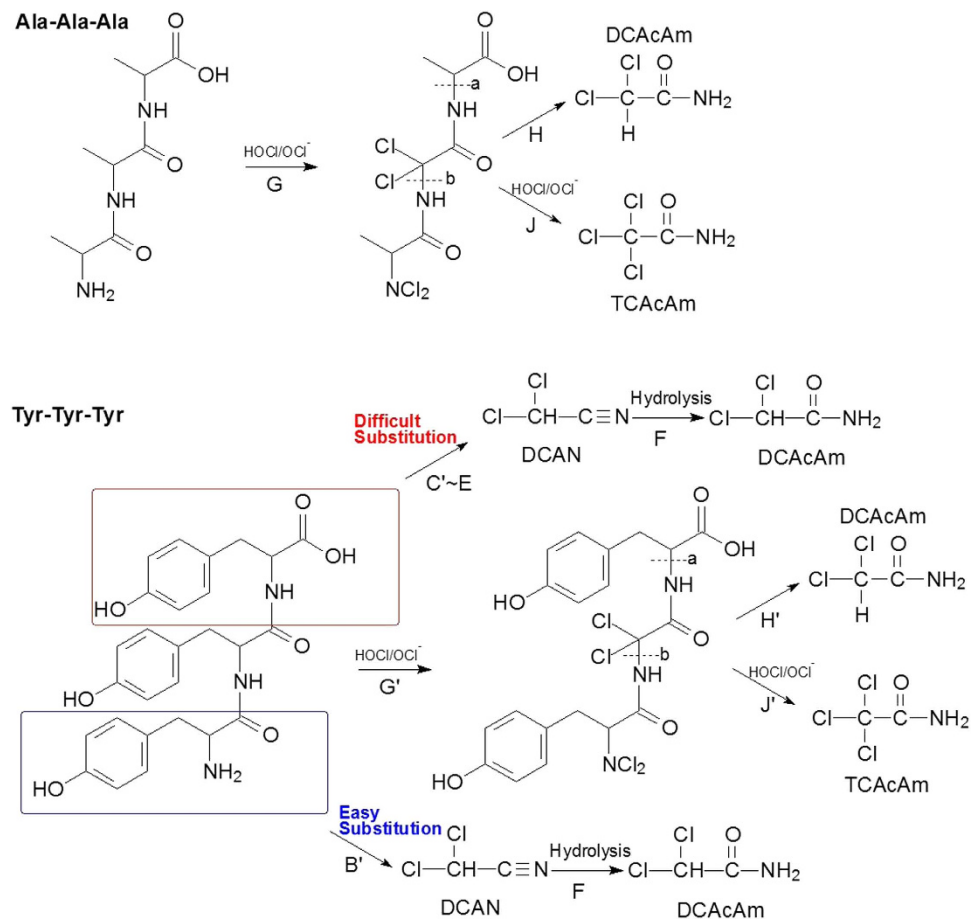


Figure 4. Proposed formation pathway of HACams from free Ala-Ala-Ala and Tyr-Tyr-Tyr.

As shown in Fig. 2C,D, Ala-Ala-Ala could not form DCAN and TCAN above the detection limit at three selected pH levels, and, DCACAm and TCACAm both decreased as pH increased. This finding indicated that the formation of DCACAm was independent of DCAN hydrolysis, which was different from the previous HACAm formation pathway (Fig. 3). Previous studies had found that chlorine substitution reaction would take place on the nitrogen atom at the amino-terminal function²⁷. However, no chlorine reactivity with the nitrogen atom at the peptide bond or the carboxyl-terminal residue was previously shown^{28–30}. Also, it has been appreciated that the hydrogen atoms of the methyl group between two carbonyl function groups are readily dissociated and the chlorine substitution is thus rapid^{31,32}. Accordingly, the methyl group between the two carbonyl function groups in Ala-Ala-Ala and Tyr-Tyr-Tyr could be substituted by chlorine (Reactions G and G' in Fig. 4), and probably form a small quantity of DCACAm and TCACAm through C–N bond breaking (bonds a and b)^{33–35} and further chlorine substitution and (Reactions H and J, H' and J' in Fig. 4). It should be noted that the proposed formation pathway of HACams during chlorination of oligopeptides was a speculative side reaction pathway. More research is needed to confirm the hypothesis.

