www.nature.com/isme



Community shifts of actively growing lake bacteria after *N*-acetyl-glucosamine addition: improving the BrdU-FACS method

Yuva Tada^{1,2} and Hans-Peter Grossart^{1,3}

¹Department of Limnology of Stratified Lakes, Leibniz-Institute of Freshwater Ecology and Inland Fisheries (IGB), Stechlin-Neuglobsow, Germany; ²Atmosphere and Ocean Research Institute, The University of Tokyo, Kashiwa-shi, Japan and ³Institute for Biochemistry and Biology, Potsdam University, Potsdam, Germany

In aquatic environments, community dynamics of bacteria, especially actively growing bacteria (AGB), are tightly linked with dissolved organic matter (DOM) quantity and quality. We analyzed the community dynamics of DNA-synthesizing and accordingly AGB by linking an improved bromodeoxyuridine immunocytochemistry approach with fluorescence-activated cell sorting (BrdU-FACS). FACS-sorted cells of even oligotrophic ecosystems in winter were characterized by 16S rRNA gene analysis. In incubation experiments, we examined community shifts of AGB in response to the addition of N-acetyl-glucosamine (NAG), one of the most abundant aminosugars in aquatic systems. Our improved BrdU-FACS analysis revealed that AGB winter communities of oligotrophic Lake Stechlin (northeastern Germany) substantially differ from those of total bacteria and consist of Alpha-, Beta-, Gamma-, Deltaproteobacteria, Actinobacteria, Candidatus OP10 and Chloroflexi. AGB populations with different BrdU-fluorescence intensities and cell sizes represented different phylotypes suggesting that single-cell growth potential varies at the taxon level. NAG incubation experiments demonstrated that a variety of widespread taxa related to Alpha-, Beta-, Gammaproteobacteria, Bacteroidetes, Actinobacteria, Firmicutes, Planctomycetes, Spirochaetes, Verrucomicrobia and Chloroflexi actively grow in the presence of NAG. The BrdU-FACS approach enables detailed phylogenetic studies of AGB and, thus, to identify those phylotypes which are potential key players in aquatic DOM cycling.

The ISME Journal (2014) 8, 441–454; doi:10.1038/ismej.2013.148; published online 29 August 2013

Subject Category: Microbial ecology and functional diversity of natural habitats

Keywords: actively growing bacteria (AGB); bromodeoxyuridine (BrdU) immunocytochemistry; fluorescence-activated cell sorting (FACS); bacterial community composition; *N*-acetyl-glucosamine (NAG); Lake Stechlin

Introduction

In aquatic environments, dissolved organic matter characteristics greatly affect bacterial diversity and metabolic activity (Azam, 1998). During the past two decades, 16S rRNA gene-based techniques have been intensively used to characterize bacterioplankton communities and their spatio-temporal fluctuations in aquatic ecosystems. However, communities of actively growing bacteria (AGB), which are mainly responsible for dissolved organic matter turnover and cycling, remain largely unknown.

To determine community composition and to assess the phylotype-specific substrate uptake at

the single-cell level, the microautoradiographyfluorescence in situ hybridization method is a powerful tool in aquatic microbial ecology (Lee et al., 1999; Cottrell and Kirchman, 2000; Alonso and Pernthaler, 2005). A major advantage of this method is the flexible usage of a variety of organic substrates as tracers (Fuhrman and Azam, 1982; Kirchman et al., 1985; Cottrell and Kirchman, 2000; Alonso-Sáez and Gasol, 2007). However, fluorescence in situ hybridization (FISH)-based methods may encompass a broad range of the 'true' phylogenetic diversity due to the limited number of available FISH probes. Depending on the phylogenetic coverage of the oligonucleotide probes, various phylotypes and even ecotypes with greatly differing growth rates may be detected. Thus, it is difficult to evaluate the AGB community composition at the taxon level by just using FISH-based methods. In microautoradiography-fluorescence in situ hybridization, the relative 16S rRNA abundance and the ratio of 16S rRNA to total rRNA genes

Correspondence: H-P Grossart, Department of Limnology of Stratified Lakes, Leibniz-Institute of Freshwater Ecology and Inland Fisheries (IGB), Alte Fischerhütte 2 D-16775, Stechlin-Neuglobsow, Germany.

E-mail: hgrossart@igb-berlin.de

Received 18 March 2013; revised 18 July 2013; accepted 24 July 2013; published online 29 August 2013





have been used as indices for activity and for the potential growth rates of specific taxa in complex. marine bacterial communities (Schäfer et al., 2001; Gentile et al., 2006; Campbell et al. 2009, 2011; Lami et al., 2009). Growth and metabolic activity of individual cells may be also represented by the number of ribosomes per cell (Kemp et al., 1993; Fegatella et al., 1998; Kerkhof and Kemp, 1999). Thus, this method allows to determine the potential growth of bacterial assemblages without using tracers and hence incubation biases. However, the relationship between rRNA copy number and growth is of low reliability in complex bacterial assemblages of natural environments, which often consist of a vast number of bacterial taxa with different rRNA-growth relationships and growth stages.

In this study, we used a novel single-cell-based method that combines bromodeoxyuridine immunocytochemistry and fluorescence-activated cell sorting (BrdU-FACS). Bromodeoxyuridine (BrdU), a halogenated nucleoside, serves as a thymidine analog and has been used as a tracer of bacterial de novo DNA synthesis, presumably of AGB (Taniguchi and Hamasaki, 2008). BrdU incorporation and fluorescent-labeled antibody detection techniques have been frequently used for identifying the AGB in aquatic environments (Steward and Azam, 1999; Urbach et al., 1999; Pernthaler et al., 2002; Hamasaki et al., 2004; Warnecke et al., 2005; Tada et al., 2010, 2011). In addition, the method can be combined with FACS for 16S rRNA gene analysis, and, thus, provides a powerful tool for phylogenetic characterization of BrdU-fluorescence-labeled cells (Mow et al., 2007). Thereby, BrdU-FACS enables to evaluate the AGB community composition at the single-taxon level.

The BrdU-FACS has been successfully applied to bacterioplankton in eutrophic coastal seawater (Mou et al., 2007). As it can be assumed that bacterial growth in oligotrophic regions should be much lower than in the eutrophic coastal ocean, an improved method is required for analyzing the AGB community composition in oligotrophic environments. Therefore, we improved the BrdU-detection sensitivity by using the tyramide signal amplification technique. The tyramide signal amplification system enhances the fluorescence intensity of the BrdU-detection and consequently increases the sensitivity of FACS of bacterial cells in oligotrophic environments during winter.

Aminosugars such as *N*-acetyl-glucosamine (NAG) (C₈H₁₅NO₆) represent a major fraction of the natural dissolved organic matter pool and serve as important bacterial carbon and nitrogen sources in a variety of aquatic systems (Nedoma *et al.* 1994). In previous studies, high-NAG concentrations have been found in lakes, possibly due to algal excretion (Giroldo *et al.*, 2003), fecal pellets and exoskeleton of zooplankton (Lee and Fisher, 1992; Tang *et al.*, 2009), viral lysis or protozoan grazing of bacterial cells (Jørgensen *et al.*, 2003; Cloud-Hansen *et al.*,

2006) and other sources (Wurzbacher and Grossart, 2012). Recent microautoradiography-fluorescence *in situ* hybridization studies show that several bacterial phylotypes can utilize NAG as a carbon and nutrient source and that NAG has the potential for niche separation of closely related bacterial taxa (Beier and Bertilsson, 2011; Eckert *et al.*, 2012). However, information on bacterial taxa actively growing in response to NAG supply is still scarce.

The purpose of this study was to examine the AGB community composition of an oligotrophic lake in winter and their growth response to NAG addition by using an improved BrdU-FACS method.

Materials and methods

Study site and sample collection

Surface water (0 m) from oligotrophic Lake Stechlin, northeastern Germany (53°10′N, 13°02′E) was collected in triplicates on 28 February 2012 by using 21 pre-combusted glass bottles (rinsed three times with surface water). After sampling, incubation experiments were immediately set up by adding 35 ml water to 50 ml of sterilized centrifugation tubes (three times rinses with surface water). Then, BrdU (1 µM final concentration; Sigma-Aldrich, St Louis, MO, USA) or BrdU+NAG (1 µM and 10 µM final concentration, respectively) were added to the respective tubes. Control tubes were also established. All water samples were incubated at 4 °C (in situ temperature ± 2 °C) for 48 h. At the end of the incubation, 10 ml samples were filtered onto 0.1-µm pore-size polycarbonate membrane filters (25-mm Nuclepore Track-Etch polycarbonate membrane, 110605, Whatman) and fixed with 50% ethanol for 1 h. Filters were stored at -30 °C until further analysis. All incubations were carried out in the dark using triplicates. For determining bacterial abundance, bacteria on the 0.1-µm membrane filters were stained with 4', 6-diamidino-2-phenilindole (DAPI, $1 \mu g \, ml^{-1}$, for $5 \, min$) and counted using epifluorescence microscopy.

Immunodetection of BrdU-incorporating cells

Before BrdU-FACS analysis, we improved the BrdU-immunodetection procedure in terms of BrdU-detection buffer and anti-BrdU antibody concentration (see Supplementary Material). For the BrdU assay, all treatments were directly carried out in the glass vacuum filter holders (16315, Sartorius, Goettingen, Germany). Bacterial cells on the membrane filters were dehydrated with serial treatments in 80% and 100% ethanol each for 1 min. Filters were then treated with 0.01 mol l⁻¹ HCl for 5 min at room temperature and with a pepsin solution (0.5 mg ml⁻¹ in 0.01 N HCl) for 2 h at 37 °C. Thereafter, cells were washed three times with 15 ml phosphate-buffered saline (PBS) for 10 min and then treated with lysozyme (10 mg ml⁻¹ in Tris-EDTA

buffer; 10 mmol l⁻¹ Tris-HCl, 1 mmol l⁻¹ EDTA; pH 8.0) for 15 min at room temperature. After the permeabilization steps, intracellular DNA was denatured by a nuclease treatment (1:100 in incubation buffer with the BrdU Labeling and Detection Kit III, 1444611, Roche, Mannheim, Germany) for doublestranded DNA for 2h at 37 °C and washed three times with 15 ml PBS for 10 min. Thereafter, anti-BrdU monoclonal antibodies conjugated with peroxidase were diluted 1:200 (final) in freshly prepared antibody reaction buffer (0.1% Tween-20, and 0.5% acetylated bovine serum albumin in PBS buffer). Samples were incubated with the antibody solution for 120 min at 37 °C, which then was washed away (three times) with 10 ml phosphate-buffered saline with Tween-20 (0.05% Triton X-100 in PBS). The antibody signal was amplified by incubating the filters with a Alexa488-labeled tyramide diluted 1:500 in amplification buffer (10% [w/v] dextran sulfate, 2 M NaCl, 0.1% [v/v] blocking reagent and 0.0015% [v/v] H_2O_2 in PBS) for 45 min at 46 °C. Filter pieces were then washed three times with 15 ml phosphatebuffered saline with Tween-20 buffer for 10 min. Bacterial cells were counterstained with DAPI (1 µg ml⁻¹) for determining total bacterial numbers. The cells on the membrane were resuspended by shaking filters with vortex (maximum speed) twice in 1.5 ml phosphate-buffered saline with Tween-20 for 15 min at room temperature.

Flow cytometry and cell sorting

Sorting of BrdU-positive cells was performed with a FACSAria II flow cytometer (Sorb, Becton Dickinson, Heidelberg, Germany). The sheath solution consisted of 0.2-µm filtered and sterile PBS. BrdU-positive cells were detected by their green fluorescence emitted from Alexa488 (488 nm excitation and 515-545 nm emission), and fluorescence intensities were used as a proxy of growth rate. Bacterial cells (total) were detected by their blue fluorescence after ultraviolet excitation (405 nm excitation and 430–470 nm emission). Gate notation was based on the extent of BrdU-fluorescence intensity (green fluorescence intensity) and cell size (side scatter) (Figure 1a: P1, low; P2, medium; and P3, high). Bacterial cells were sorted by FACS into sterile 96-well plates containing 50 µl MQ water. Sorting was terminated when the number of sorted cells was >100000 counts for each fluorescence intensity fraction. The DAPI-stained cells of the control amendment were sorted as a control for denaturing gradient gel electrophoresis (DGGE) and clone library analyses (Figure 1b).

PCR amplification of the partial 16S rRNA gene for DGGE and clone library analyses

Sorted cells from each fraction were transferred from 96-well plates to sterile 0.2 ml tubes, and 3-5 freeze-thaw cycles were applied for cell lysis. The resulting lysates served as templates for the PCR amplification. For checking the replications of incubation experiments, FACS, and PCR amplification, we used the PCR-DGGE method with partial 16S rRNA gene amplification. Thereafter, the sequences of sorted cells were determined by the PCR-cloning method. For PCR-DGGE and PCRcloning analyses, partial 16S rRNA gene amplification was carried out with the eubacterial primer set (Schäfer and Muyzer, 2001). Protocols for PCR-DGGE and PCR-cloning are given in the Supplementary material. Sequences were aligned to known sequences in the DNA Data Bank of Japan (DDBJ) using BLAST (http://blast.ddbj.nig.ac.jp/blast/blastn).

Phylogenetic relationships were inferred from multiple alignments by using the ATGC 6.04 software (Genetyx Co., Tokyo, Japan) and phylogenetic trees were calculated using the neighbor-joining method and the MEGA 5.05 software (http:// www.megasoftware.net/; Tamura et al., 2011). All sequences were checked by the DECIPHER program (Wright et al., 2012) and the Mallard (http://www.bioinformatics-toolkit.org/ Mallard, Cardiff, UK). All nucleotide sequences have been deposited in the DDBJ nucleotide sequence database under accession numbers AB781370-781443 and AB831234-831258.

Results

Methodological improvements

Before combining BrdU-immunodetection and FACS, we optimized the BrdU-detection buffer and anti-BrdU antibody concentrations (Supplementary material). The BrdU-immunodetection efficiency with the new solutions was two to three fold higher than those previously published (Supplementary Figure S1). In addition, the modification of the antibody concentration resulted in the elimination of all 'false' BrdU-positive cells and yielded a proportion of BrdU-positive cells of $36 \pm 1\%$ of total bacteria (Supplementary Figure S2). BrdU-fluorescence signals from cells treated with the peroxidaselabeled antibody (the tyramide signal amplification system) were enhanced as compared with those with the fluorescein isothiocyanate-labeled antibody (Supplementary Figure S3). After all BrdU immunocytochemical reactions in the glass vacuum filter holder, the residual cells on the filter still accounted for $76 \pm 6\%$ and $84 \pm 13\%$ of the initial cells of the control and BrdU amendments, respectively (Supplementary Figure S4).

FACS sorting of BrdU-incorporating cells

Bacterial abundance in the Lake Stechlin winter sample was $7.7 \pm 1.7 \times 10^5$ cells ml⁻¹. Total cell numbers and the proportion of BrdU-positive cells of DAPI-stained cells after 48 h incubation are given in Table 1. On the basis of the levels of



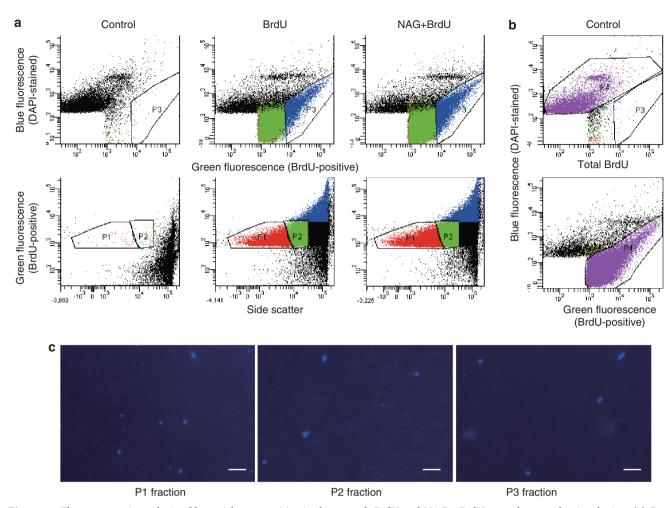


Figure 1 Flow-cytometric analysis of bacterial communities in the control, BrdU and NAG+BrdU amendments after incubation. (a) Gate notation (P1, low; P2, medium; and P3, high) based on the extent of BrdU-fluorescence intensity (Green fluorescence intensity) and cell size (Side scatter). (b) Gates for sorting the control fraction and counting total BrdU-positive cells. (c) Photomicrograph of the sorted cells. Bar = $5 \mu m$.