Under a typical water treatment relevant pH, Cl₂ completely hydrolyzes, and the primary active chlorine species include HOCl and OCl⁻ (Equation S2). Since the equilibrium constant (K) for Equation (S3) is 2.9×10^{-8} at 25 °C, HOCl and OCl⁻ are the dominant species at pH 4.0–8.0 and 8.0–10.0, respectively³². Since HOCl is more reactive than OCl⁻ in water³², the chlorine substitution was faster at lower pH level (pH = 6.5) than at higher pH level (pH = 8.5). Consequently, this probably resulted in more production of DCACAm and TCACAm from Ala-Ala-Ala at pH 6.5 than at pH 8.5. Similarly, more TCACAm was formed from chlorination of Tyr-Tyr-Tyr at lower pH level. However, DCACAm formation from Tyr-Tyr-Tyr did not show a similar pattern with TCACAm, probably because the formation of DCACAm was not only from the chlorine substitution reaction adjacent to peptide bond (Reaction G' in Fig. 4), but also from the hydrolysis of DCAN (Reaction B' in Fig. 4) which is similar to the formation of DCACAm from Tyr-Ala (Reaction B in Fig. 3).

Impacts of peptide bonds on HACAm speciation during chlorination. The formation of brominated HACams is of particular interest, since they are more toxic than their chlorinated analogues^{2,6}. In order to examine the effect of peptide bonds on HACAm speciation from the selected free AAs and

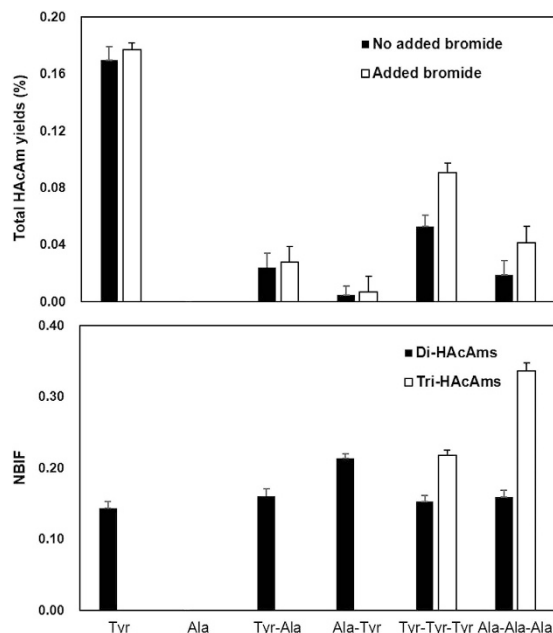


Figure 5. Total yields (A) and NBIF values (B) of HACams during chlorination of the selected AAs. AAs molar concentration = 0.05mM, Cl_2/N in model precursor (molar ratio) = 20, pH = 7.5, bromide/N in model precursor (molar ratio) = 2, except as noted. The bars represent the standard deviation of replicate measurements (n = 3).

combined AAs (oligopeptides), the AA water solution was added with bromide. As shown in Fig. 5A, bromide did not significantly change the yields of total HACams from chlorination of free Tyr and dipeptides (Tyr-Ala and Ala-Tyr), but it increased the yields of total HACams from the tripeptides (Tyr-Tyr-Tyr and Ala-Ala-Ala). Generally, bromide can form HOBr during chlorination as shown in Equation (S4)^{36,37}. Compared with HOCl, HOBr has a low dissociation of degree and a high oxidizability than $\text{HOCl}^{20,32}$, and thus the methyl group between the two carbonyl functions in Ala-Ala-Ala and Tyr-Tyr-Tyr was more easily substituted by HOBr than HOCl (Reactions G and G' in Fig. 4), and probably form more brominated HACams.

In order to further investigate the speciation of HACams from the selected free AAs and oligopeptides, bromine incorporation factors (BIF) for HACams were calculated as in studies of other DBPs^{38,39}, the BIF being used as an index to describe the proportion of the HACams that can be partially or totally substituted with bromine atoms. The following formulae were applied to calculate BIF (equations (2) and (3)), where all concentrations are on a molar basis:

$$\text{BIF}(\text{di-HACams}) = \frac{[\text{BCAcAm}] + 2[\text{DBAcAm}]}{[\text{DCAcAm}] + [\text{BCAcAm}] + [\text{DBAcAm}]} \quad (2)$$

$$\text{BIF}(\text{tri-HACams}) = \frac{[\text{BDCAcAm}] + 2[\text{DBCACAm}] + 3[\text{TBACAm}]}{[\text{TCACAm}] + [\text{BDCAcAm}] + [\text{DBCACAm}] + [\text{TBACAm}]} \quad (3)$$

BIFs for di-HACams ranged from 0 (all DCACAm) to 2 (all DBACAm), and BIFs for tri-HACams ranged from 0 (all TCACAm) to 3 (all TBACAm). A tri-HACAm BIF of 1.0 means that the average tri-HACAm species is BDCACAm. To better compare BIFs, each was normalized by the number of halogens, where the normalized BIF (NBIF) for di-HACams was its BIF divided by 2 and the NBIF for tri-HACams was its BIF divided by 3 (i.e., both NBIFs range from 0 to 1), as shown in Fig. 5B. The NBIF values for all selected AAs were all between 0.1 and 0.35, which is in agreement with a recent study²³. The recent study investigated the NBIFs of HACams formed from the chlorination of several natural waters containing bromide at 50~200 $\mu\text{g}/\text{L}^{23}$. Of note, there was more bromine incorporation into tri-HACams than in di-HACams during chlorination of Tyr-Tyr-Tyr and Ala-Ala-Ala. For chlorinated HACams, more di-HACams (0.043% for Tyr-Tyr-Tyr, 0.015% for Ala-Ala-Ala) were formed than tri-HACams (0.018% for Tyr-Tyr-Tyr, 0.011% for Ala-Ala-Ala). Whereas the yields of brominated tri-HACams (0.029% for Tyr-Tyr-Tyr, 0.019% for Ala-Ala-Ala) were higher than that of brominated di-HACams (0.019% for Tyr-Tyr-Tyr, 0.007% for Ala-Ala-Ala). Especially for Ala-Ala-Ala, the NBIF (tri-HACams) was significantly higher than NBIF (di-HACams). As discussed earlier, unlike Tyr-Tyr-Tyr, Ala-Ala-Ala formed di-HACams and tri-HACams only through a single halogenating reaction adjacent to the peptide bond

(Fig. 4), and the yields of di-HAcAms (DCAcAm) and tri-HAcAms (TCAcAm) were similar (about 0.02%) when bromide was not added (Fig. 2A,B). This results indicated that brominated tri-HAcAms was more easily to be formed by halogen (chlorine and bromine) substitution reaction adjacent to the peptide bond than brominated di-HAcAms. As reported, brominated tri-HAcAms are more cytotoxic and genotoxic than their di-HAcAm analogues, therefore, DWTPs should give attention to the formation of HAcAms in those algal- and wastewater-impacted waters rich in peptide bonds and bromide.

Conclusions

The use of wastewater-impacted water as drinking water sources increases concerns of the exposure of N-DBPs (e.g., HAcAms), because wastewater-induced DON plays a key role as N-DBP precursors. In the previous studies to investigate the nitrogen origin of HAcAms, α -amine terminus of free AAs was focused on. However, free AAs only accounts for a small fraction of dissolved organic nitrogen (DON) pool in source waters. Due to the low HAcAm yields (<0.2%) from the chlorination of AAs, low-level free AAs cannot supply enough nitrogen in HAcAms in chlorinated drinking water. Combined AAs could be an important nitrogen source in HAcAm formation during chlorination, especially in algal- and wastewater-impacted water.