Table 1 Total cell numbers and proportion of BrdU-positive cells of sorted fractions

Treatments	Total cells ($ imes 10^5$ cells ml $^{-1}$)		Resuspension rate (%)	${\it Proportion~of~BrdU-positive~cells~of~DAPI-stained~cells}$			
	Direct count	FCM count		P1: low (%)	P2: medium (%)	P3: high (%)	Total (%)
Control BrdU NAG+BrdU	12.2 ± 1.4 19.5 ± 2.7 23.6 ± 4.5	2.6 ± 0.5 13.3 ± 2.2 16.9 ± 5.6	21 68 72	0.5 ± 0.2 19.7 ± 3.2 22.4 ± 2.9	1.0 ± 0.4 28.1 ± 6.0 27.7 ± 1.8	0.4 ± 0.1 9.5 ± 1.4 9.9 ± 0.6	2.0 ± 0.6 63.9 ± 7.0 69.2 ± 3.6

Abbreviations: BrdU, bromodeoxyuridine; DAPI, 4', 6-diamidino-2-phenilindole; FCM, flow cytometry; NAG, N-acetyl-glucosamine.

BrdU-fluorescence intensity which is positively related to growth potential (Tada et al., 2010) and cell sizes, cells in the BrdU and NAG+BrdU amendments were categorized into three fractions, low (P1), medium (P2) and high (P3) BrdU-fluorescence intensity and cell size populations (Figure 1a). BrdU-positive cells of each fraction in the controls representing 'false' positives accounted for <2% of total cells. Cell abundances in the BrdU and NAG+ BrdU amendments were higher than in the controls, indicating active bacterial growth after tracer addition. BrdU-positive cells in the BrdU and NAG+BrdU

amendments accounted for 64% and 69% of total cells, respectively. BrdU-positive cells for total bacteria were a bit higher than the sum of the P1, P2, and P3 fractions, as not all cells were gated for the FACS analysis.

DGGE analysis of the sorted populations

DGGE-banding patterns of 16S rRNA genes of the sorted bacterial populations revealed that the community composition of control and AGB substantially differed (Figure 2). Moreover, the addition

of NAG resulted in a clear community shift, in particular of AGB.

Phylogenetic analysis of AGB community

To analyze AGB community compositions, eight clone libraries were constructed (Table 2). These included the initial, control and samples of all three fractions (P1, P2 and P3) of each BrdU and NAG+ BrdU amendment. The 19-41 sequences obtained for each library (average length of 580 bp) were classified into 4-25 operational taxonomic units at an evolutionary distance of 0.03 (Table 2). Operational taxonomic units are distributed among 13 major taxa which are commonly found in freshwater lakes (Figures 3–6, and, Table 3). In the clone library of initial amendment, sequences related to Actinobacteria (including acl and aclV lineages) and Bacteroidetes (including bacI lineage) frequently appeared and accounted for 32% and 29% of total sequences, respectively (Figure 3 and Table 3). Bacterial taxa which appeared in the BrdU amendment differed from those in the control (Table 3). Whereas Alphaproteobacteria represented the frequent sequences (37% of total sequences) in the clone library of the control (Figure 3), Betaproteobacteria accounted for the largest fraction (53% of total sequences) of total sequences in that of the BrdU amendment. Furthermore, the AGB community composition of the NAG+BrdU amendment greatly differed from those of the control and BrdU amendments (Figure 3 and Table 3). The most prevailing taxon in the NAG+ BrdU clone library consisted of Gammaproteobacteria (31% of total sequences). Betaproteobacterial sequences were the second most frequent ones (28% of total sequences) in the NAG+BrdU clone library. Sequences related to Alphaproteobacteria, Bacteroidetes, Actinobacteria, Firmicutes, Plancto-Spirochaetes, Verrucomicrobia mvcetes, *Chloroflexi* also appeared in the NAG+BrdU clone library, but only accounted for 9%, 6%, 9%, 6%, 5%, 2%, and 2% of the total sequences, respectively. Nitrospira sequences appeared only in the clone library of the initial amendment and accounted for 5% of total sequences.

Apparent differences between BrdU and NAG+BrdU amendments were related to Alpha-, Beta-, Gammaproteobacteria, Bacteroidetes, Planctomycetes, Spirochaetes, Verrucomicrobia and Firmicutes taxa (Table 3). Alphaproteobacterial sequences related to Sphingomonas appeared in clone libraries of both amendments, whereas those of the Candidatus Pelagibacter sp. (alfV lineage) and Pseudaminobacter sp. only occurred in the NAG+BrdU amendment (Table 3 and Figure 5). Betaproteobacteria related to the betI lineage (including Acidovorax sp., Limnohabitans sp., Variovorax sp. and Hydrogenophaga sp.) and the betVII lineage (including Janthinobacterium sp. and Herbaspirillum sp.) were observed in both amendments. However, members of the

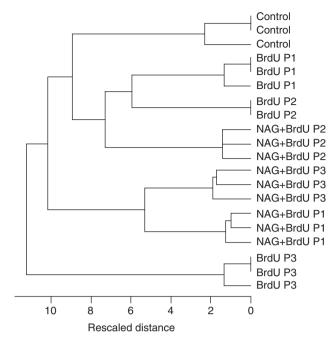


Figure 2 Cluster analysis of DGGE-banding patterns of the different BrdU-fluorescence intensity fractions (P1, P2 and P3) of the control, BrdU and NAG+BrdU amendments. The BrdU-P2 sorted samples represent duplicates, all other samples represent triplicates. The DGGE-banding patterns of sorted cells indicate very good within-treatment reproducibility (incubated, sorted and amplified independently).

 Table 2
 Distributions of sequences and OTUs from clone

 libraries

	Initial	Control	BrdU		NA	NAG + BrdU		
			P1	P2	Р3	P1	P2	РЗ
No. of sequences No. of OTUs ^a	41 25	23 14				21 13	24 12	19 13

Abbreviations: BrdU, bromodeoxyuridine; NAG, *N*-acetyl-glucosamine; OTUs, Operational taxonomic units. OTUs was defined with CD-HIT (Li and Godzik, 2006) assuming a 97% sequence similarity level.

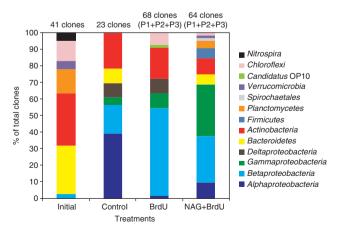


Figure 3 Percentages of clones represented by the major phylogenetic groups of bacteria in libraries of 16S rRNA genes in the initial, control, BrdU and NAG+BrdU amendments.

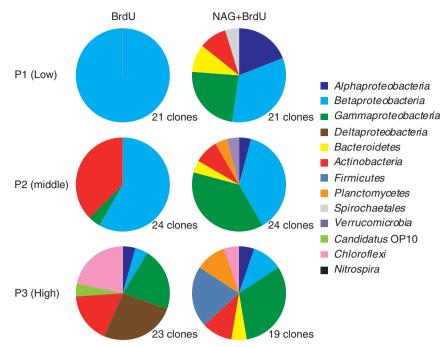


Figure 4 Bacterial community composition (clone libraries) in different BrdU-fluorescence intensity fractions (P1, P2 and P3) of both BrdU and NAG+BrdU amendments.

lineage including Methylophilus sp. exclusively appeared in the NAG+BrdU amendment. Whereas Gammaproteobacteria affiliated to Acinetobacter sp. were found in both BrdU and NAG+BrdU amendments, those affiliated to Pseudomonas sp. were solely present in the NAG+BrdU amendment. Actinobacteria related to acI, acII, and acIV as well as members of the Firmicutes, Bacteroidetes, Spirochaetes, Verrucomicrobia and Planctomycetes appeared in the NAG+BrdU amendment (Table 3 and Figure 6). In contrast, members of the Deltaproteobacteria and Candidatus OP10 did not appear in the NAG + BrdU, but in the BrdU amendment.

Phylogenetic analysis of sorted AGB based on BrdUfluorescence intensities and cell sizes

AGB community composition of the three sorted fractions with different BrdU-fluorescence intensities and cell sizes (P1, P2 and P3) were determined both in BrdU and NAG+BrdU amendments (Figure 4). In the BrdU amendment, Betaproteobacteria formed the largest portion of total sequences in the P1 and P2 fractions (100% and 58%, respectively). In particular, sequences related to Rodoferax sp. (betI lineage) and Janthinobacterium sp. (betVII lineage) were found in the P1 fraction (Figure 5 and Table 3). The proportion of Actinobacteria increased in the P2 fraction (38%), and especially members of the Nocardioides and Propionibacterium appeared in the P2 and P3 fractions. The number of operational taxonomic units in the P3 fraction was higher than in all other fractions, in particular the proportion of Deltaproteobacteria in the BrdU amendment increased in this fraction (26%).

Moreover, members of the Alphaproteobacteria, Candidatus OP10, and Chloroflexi were exclusively observed in the P3 fraction and accounted for 4%, 4%, and 22% of total sequences, respectively.

The AGB community in the NAG+BrdU amendment also consisted of major members of common bacterial freshwater clusters, namely Alpha-, Beta-, Gammaproteobacteria, Bacteroidetes, Actinobacteria, Firmicutes, Planctomycetes, Spirochaetes, Verrucomicrobia and Chloroflexi (Figures 5, 6, and Table 3). Betaproteobacteria represented the most prevailing taxon in the P1 fraction (33% of total sequences). Sequences related to Rhodoferax sp., Methylophilus sp. and Janthinobacterium sp. belonging to the betI, betIV and betVII lineages, respectively were observed in the NAG+BrdU clone library of the P1 fraction (Figure 5). Furthermore, sequences of Limnohabitans sp. and Hydrogenophaga sp. both belonging to the betI lineage appeared in the P3 fraction. The proportion of *Actinobacteria* in the P1, P2 and P3 fractions of the NAG+BrdU amendment accounted for 10%, 8%, and 11% of all sequences, respectively, whereby acI and acIV Actinobacteria mainly appeared in the P2 and P3 fractions, respectively (Figure 6). Sequences of the acII lineage were observed in both P1 and P2 fractions. The proportion of Gammaproteobacteria in the P1, P2 and P3 fractions accounted for 24%, 38%, and 32% of total sequences, respectively. Members of the Bacteroidetes were observed in all fractions and contributed between 4 and 10% of total sequences. In contrast, a single member of the *Firmicutes*, solely observed in the P3 fraction, accounted for 21% of the total sequences. Spirochaetes, Verrucomicrobia and Chloroflexi were solely observed in the P1, P2



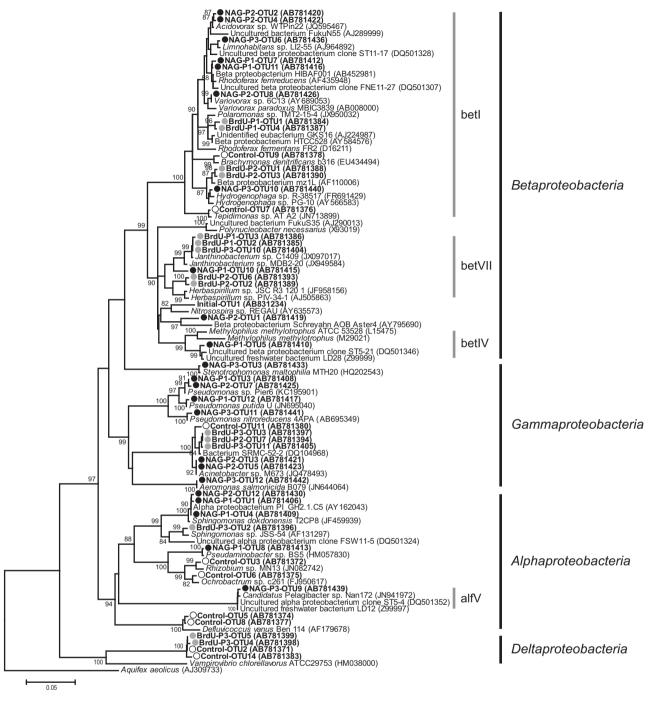


Figure 5 Phylotype distribution in the initial sample, control and fractions of different BrdU-fluorescence intensity (P1, P2 and P3) of BrdU and NAG+BrdU amendments: phylogenetic tree of all obtained 16S rRNA gene sequences (clone libraries; 0.03 distance OUTs) related to known Proteobacteria. Bootstrap values >80% are indicated above the branches. The scale bar represents the estimated 5% sequence divergence. Open circles indicate operational taxonomic units (OTUs) from control amendment, gray circles indicate OTUs from the BrdU amendment and solid circles indicate OTUs from the NAG+BrdU amendment.

and P3 fraction of the NAG+BrdU amendment, respectively.

Discussion

Methodological improvements

For the BrdU immunocytochemical detection, loss of cells is one of the most serious problems. In the

present study, >75% of all cells could be kept on the filter (Supplementary Figure S4) after performing all BrdU immunocytochemical steps directly in the glass vacuum filter holders to minimize cell loss. In previous BrdU studies, adhesive materials such as agarose (Pernthaler et al., 2002) and poly-l-lysin (Tada et al., 2010) were used to keep the cells on the filter. For BrdU-FACS application, however, cell



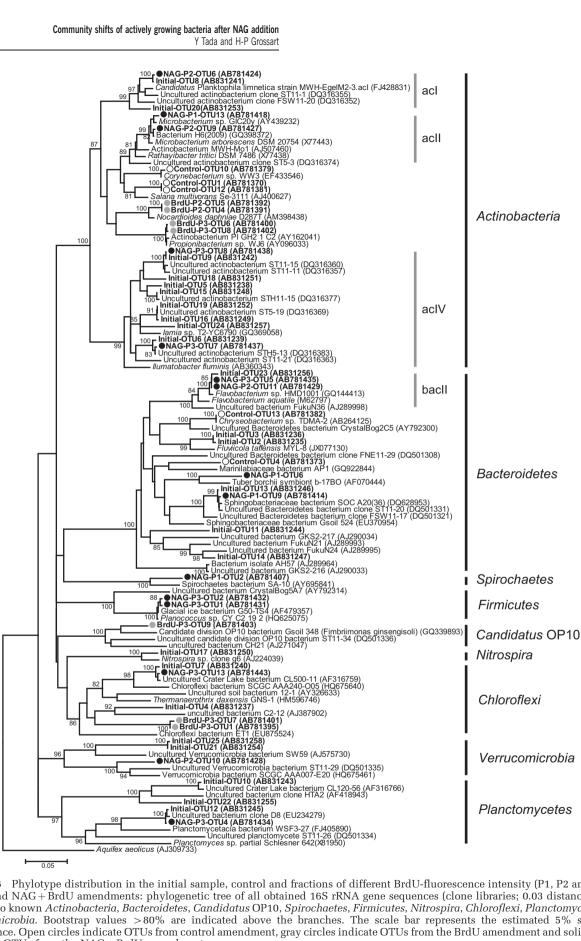


Figure 6 Phylotype distribution in the initial sample, control and fractions of different BrdU-fluorescence intensity (P1, P2 and P3) of BrdU and NAG+BrdU amendments: phylogenetic tree of all obtained 16S rRNA gene sequences (clone libraries; 0.03 distance OUTs) related to known Actinobacteria, Bacteroidetes, Candidatus OP10, Spirochaetes, Firmicutes, Nitrospira, Chloroflexi, Planctomycetes and Verrcomicrobia. Bootstrap values >80% are indicated above the branches. The scale bar represents the estimated 5% sequence divergence. Open circles indicate OTUs from control amendment, gray circles indicate OTUs from the BrdU amendment and solid circles indicate OTUs from the NAG+BrdU amendment.