This study firstly revealed that the HAcAm formation from AAs in more complex structures (oligopeptides) was different to the formation from free AAs. Compared to free AAs, the peptide bonds in oligopeptides, including dipeptides and tripeptides, reduced the contribution of the combined AAs on the DCAcAm formation. However, the peptide bond in tripeptides produced more TCAcAm compared to free AAs which were unable to form TCAcAm. These results implied that the peptide bonds contributed to the formation of HAcAms, and thus played a likely more important role in prediction of specific N-DBPs (e.g., HAcAm) concentrations upon chlorination.

Apart from the most frequently and abundantly detected di-HAcAms, DWTPs should also consider the formation of tri-HAcAms in those algal- and wastewater-impacted waters rich in peptide bonds and bromide, because bromide probably promoted the formation of total HAcAms (esp., brominated tri-HAcAms), and bromine-containing tri-HAcAms have been shown to be more cytotoxic and genotoxic than di-HAcAm analogues. A benefit of improving the removal of combined AAs before chlorination disinfection is reduced formation of HAcAms and thus reduced health concerns.

References

- Shah, A. D. & Mitch, W. A. Halonitroalkanes, halonitriles, haloamides, and N-nitrosamines: A critical review of nitrogenous disinfection by-product formation pathways. *Environ. Sci. Technol.* **46**, 119–131 (2012).
- Richardson, S. D., Plewa, M. J., Wagner, E. D., Schoeny, R. & DeMarini, D. M. Occurrence, genotoxicity, and carcinogenicity of regulated and emerging disinfection by-products in drinking water: A review and roadmap for research. *Mutation. Res.* **636**, 178–242 (2007).
- McCurry, D. L. *et al.* Superior removal of disinfection by-product precursors and pharmaceuticals from wastewater in a staged anaerobic fluidized membrane bioreactor compared to activated sludge. *Environ. Sci. Technol. Lett.* **1**, 459–464 (2014).
- Plewa, M. J., Wagner, E. D., Muellner, M. G., Hsu, K. M. & Richardson, S. D. Comparative mammalian cell toxicity of N-DBPs and C-DBPs. In *Occurrence, Formation, Health Effects and Control of Disinfection By-Products in Drinking Water*. Karanfil, T., Krasner, S. W., Westerhoff, P., Xie, Y. Eds, American Chemical Society: Washington, DC, 233, pp 372 (2007).
- Richardson, S. D. & Ternes, T. A. Water analysis: Emerging contaminants and current issues. *Anal. Chem.* **86**, 2813–2848 (2014).
- Plewa, M. J. *et al.* Occurrence, synthesis, and mammalian cell cytotoxicity and genotoxicity of haloacetamides: An emerging class of nitrogenous drinking water disinfection byproducts. *Environ. Sci. Technol.* **42**, 955–961 (2008).
- Krasner, S. W. *et al.* Occurrence of a new generation of disinfection byproducts. *Environ. Sci. Technol.* **40**, 7175–7185 (2006).