 $\textbf{Table 3} \ \ \textbf{The presence and absence of phylogenetic group of 16S rRNA gene sequences at each sample}$

Phylum	Sequence names and OTUs	Closest sequences	Treatments			
			Initial	Control	BrdU	NAG+BrdU
Alphaproteobacteria	Control-OTU3	Rhizobium sp. MN13		+		
	Control-OTU5, OTU8	Defluvicoccus vanus Ben 114		+		
	Control-OTU6	Ochrobactrum sp. c261		+		
	BrdU-P3-OTU2 NAG-P1-OTU1, OTU4/NAG-P2-OTU12	Sphingomonas sp. JSS-54 Sphingomonas dokdonensis			+	1
	NAG-P1-OTU8	Pseudaminobacter sp. BS5				+ +
	NAG-P3-OTU9	Candidatus Pelagibacter sp. Nan172				+
Betaproteobacteria	Initial-OTU1	Nitrosospira sp. REGAU	+			
	Control-OTU7	Tepidimonas sp. AT A2		+		
	Control-OTU9 BrdU-P1-OTU1, OTU4	Brachymonas denitrificans b316 Polaromonas sp. TMT2-15-4		+	+	
	BrdU-P1-OTU2, OTU3/ BrdU-P3- OTU10/NAG-P1-OTU10	Janthinobacterium sp. C1409			+	+
	BrdU-P2-OTU1, OTU3	Beta proteobacterium mz1L			+	
	BrdU-P2-OTU2, OTU6	Herbaspirillum sp. JSC R3 120 1			+	1
	NAG-P1-OTU5	Methylophilus methylotrophus ATCC 53528				+
	NAG-P1-OTU7, OTU11	Rhodoferax ferrireducens				+
	NAG-P2-OTU1	Beta proteobacterium Schreyahn				+
	NAC DO OFFICE OFFICE	AOB Aster4				
	NAG-P2-OTU2, OTU4 NAG-P2-OTU8	Acidovorax sp. WTPin22 Variovorax sp. 6C13				+
	NAG-P3-OTU6	Limnohabitans sp. LI2–55				+ +
	NAG-P3-OTU10	Hydrogenophaga sp. R-38517				+
Gammaproteobacteria	Control-OTU11/NAG-P2-OTU3, OTU5	Acinetobacter sp. M673		+		+
	BrdU-P2-OTU7, BrdU-P3-OTU3, OTU11				+	
	NAG-P1-OTU3/NAG-P2-OTU7 NAG-P1-OTU12	Pseudomonas sp. Pier6 Pseudomonas putida U				+ +
	NAG-P3-OTU3	Stenotrophomonas maltophilia MTH20				+
	NAG-P3-OTU11	Pseudomonas nitroreducens 4APA				+
Delta	NAG-P3-OTU12	Aeromonas salmonicida B079				+
Deltaproteobacteria Bacteroidetes	Control-OTU2, OTU14/BrdU-P3-OTU4, OTU5 Initial-OTU2, OTU3	Vampirovibrio chlorellavorus ATCC:29753 Fluviicola taffensis MY-8	+	+	+	
Datitionation	Control-OTU4	Marinilabiaceae bacterium AP1	'	+		
	Control-OTU13	Chryseobacterium sp. TDMA-2		+		
	NAG-P1-OTU6	Tuber borchii symbiont b-17BO				+
	Initial-OTU13/NAG-P1-OTU9	Sphingobacteriaceae bacterium SOC A20(36)	+			+
	Initial-OTU23/NAG-P2-OTU11/ NAG-P3-OTU5 Initial-OTU11	Flavobacterium sp. HMD1001 Uncultured bacterium GKS2–217	+			+
	Initial-OTU14	Uncultured bacterium FukuN24	+			
Actinobacteria	Initial-OTU5, OTU15	Uncultured actinobacterium STH11–15	+			
	Initial-OTU16, OTU19	Uncultured actinobacterium ST5–19	+			
	Initial-OTU24	Iamia sp. T2-YC6790	+			
	Control-OTU1, OTU12 Control-OTU10	Salana multivorans Se-3111 Corynebacterium sp. WW3		+ +		
	BrdU-P2-OTU4, OTU5	Nocardioides daphniae D287T			+	
	BrdU-P3-OTU6, OTU8	Propionibacterium sp. WJ6			+	
	NAG-P1-OTU13 Initial-OTU6, OTU20/NAG-P2-OTU6	Microbacterium sp. GIC20y Candidatus Planktophila limnetica	+			++
	NAG-P2-OTU9 Initial-OTU6/NAG-P3-OTU7	MWH-EgelM2-3.acI Bacterium H6 (2009) Uncultured actinobacterium	+			+ +
	Initial-OTU9, OTU18/NAG-P3-OTU8	STH5–13 Uncultured actinobacterium	+			+
Firmicutes	NAG-P3-OTU1, OTU2	ST11-11 Planococcus sp. CY C2 19 2				+
Planctomycetes	Initial-OTU10	Uncultured Crater Lake bacterium XL120-56	+			
	Initial-OTU22 Initial-OTU12/NAG-P3-OTU4	Uncultured bacterium clone HTA2 Uncultured bacterium clone D8	+			
Spirochaetes	NAG-P1-OTU2	Spirochaetes bacterium SA-10	+			+ +
Verrucomicrobia	Initial-OTU21, OTU25	Uncultured Verrucomicrobia bacterium SW59	+			1
	NAG-P2-OTU10	Verrucomicrobia bacterium SCGC AAA007-E20				+





Table 3 (Continued)

Phylum	Sequence names and OTUs	Closest sequences	Treatments				
			Initial	Control	BrdU	NAG + BrdU	
Candidatus OP10	BrdU-P3-OTU9	Candidate division OP10 bacterium Gsoil 348			+		
Chloroflexi	Initial-OTU4 BrdU-P3-OTU1, OTU7	Uncultured bacterium C2-12 Thermanaerothrix daxensis GNS-1	+		+		
	Initial-OTU7/NAG-P3-OTU13	Uncultured Crater Lake bacterium CL500-11	+			+	
Nitrospira	Initial-OTU17	Nitrospira sp. clone g6	+				

Abbreviations: BrdU, bromodeoxyuridine; NAG, N-acetyl-glucosamine; OTUs, Operational taxonomic units.

adhesive materials exert negative effects for cell sorting owing to related changes of the cell's surface properties. Our method circumvents this critical

The cell resuspension rate of the control was lower than those of BrdU and NAG+BrdU amendments both with higher percentages of BrdU-positive cells (Table 1). This indicates that AGB might better detach from the filter than the inactive or dormant cells. Furthermore, bacterial community composition differed between the treatments (Table 3), whereby surface characteristics of cells (for example, cell wall and extracellular matrix) may differ between phylotypes, suggesting that the resuspension rate varies with the changing bacterial community composition.

We optimized antibody concentration and antibody reaction buffer for the improved BrdU-immunodetection (Supplementary Figures S1 and S2). These improvements greatly enhanced the BrdU-detection sensitivity as compared with previous studies (Supplementary Figure S3). These methodological changes thus allowed us to apply the BrdU-FACS method for phylogenetic analyses of AGB in oligotrophic waters during winter even at the taxon level.

In the BrdU amendment, about 63.9% of the total cells (DAPI-stained) were able to incorporate the BrdU (Table 1), which was relatively similar to the proportion of AGB in the NAG+BrdU amendment (69.2%). We added the BrdU at a lower concentration (final 1 µM) than for other BrdU studies. Although the added BrdU did not substantially contribute to the total dissolved organic carbon pool in oligotrophic Lake Stechlin, we cannot rule out that the added BrdU has served as an additional bacterial substrate. Thus, AGB should include bacteria with the potential for active growth.

Biases during incubation experiments

The bacterial community composition of the control amendment differed from that of the initial sample (Table 3), indicating that there were some biases during our incubation experiments. Possible biases should be a confinement bias and long incubation time. However, DGGE analysis revealed a good

replication of incubation experiments (Figure 2), indicating that these biases exist in a consistent manner in all bottles.

AGB community composition

Clone library analyses revealed that AGB community composition substantially differed from that of total bacteria in the control, suggesting that community composition of total bacteria is the result of an interplay of many factors including bacterial dormancy (Jones and Lennon, 2010), protozoan grazing (Šimek et al., 2001; Pernthaler et al., 2001; Eiler and Bertilsson, 2007; Jones et al., 2009; Salcher et al., 2010; Eckert et al., 2012) and viral lysis (Weinbauer and Höfle, 1998; Thingstad, 2000). AGB with different BrdU-fluorescence intensities composed of different bacterial taxa (Figures 4, 5, and 6) indicate that single-cell growth potential varies at the taxon level.

In the BrdU amendment, *Polaromonas* sp. as well as Hydrogenophaga sp. (both belonging to the betI lineage) and Janthinobacterium sp. (betVII lineage) frequently appeared, particularly in the P1 and P2 fractions. This notion indicates that these phylotypes should grow in spite of lower temperature. In previous FISH studies, however, their population size was tightly correlated with water temperature (Hahn et al., 2005; Wu and Hahn, 2006a, b). In contrast, our incubation experiments were performed at a low temperature (ca. 4 °C), suggesting that low nutrient availability was the critical factor in the control, whereas the low water temperature did not limit the bacterial growth in both BrdU and NAG + BrdU amendments.

Actinobacteria represent one of the most prominent lineages in surface waters of temperate, stratified lakes (Glöckner et al., 2000; Hahn et al., 2003; Warnecke et al., 2005; Allgaier et al., 2007; Salcher et al., 2010; Rösel and Grossart, 2012; Rösel et al., 2012). In our study, the actinobacterial sequences were frequently observed in clone libraries of the P2 and P3 fractions (Figures 3 and 4). Especially, *Propionibacterium* sp. observed in the P3 fraction expressed a high growth potential and thus are of potential importance for organic matter cycling in oligotrophic Lake Stechlin in winter.

In addition, our results revealed that sequences of Candidatus OP10 and Chloroflexi, which are known to contribute to only a minor fraction of lake bacteria (Newton et al. 2011) occur in the P3 fraction, suggesting that these quantitatively minor groups may be highly active and hence contribute to the organic matter cycling in Lake Stechlin in winter.

BrdU-immunofluorescence intensity should be influenced by genome size and GC content of individual cells. Genome sizes of marine and freshwater bacteria greatly vary at the species level, and range from 1 to 10 Mbp (for example, Giovannoni et al., 2005; Zeng et al., 2012; Garcia et al., 2013). For instance, the genome size of acl Actinobacteria is small (<2 Mbp) and their GC content high (ca. 40%) (Garcia et al., 2013), suggesting that their real growth potential might be underestimated by BrdU incorporation. Such differences should affect the variation of BrdU-fluorescence intensities of specific phylogenetic taxa (potentially within the same taxon level) and hence should reflect their actual activity level.

Community shift of AGB in response to NAG addition Several FISH-based studies have shown that NAG supply in lakes lead to bacterial community shifts at the phylum or genus level (Beier and Bertilsson, 2011; Eckert et al., 2012). However, closely related bacterial phylotypes exhibit variable preferences for organic substrates (Hunt et al., 2008; Alonso et al., 2009; Buck et al., 2009). Thus, the analysis of community shifts at a higher phylogenetic resolution is a prerequisite to elucidate the effects of specific organic substrates, for example, NAG, in freshwater lakes. The present study revealed a pronounced growth response of various limnetic bacterial phyla and classes to NAG supply (Figures 5, 6, and Table 3). This notion is consistent with previous findings indicating that widespread taxa of marine and freshwater bacteria have the ability to take up NAG (Nedoma et al., 1994; Riemann and Azam, 2002; Cottrell and Kirchman, 2000; Beier and Bertilsson, 2011).

Particularly, Betaproteobacteria related to Acidovorax, Rhodoferax, Limnohabitans, Variovorax, Hydrogenophaga (betI lineage), Methylophilus (betIV lineage) and Janthinobacterium (betVII lineage) formed substantial portions of the P1 and P2 fractions of the NAG+BrdU amendment (Figures 4 and 5). Sequences related to the widespread Limnohabitans and Hydrogenophaga (betl lineage) also occurred in the P3 fraction, indicating that these bacteria grow well on NAG or at least on NAG degradation products. This is in accordance to microautoradiography-fluorescence in situ hybridization results in mesotrophic Lake Zürich (Eckert et al., 2012). However, genomic and physiological studies on type strains of Limnohabitans and Variovorax do not support active NAG utilization (Yoon et al., 2006; Miwa et al., 2008; Hahn et al.,

2010a, b; Kasalický *et al.*, 2010). A recent study has revealed that members of *Limonohabitans* perform well on a variety of alga-derived substrates; and regarding their genome size they have a more substrate-responsive ability 'high metabolic IQ' (Kasalický et al., 2013) rather than abundant acI and *Polynucleobacter* groups. Thus, it seems likely that this lineage solely should take up the NAG and alga-derived hydrolysis products and, therefore, depends on the chitinolytic activity of other phylogenetic groups.

NAG utilization by Polynucleobacter (betII lineage), especially PnecC, has been demonstrated (Alonso et al., 2009). However, sequences related to Polynucleobacter were absent in all clone libraries. A possible reason for this notion could be a limited resolution of our phylogenetic analyses (Table 2). Another reason could be the relatively low water temperature, which has been indicated as a critical factor regulating *Polynucleobacter* populations (Hahn et al., 2005; Wu and Hahn, 2006b). Our findings thus suggest that the contribution of the betII lineage to NAG cycling is diminished during winter.

Gammaproteobacteria frequently appeared in the P2 and P3 fractions of the NAG+BrdU amendment, indicating that they have a key role in the NAG degradation during winter. A large part of Gammaproteobacteria utilizing NAG was closely related to potentially chitinolytic soil bacteria (Figure 5). Their active growth suggests a tight terrestrial-aquatic coupling and point to a possibly important biogeochemical role, for example, in NAG degradation, of otherwise rare soil bacteria in this lake.

Members of the Bacteroidetes were not detected in the BrdU but in the NAG + BrdU amendment (Figure 6 and Table 3). This suggests that NAG may stimulate specific members of limnic Bacteroidetes. Previous studies also revealed that Bacteroidetes (including Cytophaga-Flavobacteria) take up NAG (Beier and Bertilsson, 2011; Eckert et al., 2012). In contrast to Lake Zürich (where NAG-utilizing Bacteroidetes predominantly belonged to the bacVI lineage) (Eckert et al., 2012), in our study, they mainly belonged to the bacII lineage. The phylum Bacteroidetes, however, expresses a huge diversity in freshwater ecosystems (Newton et al., 2011; Eckert et al., 2012). On the basis of our results it seems likely that key lineages of NAG-utilizing Bacteroidetes differ with environmental variables such as season and physiological state of the phyto- and zooplankton.

Members of Actinobacteria are known as important consumers of NAG in freshwater lakes (Beier and Bertilsson, 2011; Eckert et al., 2012). Our results revealed that Microbacterium sp., uncultured lineages of acI and acIV formed the majority of Actinobacteria in the NAG+BrdU treatment. Especially, the acIV sequences frequently appeared in the P3 fraction, implying that this lineage represents highly active Actinobacteria involved in NAG degradation. AcI and acIV Actinobacteria are



highly abundant and widespread (Glöckner et al., 2000; Burkert et al., 2003; Warnecke et al., 2005; Allgaier and Grossart, 2006; Newton et al., 2007; Buck et al., 2009). However, little is known about their ecological role, especially as no pure cultures of these lineages are available (Newton et al., 2011). In this study, acI and acIV lineages appeared in the P2 or P3 fraction. Furthermore, single-cell genome analysis has revealed that acI Actinobacteria have a chitinase-like gene (Garcia et al., 2013). This suggests that acI Actinobacteria are actively involved in chitin degradation. Our results indicate that NAG may be an important energy source for members of acI and acIV Actinobacteria and implies participation in NAG cycling in Lake Stechlin.

In general, members of *Verrucomicrobia* and *Planctomycetes* are known as common but minor phyla in freshwater lakes and little is known about their diversity and ecological role (Newton *et al.*, 2011). Our results suggest that these groups could also be involved in NAG degradation in Lake Stechlin. However, their overall contribution to freshwater NAG cycling still remains to be determined.

Conclusions

Combining an improved BrdU-FACS method with clone library analysis revealed an AGB winter community in an oligotrophic lake. AGB in fractions with different BrdU-fluorescence intensities and cell sizes revealed that single-cell growth potential varies at the taxon level, which might be used to identify bacterial groups with different growth rates using flow cytometry. Moreover, NAG incubation experiments revealed that widespread limnetic phyla and classes actively grow in response to NAG addition even at low temperature. These results imply that NAG cycling can be important for organic matter cycling and bacterial community composition in freshwater ecosystems in winter.