- Chu, W. H., Gao, N. Y., Yin, D. Q., Krasner, S. W. & Templeton, M. R. Trace determination of 13 haloacetamides in drinking water using liquid chromatography triple quadrupole mass spectrometry with atmospheric pressure chemical ionization. *J. Chromatogr. A*. **1235**, 178–181 (2012).
- Bond, T., Templeton, M. R. & Graham, N. Precursors of nitrogenous disinfection by-products in drinking water: A critical review and analysis. *J. Hazard. Mater.* **235**, 1–16 (2012).
- Bronk, D. A. Dynamics of DON. In *Biogeochemistry of Marine Dissolved Organic Matter*. Hansell, D. A., Carlson, C. A., Eds, Academic Press: San Diego, CA, pp 153–247 (2002).
- Dotson, A. & Westerhoff, P. Occurrence and removal of amino acids during drinking water treatment. *J. AWWA*. **101**, 101–115 (2009).
- Mandalakis, M., Apostolaki, M., Tziaras, T., Polymenakou, P. & Stephanou, E. G. Free and combined amino acids in marine background atmospheric aerosols over the Eastern Mediterranean. *Atmos. Environ.* **45**, 1003–1009 (2011).
- Lundeen, R. A. & McNeill, K. Reactivity differences of combined and free amino acids: Quantifying the relationship between three-dimensional protein structure and singlet oxygen reaction rates. *Environ. Sci. Technol.* **47**, 14215–14223 (2013).
- Xie, P. C. *et al.* Comparison of permanganate preoxidation and preozonation on algae containing water: cell integrity, characteristics, and chlorinated disinfection byproduct formation. *Environ. Sci. Technol.* **47**, 14051–14061 (2013).
- Chu, W. H., Gao, N. Y., Deng, Y. & Krasner, S. W. Precursors of dichloroacetamide, an emerging nitrogenous DBP formed during chlorination or chloramination. *Environ. Sci. Technol.* **44**, 3908–3912 (2010).
- Huang, H., Wu, Q. Y., Hu, H. Y. & Mitch, A. W. Dichloroacetonitrile and dichloroacetamide can form independently during chlorination and chloramination of drinking waters, model organic matters, and wastewater effluents. *Environ. Sci. Technol.* **46**, 10624–10631 (2012).
- Chu, W. H., Gao, N. Y., Krasner, S. W., Templeton, M. R. & Yin, D. Q. Formation of halogenated C-, N-DBPs from chlor(am)ination and UV irradiation of tyrosine in drinking water. *Environ. Pollut.* **61**, 8–14 (2012).
- Chu, W. H., Gao, N. Y., Deng, Y. & Dong, B. Z. Formation of chloroform during chlorination of alanine in drinking water. *Chemosphere* **77**, 1346–1351 (2009).
- How, Z. T. *et al.* Analysis of free amino acids in natural waters by liquid chromatography-tandem mass spectrometry. *J. Chromatogr. A* **1370**, 135–146 (2014).
- Richardson, S. D. & Thruston, A. D. Tribromopyrrole, brominated acids and other disinfection by-products produced by disinfection of drinking water rich in bromide. *Environ. Sci. Technol.* **37**, 3782–3793 (2003).