Acknowledgements

We thank Solvig Pinnow for excellent technical assistance and Claudia Dziallas for valuable discussions. We also thank Koji Hamasaki, Koji Suzuki and Hisashi Endo for their great support for this study. We appreciate the help by members of the MIBI group and the technical stuff of IGB Neuglobsow. This study was supported by a fellowship of the Leibniz Institute of Freshwater Ecology and Inland Fisheries (IGB) and by a grant of the German Science Foundation (DFG-GR1540/17-1).

References

Allgaier M, Grossart H-P. (2006). Diversity and seasonal dynamics of *Actinobacteria* populations in four lakes in northeastern Germany. *Appl Environ Microbiol* **72**: 3489–3497.

- Allgaier M, Brückner S, Jaspers E, Grossart H-P. (2007). Intra- and inter-lake variability of free-living and particle-associated *Actinobacteria* communities. *Environ Microbiol* 9: 2728–2741.
- Alonso C, Pernthaler J. (2005). Incorporation of glucose under anoxic conditions by bacterioplankton from coastal North Sea surface waters. *Appl Environ Microbiol* 71: 1709–1716.
- Alonso C, Zeder M, Piccini C, Conde D, Pernthaler J. (2009). Ecophysiological differences of betaproteobacterial populations in two hydrochemically distinct compartments of a subtropical lagoon. *Environ Micro*biol 11: 867–876.
- Alonso-Sáez L, Gasol JM. (2007). Seasonal variation in the contribution of different bacterial groups to the uptake of low molecular weight-compounds in NW Mediterranean coastal waters. *Appl Environ Microbiol* **73**: 3528–3535.
- Azam F. (1998). Microbial control of oceanic carbon flux: the plot thickens. *Science* **280**: 694–696.
- Beier S, Bertilsson S. (2011). Uncoupling of chitinase activity and uptake of hydrolyses products in freshwater bacterioplankton. *Limnol Oceanogr* **56**: 1179–1188.
- Buck U, Grossart H-P, Amann R, Pernthaler J. (2009). Substrate incorporation patterns of bacterioplankton populations in stratified and mixed waters of a humic lake. *Environ Microbiol* 11: 1854–1865.
- Burkert U, Warnecke F, Babenzien D, Zwirnmann E, Pernthaler J. (2003). Members of a readily enriched β-proteobacterial clade are common in surface waters of a humic lake. *Appl Environ Microbiol* **69**: 6550–6559.
- Campbell BJ, Yu LY, Straza TRA, Kirchman DL. (2009). Temporal changes in bacterial rRNA and rRNA genes in Delaware coastal waters. *Aquat Microb Ecol* **57**: 123–135.
- Campbell BJ, Yu L, Heidelberg JF, Kirchman DL. (2011). Activity of abundant and rare bacteria in a coastal ocean. *Proc Natl Acad Sci USA* **108**: 12776–12781.
- Cloud-Hansen KA, Peterson SB, Stabb EV, Goldman WE, McFall-Ngai MJ, Handelsman J. (2006). Breaching the great wall: peptidoglycan and microbial interactions. *Nat Rev Microbiol* **4**: 710–716.
- Cottrell MT, Kirchman DL. (2000). Natural assemblages of marine proteobacteria and members of the *Cytophaga–Flavobacter* cluster consuming low- and high-molecular-weight dissolved organic matter. *Appl Environ Microbiol* **66**: 1692–1697.
- Eckert EM, Salcher MM, Posch T, Eugster B, Pernthaler J. (2012). Rapid successions affect microbial N-acetyl-glucosamine uptake patterns during a lacustrine spring phytoplankton bloom. Environ Microbiol 14: 794–806.
- Eiler A, Bertilsson S. (2007). Flavobacteria blooms in four eutrophic lakes: linking population dynamics of freshwater bacterioplankton to resource availability. Appl Environ Microbiol 73: 3511–3518.
- Fegatella F, Lim J, Kjelleberg S, Cavicchioli R. (1998). Implications of rRNA operon copy number and ribosome content in the marine oligotrophic ultramicrobacterium Sphingomonas sp. strain RB2256. Appl Environ Microbiol 64: 4433–4438.
- Fuhrman JA, Azam F. (1982). Thymidine incorporation as a measure of heterotrophic bacterioplankton production in marine surface waters: evaluation and field results. *Mar Biol* **66**: 109–120.

- Garcia SL, McMahon KD, Martinez-Garcia M, Srivastava A, Sczyrba A, Stepanauskas R et al. (2013). Metabolic potential of a single cell belonging to one of the most abundant lineages in freshwater bacterioplankton. ISME J **7**: 137–147.
- Gentile G, Giuliano L, D'Auria G, Smedile F, Azzaro M, De Domenico M et al. (2006). Study of bacterial communities in Antarctic coastal waters by a combination of 16S rRNA and 16S rDNA sequencing. Environ Microbiol 8: 2150-2161.
- Giovannoni SJ, Tripp HJ, Givan S, Podar M, Vergin KL, Baptista D et al. (2005). Genome streamlining in a cosmopolitan oceanic bacterium. Science 309: 1242-1245.
- Giroldo D, Vieira AAH, Paulsen BS. (2003). Relative increase of deoxy sugars during microbial degradation of an extracellular polysaccharide released by a tropical freshwater Thalassiosira sp. (Bacillariophyceae). J Phycol **39**: 1109–1115.
- Glöckner F-O, Zaichikov E, Belkova N, Denissova L, Pernthaler J, Pernthaler A et al. (2000). Comparative 16S rRNA analysis of lake bacterioplankton reveals globally distributed phylogenetic clusters including an abundant group of actinobacteria. Appl Environ Microbiol 66: 5053-5065.
- Hahn MW, Lünsdorf H, Wu Q, Schauer M, Höfle MG, Boenigk J et al. (2003). Isolation of novel ultramicrobacteria classified as actinobacteria from five freshwater habitats in Europe and Asia. Appl Environ Microbiol 69: 1442-1451.
- Hahn MW, Pöckl M, Wu QL. (2005). Low intraspecific diversity in a *Polynucleobacter* subcluster population numerically dominating bacterioplankton of a freshwater pond. Appl Environ Microbiol 71: 4539-4547.
- Hahn MW, Kasalický V, Jezbera J, Brandt U, Jezberová J, Šimek K. (2010a). *Limnohabitans curvus* gen. nov., sp. nov., a planktonic bacterium isolated from a freshwater lake. Int J Syst Evol Microbiol 60: 1358-1365.
- Hahn MW, Kasalický V, Jazbera J, Brandt U, Šimek K. (2010b). Limnohabitans australis sp. nov., isolated from freshwater pond, and emended description of the genus Limnohabitans. Int J Syst Evol Microbiol 60: 2946-2950.
- Hamasaki K, Long RA, Azam F. (2004). Individual cell growth rates of marine bacteria, measured by bromodeoxyuridine incorporation. Aquat Microb Ecol 35: 217-227.
- Hunt DE, David LA, Gevers D, Preheim SP, Alm EJ, Polz MF. (2008). Resource partitioning and sympatric differentiation among closely related bacterioplankton. Science 320: 1081-1085.
- Jones SE, Newton RJ, McMahon KD. (2009). Evidence for structuring of bacterial community composition by organic carbon source in temperate lakes. Environ Microbiol 11: 2463-2472.
- Jones SE, Lennon JT. (2010). Dormancy contributes to the maintenance of microbial diversity. Proc Natl Acad Sci USA 107: 5881-5886.
- Jørgensen NOG, Stepanaukas R, Pedersen AGU, Hansen M, Nybroe O. (2003). Occurrence and degradation of peptidoglycan in aquatic environments. FEMS Microbiol Ecol 46: 269-280.
- Kasalický V, Jezbera J, Šimek K, Hahn MW. (2010). planktonicus Limnohabitans sp. nov.. Limnohabitans parvus sp. nov., two novel planktonic Betaproteobacteria isolated from a freshwater reservoir. Int J Syst Evol Microbiol 60: 2710–2714.

- Kasalický V, Jezbera J, Šimek K, Hahn MW. (2013). The diversity of the *Limnohabitans* genus, an important group of freshwater bacterioplankton, by characterization of 35 isolated strains. PLoS One 8: e58205.
- Kemp PF, Lee S, LaRoche J. (1993). Estimating the growth rate of slowly growing marine bacteria from RNA content. Appl Environ Microbiol 59: 2594-2601.
- Kerkhof L, Kemp P. (1999). Small ribosomal RNA content in marine Proteobacteria during non-steady-state growth. FEMS. Microbiol Ecol 30: 253-260.
- Kirchman D, K'Ness E, Hodson R. (1985). Leucine incorporation and its potential as a measure of protein synthesis by bacteria in natural aquatic systems. Appl Environ Microbiol 49: 599-607.
- Lami R, Ghiglione JF, Desdevises Y, West NJ, Lebaron P. (2009). Annual patterns of presence and activity of marine bacteria monitored by 16S rDNA-16S rRNA fingerprints in the coastal NW Mediterranean Sea. *Aquat Microb Ecol* **54**: 199–210.
- Lee BG, Fisher NS. (1992). Decomposition and release of elements from zooplankton debris. Mar Ecol Prog Ser **88**: 117-128.
- Lee N, Nielsen PH, Andreasen KH, Juretschko S, Nielsen JL, Schleifer KH et al. (1999). Combination of fluorescent in situ hybridization and microautoradiography - a new tool for structure-function analyses in microbial ecology. Appl Environ Microbiol 65: 1289-1297.
- W, Godzik A. (2006). Cd-hit: a fast program Li for clustering and comparing large sets of protein nucleotide sequences. **Bioinformatics** or 1658-1659.
- Miwa H, Ahmed I, Yoon J, Yokota A, Fujiwara T. (2008). Variovorax boronicumulans sp. nov., a boron-accumulating bacterium isolated from soil. Int J Syst Evol Microbiol 58: 286-289.
- Mou X, Hodson RE, Moran MA. (2007). Bacterioplankton assemblages transforming dissolved compounds in coastal seawater. Environ Microbiol 9: 2025-2037.
- Nedoma J, Vrba J, Hejzlar J, Šimek K, Straskrabová V. (1994). N-acetylglucosamine dynamics in fresh-warer environments - concentration of amino-sugars, extracellular enzyme-activities, and microbial uptake. Limnol Oceanogr 39: 1088-1100.
- Newton RJ, Jones SE, Helmus MR, McMahon KD. (2007). Phylogenetic ecology of the freshwater Actinobacteria acI lineage. Appl Environ Microbiol 73: 7169-7176.
- Newton RJ, Jones SE, Eiler A, McMahon KD, Bertilsson S. (2011). A guide to the natural history of freshwater lake bacteria. Microbiol Mol Biol Rev 75: 14-49.
- Pernthaler J, Posch T, Šimek K, Vrba J, Pernthaler A, Glöckner FO et al. (2001). Predator-specific enrichment of actinobacteria from a cosmopolitan freshwater clade in mixed continuous culture. Appl Environ Microbiol 67: 2145-2155.
- Pernthaler A, Pernthaler J, Schattenhofer M, Amann R. (2002). Identification of DNA-synthesizing bacterial cells in coastal North Sea plankton. Appl Environ Microbiol 68: 5728-5736.
- Riemann L, Azam F. (2002). Widespread N-acetyl-dglucosamine uptake among pelagic marine bacteria and its ecological implications. Appl Environ Micro*biol* **68**: 5554–5562.
- Rösel S, Grossart H-P. (2012). Contrasting dynamics in activity and community composition of free-living and particle-associated bacteria in spring. Aquat Microb *Ecol* **66**: 169–181.



- Rösel S, Allgaier M, Grossart H-P. (2012). Long-term characterization of free-living and particle-associated bacterial communities in Lake Tiefwaren reveals distinct seasonal patterns. *Microb Ecol* **64**: 571–583.
- Salcher MM, Pernthaler J, Posch T. (2010). Spatiotemporal distribution and activity patterns of bacteria from three phylogenetic groups in an oligomesotrophic lake. *Limnol Oceanogr* **55**: 846–856.
- Schäfer H, Bernard L, Courties C, Lebaron P, Servais P, Pukall R et al. (2001). Microbial community dynamics in Mediterraneen nutrient-enriched seawater mesocosms: changes in the genetic diversity of bacterial populations. FEMS Microbiol Ecol 34: 243–253.
- Schäfer H, Muyzer G. (2001). Denaturing Gradient Gel Electrophoresis in Marine Microbial Ecology. Academic Press: San Diego, CA, USA.
- Šimek K, Pernthaler J, Weinbauer MG, Hornak K, Dolan JR, Nedoma J et al. (2001). Changes in bacterial community composition and dynamics and viral mortality rates associated with enhanced flagellate grazing in a mesoeutrophic reservoir. Appl Environ Microbiol 67: 2723–2733.
- Steward GF, Azam F. (1999). Bromodeoxyuridine as an alternative to ³H-thymidine for measuring bacterial productivity in aquatic samples. *Aquat Microb Ecol* **19**: 57–66.
- Tada Y, Taniguchi A, Hamasaki K. (2010). Phylotypespecific growth rates of marine bacteria measured by bromodeoxyuridine immunocytochemistry and fluorescence in situ hybridization. Aquat Microb Ecol 59: 229–238.
- Tada Y, Taniguchi A, Nagao I, Miki T, Uematsu M, Tsuda A et al. (2011). Differing growth responses of major phylogenetic groups of marine bacteria to natural phytoplankton blooms in the western North Pacific Ocean. Appl Environ Microbiol 77: 4055–4065.
- Tamura K, Peterson D, Peterson N, Stecher G, Nei M, Kumar S. (2011). MEGA5: molecular evolutionary genetics analysis using maximum likelihood, evolutionary distance, and maximum parsimony methods. *Mol Biol Evol* **28**: 2731–2739.
- Taniguchi A, Hamasaki K. (2008). Community structures of actively growing bacteria shift along a north-south transect in the western North Pacific. *Environ Microbiol* **10**: 1007–1017.

- Tang KW, Bickel SL, Dziallas C, Grossart H-P. (2009). Microbial activities accompanying decomposition of cladoceran and copepod carcasses under different environmental conditions. Aquat Microb Ecol 57: 89–100.
- Thingstad TF. (2000). Elements of a theory for the mechanisms controlling abundance, diversity, and biogeochemical role of lytic bacterial viruses in aquatic systems. *Limnol Oceanogr* **45**: 1320–1328.
- Urbach E, Vergin KL, Giovannoni SJ. (1999). Immunochemical detection and isolation of DNA from metabolically active bacteria. *Appl Environ Microbiol* **65**: 1207–1213.
- Warnecke F, Sommaruga R, Seka R, Hofer J, Pernthaler J. (2005). Abundances, identity, and growth state of Actinobacteria in mountain lakes of different UV transparency. *Appl Environ Microbiol* **71**: 5551–5559.
- Weinbauer MG, Höfle MG. (1998). Significance of viral lysis and flagellate grazing as controlling factors of bacterioplankton production in a eutrophic lake. *Appl Environ Microbiol* **64**: 431–438.
- Wright ES, Yilmaz LS, Noguera DR. (2012). DECIPHER, a search-based approach to chimera identification for 16S rRNA gene sequences. *Appl Environ Microbiol* **78**: 717–725.
- Wu QL, Hahn MW. (2006a). Differences in structure and dynamics of *Polynucleobacter* communities in a temperate and a subtropical lake, revealed at three phylogenetic levels. *FEMS Microbiol Ecol* **57**: 67–79.
- Wu QL, Hahn MW. (2006b). High predictability of the seasonal dynamics of a species-like *Polynucleobacter* population in a freshwater lake. *Environ Microbiol* 8: 1660–1666.
- Wurzbacher C, Grossart H-P. (2012). Improved detection and identification of aquatic fungi and chitin in aquatic environments. *Mycologia* **104**: 1267–1271.
- Yoon J-H, Kang S-J, Oh T-K. (2006). Variovorax dokdonensis sp. nov., isolated from soil. Int J Syst Evol Microbiol 56: 811–814.
- Zeng Y, Kasalický V, Šimek K, Koblížek M. (2012). Genome sequence of two freshwater betaproteobacterial isolates, *Limnohabitans* species strains Rim 28 and Rim 47, indicate their capabilities as both photoautotrophs and ammonia oxidizers. *J Bacteriol* 194: 6302–6303.

Supplementary Information accompanies this paper on The ISME Journal website (http://www.nature.com/ismei)