21. Chu, W. H., Gao, N. Y., Yin, D. Q., Krasner, S. W. & Mitch, W. A. Impact of UV/H₂O₂ pre-oxidation on the formation of haloacetamides and other nitrogenous disinfection byproducts during chlorination. *Environ. Sci. Technol.* **48**, 12190–12198 (2014).
22. Na, C. Z. & Olson, T. M. Relative reactivity of amino acids with chlorine in mixtures. *Environ. Sci. Technol.* **48**, 12190–12198 (2014).
23. Chu, W. H., Gao, N. Y., Yin, D. Q. & Krasner, S. W. Formation and speciation of nine haloacetamides, an emerging class of nitrogenous DBPs, during chlorination or chloramination. *J. Hazard. Mater.* **260**, 806–812 (2013).
24. Reckhow, D. A., Platt, T. L., MacNeill, A. L. & McClellan, J. N. Formation and degradation of dichloroacetonitrile in drinking waters. *J. Water Supply Res. Technol.-Aqua.* **50**, 1–13 (2001).
25. Deng, Z., Yang, X., Shang, C. & Zhang, X. R. Electrospray ionization-tandem mass spectrometry method for differentiating chlorine substitution in disinfection byproduct formation. *Environ. Sci. Technol.* **48**, 4877–4884 (2014).
26. Yang, X., Fan, C., Shang, C. & Zhao, Q. Nitrogenous disinfection byproducts formation and nitrogen origin exploration during chloramination of nitrogenous organic compounds. *Water Res.* **44**, 2691–2702 (2010).
27. Armesto, X. L., Canle, L. M., Garcia, M. V., Losada, M. & Santaballa, J. A. N reactivity vs O reactivity in aqueous chlorination. *Int. J. Chem. Kinet.* **26**, 1135–1141 (1994).
28. Armesto, X. L., Canle, L. M., Garcia, M. V., Losada, M. & Santaballa, J. A. Chlorination of dipeptides by hypochlorous acid in aqueous solution. *Gazz. Chim. Ital.* **124**, 519–523 (1994).
29. Armesto, X. L. *et al.* Intracellular oxidation of dipeptides: Very fast halogenation of the amino-terminal residue. *J. Chem. Perkin Trans. 2*, 608–612 (2001).
30. Abia, L., Armesto, X. L., Canle, L. M., Garcia, M. V. & Santaballa, J. A. Oxidation of aliphatic amines by aqueous chlorine. *Tetrahedron* **54**, 521–530 (1998).
31. De, L. J., Merlet, N. & Dore, M. Chloration de composés organiques: demande en chlore et réactivité vis-à-vis de la formation des trihalométhanes. Incidence de l'azote ammoniacal. *Water Res.* **16**, 1437–1450 (1982).
32. Deborde, M. & von Gunten, U. Reactions of chlorine with inorganic and organic compounds during water treatment—kinetics and mechanisms: A critical review. *Water Res.* **42**, 13–51 (2008).
33. Nelson, N., Levy, R. B. & Catal, J. Advance in research on hydrodenitrogenation and its catalysts. *J. Catal.* **58**, 485–488 (1979).
34. Portefaix, J. L., Cattenot, M., Gueriche, M., Thivolle-Cazat, J. & Breyse, M. Conversion of saturated cyclic and noncyclic amines over a sulphided NiMo/Al₂O₃ catalyst: Mechanisms of carbon-nitrogen bond cleavage. *Catal. Today* **10**, 473 (1991).
35. Prins, R., Zhao, Y., Sivasankar, N. & Kukula, P. Mechanism of C-N bond breaking in hydrodenitrogenation. *J. Catal.* **234**, 509–512 (2005).
36. Bousher, A., Brimblecombe, P. & Midgley, D. Rate of hypobromite formation in chlorinated seawater. *Water Res.* **20**, 865–870 (1986).
37. Kumar, K. & Margerum, D. W. Kinetics and mechanism of general-acid-assisted oxidation of bromide by hypochlorite and hypochlorous acid. *Inorg. Chem.* **26**, 2706–2711 (1987).
38. Gould, J. P., Fitchhorn, L. E. & Urheim, E. Formation of brominated trihalomethanes: Extent and kinetics. In *Water Chlorination: Environmental Impact and Health Effects*. Jolley, R. L., Ed., Ann Arbor Science Publishers: Ann Arbor, MI, Vol. 4 (1983).
39. Hua, G., Reckhow, D. & Kim, J. Effect of bromide and iodide ions on the formation and speciation of disinfection byproducts during chlorination. *Environ. Sci. Technol.* **40**, 3050–3056 (2006).

Acknowledgments

This project is supported by the National Natural Science Foundation of China (51378366), and the National Major Science and Technology Project of China (2015ZX07406004 and Natural Science Foundation of Jiangsu Province, China (No. BK2012677)). Dr. Y. Deng works in this project under the support of Global Education Center at Montclair State University (New Jersey, USA). The authors also thank Stuart W. Krasner (Metropolitan Water District of Southern California, USA) for helpful suggestions in the study.

Author Contributions

H.W.C. conceived the experiments, analyzed the data, wrote the article and contributed to the critical revision of the article. X.L. performed the experiments and prepared figures. Y.N.G. provided technical support for DBP analysis. Y.D. wrote the article. Q.D.Y. conceived the experiments. M.D.L. did the experiments. F.T.C. provided support for sample pretreatment.

Additional Information

Supplementary information accompanies this paper at <http://www.nature.com/srep>

Competing financial interests: The authors declare no competing financial interests.

How to cite this article: Chu, W. *et al.* Peptide bonds affect the formation of haloacetamides, an emerging class of N-DBPs in drinking water: free amino acids versus oligopeptides. *Sci. Rep.* **5**, 14412; doi: 10.1038/srep14412 (2015).



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